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University
of Glasgow

Statistical Tools in Environmental Impact Assessment

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University of Glasgow

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School of Mathematics & Statistics

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Abstract

Environmental Impact Assessment consists of assessing the possible effects of a project of development on the environment. Statistical methods which are used in this process include the ability to detect the magnitude of an impact at the time in which an intervention has occurred, and the ability to detect when an intervention has occurred if the time of impact is not available. This impact at the time at which the intervention has occurred is also known as a 'changepoint'. This project consists of evaluating methods used to assess the magnitude of these 'impacts', and the methods which can be used to determine where an impact has occurred if the time of impact is not available.

A number of datasets have been made available for analysis within this thesis, each of these datasets are expected to contain impacts within them. By reviewing the techniques and methods which are available for statisticians, applications to these datasets were made.

The classic approach to quantifying the magnitude of change within a series when the time of the impact is known is Before-After-Control-Impact design, which uses a dummy variable within a linear regression model to determine whether a significant change in mean can be detected. This thesis reviewed the history and the approaches used by statisticians when using this method, and the adaptations that can be made. BACI was then applied to our own datasets to determine whether a significant change in mean could be detected.

Many statistical changepoint detection methods are available to statisticians when the location of a changepoint, if any, is not known. By summarising the methods available and calculating the power of various methods via a simulation study, a number of changepoint detection methods were applied to real life data.

Finally, various modelling techniques were applied to the available datasets and by incorporating terms to indicate the detected location of changepoints, we could determine whether adding these terms gives better fitting models.

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Chapter 1

Introduction to Changepoint analysis and its applications

1.1 What is Environmental Impact Assessment

Environmental Impact Assessment (EIA) is an important policy tool which is used to weigh evidence to assess environmental interventions, in both public and private projects. Adopted 25 years ago, it is a procedure that ensures that the environmental implications of decisions are taken into account before the decisions are made. Furthermore, the complexity of the process is increased by the diversity of the disciplines involved, for example – environmental science, social science and economics. Because of this diversity of disciplines, the decision making procedure and final result should not be based solely on scientific grounds but also on social viewpoints.

‘Environmental Impact Assessment’ is not strictly a statistical tool but may include some statistical elements to provide evidence for or against environmental intervention effect on the environment.

Created in 1985, the EIA directive has been changed three times- in 1997, 2003 and 2009. Primarily the focus of EIA is to determine whether projects will cause an environmental impact by predicting the effects of development. The definition of an EIA as stated by the International Association for Impact Assessment is:

"The process of identifying, predicting, evaluating and mitigating the biophysical, social, and other relevant effects of development proposals prior to major decisions being taken, and commitments made."

It also states the objectives of an EIA are as follows,

- *To ensure that environmental considerations are explicitly addressed and incorporated into the development decision making process;*
- *To anticipate and avoid, minimize or offset the adverse significant biophysical, social and other relevant effects of development proposals;*
- *To protect the productivity and capacity of natural systems and the ecological processes which maintain their functions; and*
- *To promote development that is sustainable and optimizes resource use and management opportunities*

An EIA is used to weigh up the evidence with respect to both positive and negative consequences with regards to environmental impact. In Europe, projects must go through a screening process to determine if an EIA is needed. Those projects already listed as an 'Annex I' project must have an EIA carried out; otherwise any project listed as an 'Annex II' project does not require an EIA (unless stated otherwise). If it is not listed as an Annex I or Annex II, consideration is taken on deciding whether there should be an EIA at the discretion of the state. For example, Annex I projects would include major power plants, chemical works, waste disposal incineration and major roads. Annex II projects would include quarries and opencast, some intensive livestock rearing, overhead transmission lines and in some cases, wind farms. Furthermore, the location of development is a major factor in the decision process as the environmental sensitivity is different in geographical areas. Therefore existing land use, relative abundance of natural resources and the absorption capacity of the natural environment must be taken into account in deciding if an EIA is required.

If now the statistical nature of an EIA is considered, then there are issues concerning the sampling frame over space but also over time. Data can be obtained by sampling from various locations around the project site over time, allowing us to determine through modelling whether the project has had any effect on the environment.

In a statistical sense, impact is determined through modelling of the observed data to infer whether a change has taken place which can then be linked to the development.

Detecting a change in an environmental time series can be regarded as evidence of change attributed to the development or project in question if the 'experiment' is designed properly.

There are a number of statistical methods available to identify changes in environmental time series. If the time at which an impact is expected to have occurred is available, Before-After analysis is a simple way of determining whether a significant change in a variable has been detected. When 'control' sites are available (sites which are relatively close to the impact site but are not affected by the intervention) Before-After-Control-Impact analysis can be implemented. A literature review on BACI analysis will be carried out, helping us to design our own analysis.

If the time of impact is not available a change point analysis can be carried out. Change point analysis allows us to detect the time at which an impact has occurred if it is present within a series. A literature review of various change point methods will be carried out and a subset of the methods will be assessed. Three of these methods will be used on the real life data after conducting a simulation study to assess their performance.

Once the time of impact has been located, modelling techniques can be applied including a dummy variable which can account for the time at which an impact has occurred. This allows the model to 'jump' where an impact is present.

The approaches described above will be assessed and applied in this thesis on a variety of environmental contexts including a windfarm development and a global change in air quality with the purpose of locating and identify changepoints, determining whether the change in response is significant and then modelling the data appropriately.

1.2 What is an 'Impact' and types of impact

The potentially impacted data series may or may not contain a shift or change in the statistical structure of the data; if the series does contain a shift, we need to incorporate such an effect into our models. These changes can be referred to as 'discontinuities' or 'regime shifts'. The detection of discontinuities or regime shifts, are of great interest to those who wish to model time series data efficiently. A change in parameters within a series may affect how the series should be modelled as functions may need to be added to account for the change. Regime shifts are cases of inhomogeneity from one relatively stable state to another. The shifts in the series are not necessarily natural but can also be anthropogenic. For example the introduction of a new law or the building of a new structure may affect the time series variables by increasing or decreasing the mean for some given variable for a short or long period of time, change the slope of the trend, or cause the variable of interest to be more noisy after the change point.

The methods we wish to use depend on the design of the study. There are different ways of determining whether an impact has occurred and its magnitude, dependent on the data we have. It is therefore in our interests to determine:

- 1.) Can we detect a shift in model parameters after this intervention; do the parameters revert to their original values after a period of time?
- 2.) What is the magnitude of the shift at the point of intervention?
- 3.) Is there evidence of trends before and after this intervention; are these trends different?

The time series could potentially contain a combination of steps and trends. The steps may be permanent or temporary jumps in mean values and they could be abrupt or gradual changes. There may also be trends before and after this intervention, disrupted by the intervention itself. These regime shifts could therefore consist of:

- A shift in the mean from μ_1 to some other values $\mu_2 \dots \mu_n$
- A change in regression (slope)
- The standard deviation could shift from σ_1 to some other value
- A combination of all 3

Examples of such interventions are shown below in figure 1.1 showing the series schematically before and after intervention with no variability or seasonal trends. Trends that are present are assumed to be caused by the intervention – not by natural cyclic trends in the data.

The three plots on the left of figure 1.1 show shifts in which the series do not decrease back to the original value: The top left shows a simple abrupt shift in the mean value, there is no trend in this series the mean of the series only shifts after intervention.

The plot on the left in the centre shows an abrupt shift as well as an increasing linear trend. The plot on the bottom left also shows an abrupt shift, but a non-linear increasing trend.

The three plots on the right of figure 1.1 show series that revert back to their original mean value after a period of time. The top right plot shows a temporary change in mean which increases for one time unit then reverts abruptly to its original value. The middle right plot shows an abrupt change in mean which reverts back to its original value in a linear fashion, whereas the bottom right plot reverts back to its original value in a non-linear fashion.

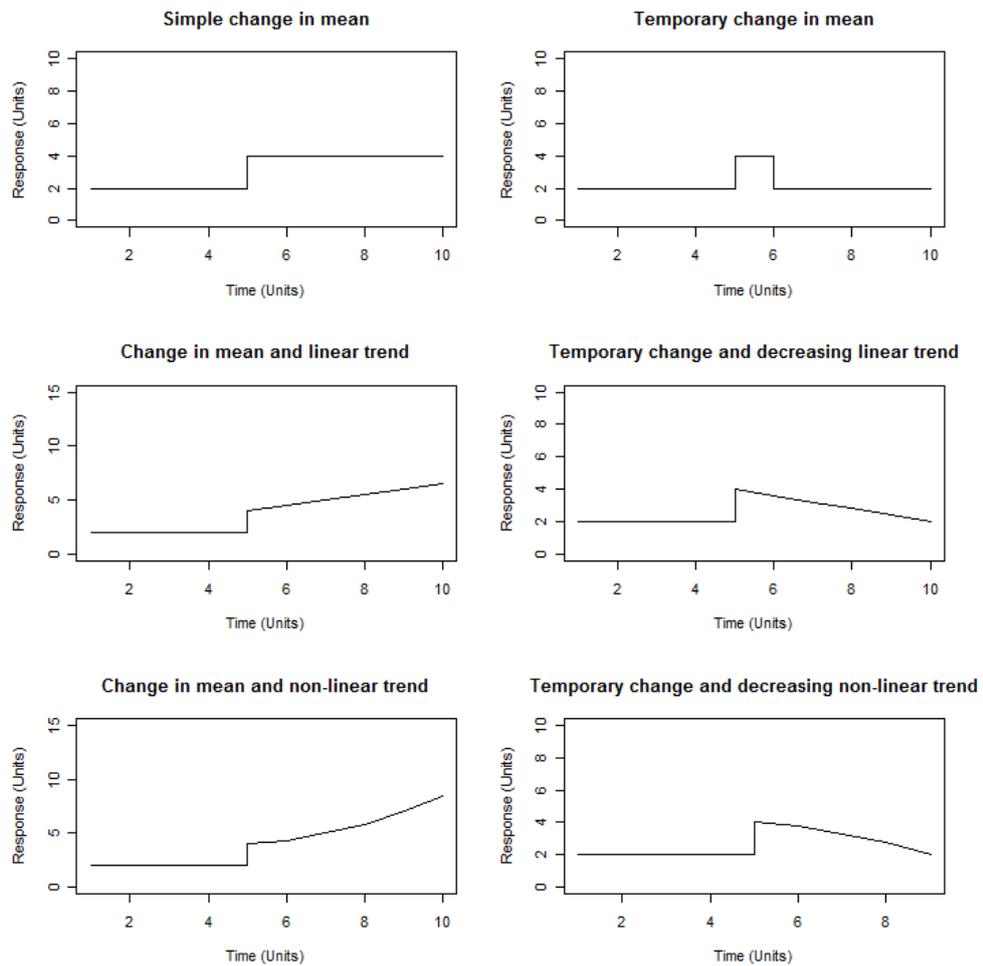


Figure 1.1: Examples of interventions

As specified before, shifts do not necessarily need to be abrupt but can be gradual. For example, a gradual increasing linear trend after time t would just be considered a linear trend attributed to the intervention. Furthermore just because a mean or trend is constantly increasing after intervention does not mean that it will always stay constant or increase, it may converge to its original value after a number of time units.

A further consideration is the amount of variability in the data. Shifts and trends caused by the intervention may be present however a large amount of variability may hide the effects of intervention. Furthermore as with all Impact assessments, we must determine whether before and after variances differ or are the same, Oaten (2001) addresses this problem in his paper 'Temporal and Spatial variation in environmental impact assessment'. It is also relevant to consider the problem of identifying whether a trend is actually caused by the intervention, or if it is natural. In longer time series it is more apparent which trends are natural and which are caused by intervention, however in short series or those that cannot be compared with control series this may prove problematic.

1.3 Application of techniques

1.3.1 Overview

This thesis will apply a variety of change point detection methods and modelling techniques to:

- 1.) European Monitoring and Evaluation Programme data on Sulphur Dioxide at various locations in Europe.
- 2.) Whitelee wind farm data on Total Phosphorus, Total Organic Carbon and Nitrate Oxide (Murray (2012)).

1.3.2 European Monitoring and Evaluation Programme (EMEP)

The EMEP measurement network monitors and evaluates concentrations of air pollutants for over 25 years at sites situated across Europe. Many datasets are available on the EMEP website which are freely accessible for non-commercial use including heavy metals, Persistent Organic Pollutants (POP's) particulate matter and Volatile Organic Compounds (VOC's) however within our analysis we will focus on an acidifying and eutrophying compound, SO₂ (Sulphur dioxide). There may be a presence of discontinuities in the time series due to an international convention which limited pollutants in Europe. By assessing this time series we can determine whether there are any statistical discontinuities within the series using a variety of methods. We will assess 2 datasets from different locations within the analysis, one from Great Britain (GB02) and one from Austria (AT02). Log SO₂ values will be used within the analysis, where the analysis will be carried out on around 30 years' worth of data. The data which has been used can be found on the following website: <http://www.emep.int/>. Locations and altitude of the 2 data sets are shown in table 1.1.

Table 1.1: Locations and altitudes of EMEP stations in Europe.

Code	Station name	Latitude	Longitude	Altitude
AT0002R	Illmitz	47 46 0 N	16 46 0 E	117
GB0002R	Eskdalemuir	55 18 47 N	3 12 15 W	243

The sampling frequency of the data shown below in figures 1.2 and 1.3 is daily. Much of the data are not available due to sampling problems and therefore has to be estimated using modelling techniques. Local linear regression with weights was chosen to estimate the missing data within each series. The two series available are shown below, green datapoints are real samples whereas red datapoints are estimated from the black line plus random variation where the variation is simulated from a normal distribution with variance roughly equal to that of the actual data (single imputation).

Daily log SO₂ over time can be seen below in figure 1.2 (red datapoints indicate simulated data). The variance looks constant over the plot however the mean curve changes around 1992, this could be because of a decrease in emissions around this time.

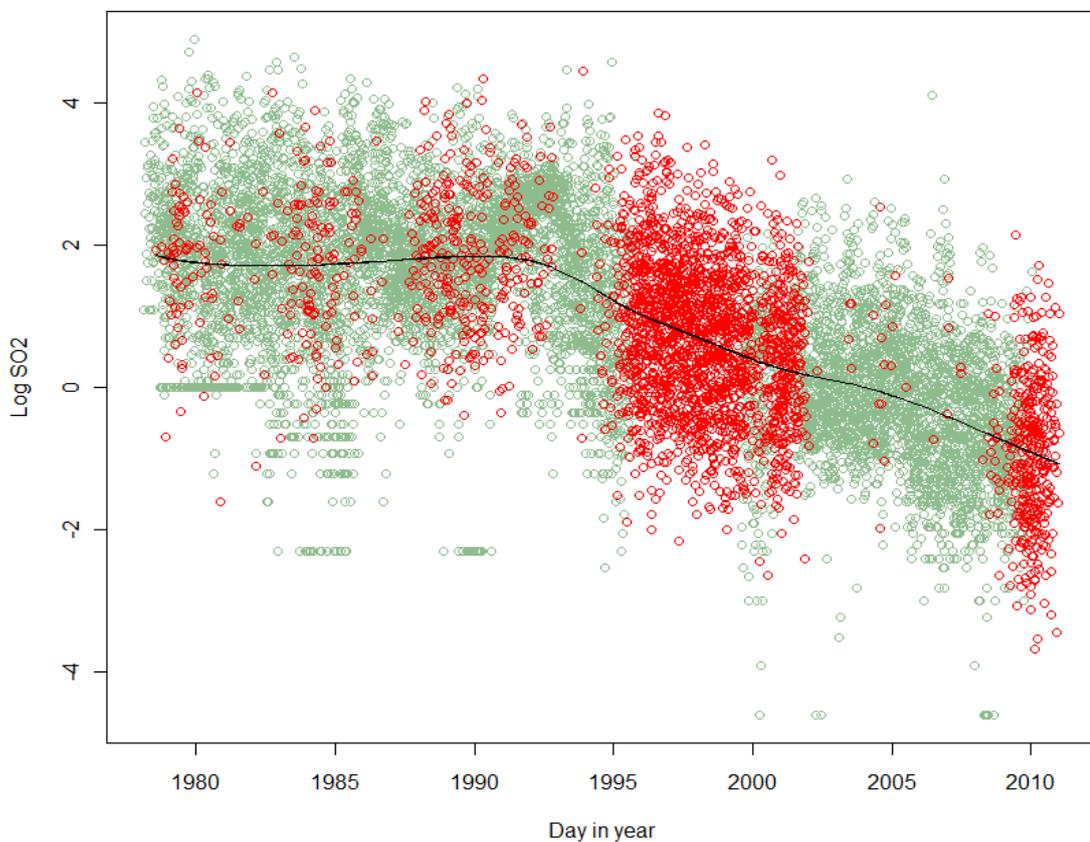


Figure 1.2: Time series of log SO₂ at AT02 (Superimposed estimated data)

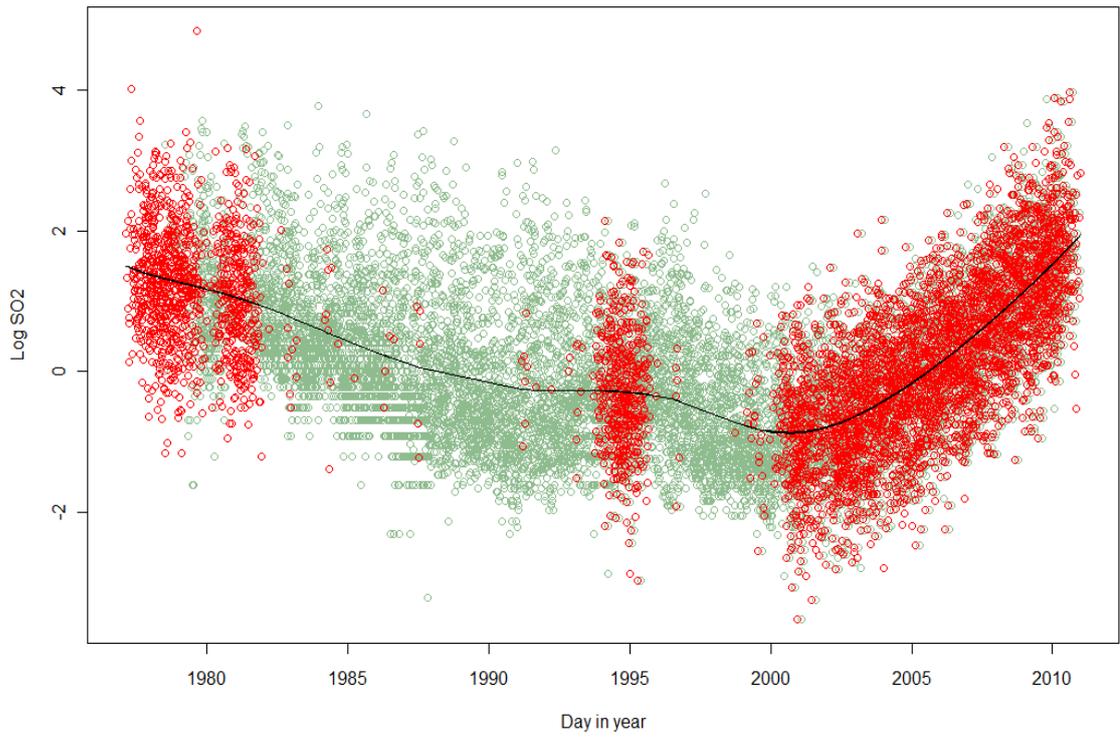


Figure 1.3: Time series of log SO2 at GB02 (Superimposed estimated data)

Figure 1.3 shows daily log SO2 at the GB02 station. Again, much of the data has been simulated as indicated by the red datapoints. There is evidence of a gradual decline until around 2000 when the mean curve then starts to increase.

1.3.3 Whitelee wind farm

Whitelee windfarm is a windfarm situated south of Glasgow and is the second largest in Europe. Whitelee wind farm's generating capacity is 322MW. It has 140 turbines, where each 65 meters high with a rotor diameter of 90m (a total height of 100 meters per turbine). The site is also primarily based on peat and wetlands, 1-7m of peat overlying 2-3m of glacial till over the basalt bedrock. The peat can be fluid just 0.5m down from the surface. Peat is a natural store of carbon and it is therefore of interest to determine if building on peat disrupts natural storage (www.whiteleewindfarm.com).

It is hypothesised that the disturbance of peat lands by developments of creation of roads, insertion of wind turbines and associated forestry may have an impact on the peat land structure and have an effect on the rate of decomposition of organic matter. The construction of these turbines in mid-2007 may have an effect on the transfer of carbon and nutrients from terrestrial to aquatic ecosystems.

Sampling is currently on-going by Anthony Waldron (The University of Edinburgh) but previous to 2011 samples were collected and analysed by Helen Murray (University of Glasgow).

The dataset itself consists of 8 variables across 11 sites giving around 88 time series in total (some series are missing). A list of the 8 variables (in mg/L for all apart from CaCO₃ for alkalinity) can be found in table 1.2.

Abreviation	Nutrient
TOC	Total organic carbon
DOC	Dissolved organic carbon
POC	Partial organic carbon
TP	Total phosphorus
SRP	Soluble reactive phosphorus
NO ₃	Nitrate NO ₃
NO ₂	Nitrate NO ₂
Alkalinity	Alkalinity

Table 1.2: Whitelee time series variables

Analysis will only be carried out on TP, TOC and NO₂. The reason for this is that these series have the least missing data. Estimating data within the other series would leave us with more estimated data than real data. Stations WL13, WL14 and WL1 were chosen since it is expected WL13 will be effected most, followed by WL14 and finally WL1. However, this is only to determine where a changepoint lies, the overall modelling of the data will include all 11 sites.

As turbines and roads started to get built around mid-2007, we expect there will possibly be a change point at or after this time. However the type of change is unknown, it could be a change in mean, trend, variance or a combination of the three. Furthermore the time series may revert to its original state after a number of years which may either be considered a second discontinuity whether it is abrupt or gradual. A map of Whitelees sites can be seen in figure 2.4.

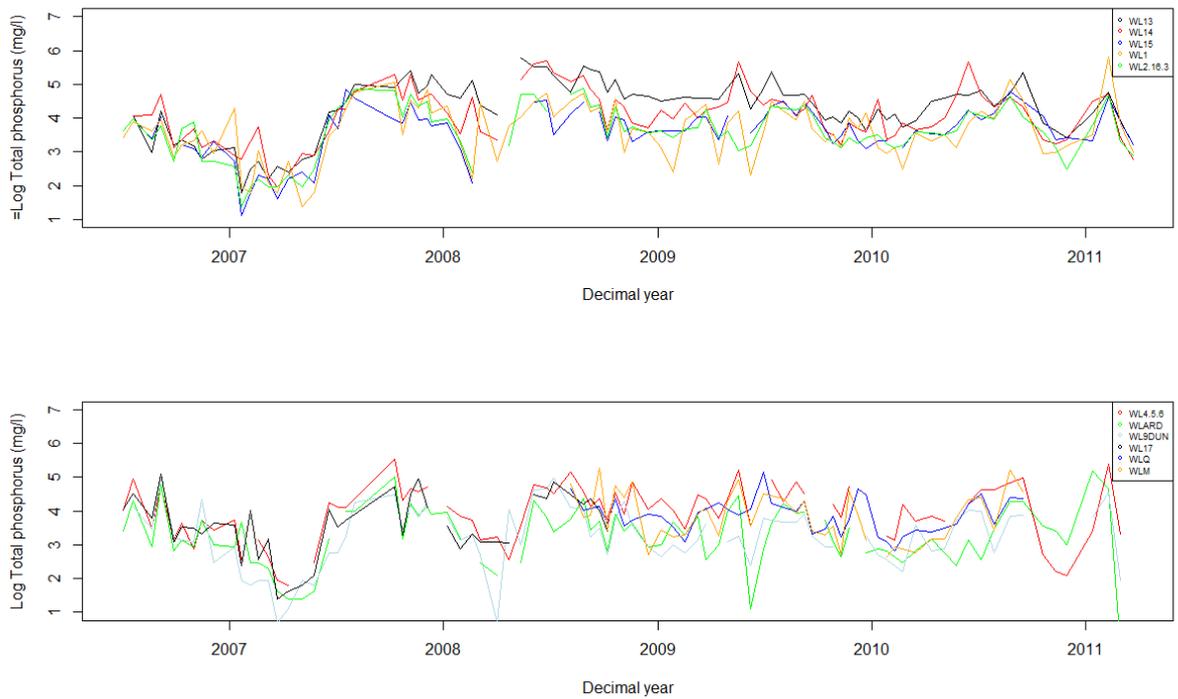


Figure 1.4 a, b: Time series of log Total Phosphorus

Figure 1.4 above shows log Total Phosphorus over time at the 11 sites, split into two plots a and b. Around mid-2007 there does seem to be a jump in mean value however it is not clear whether the variance or trend parameters change after this point

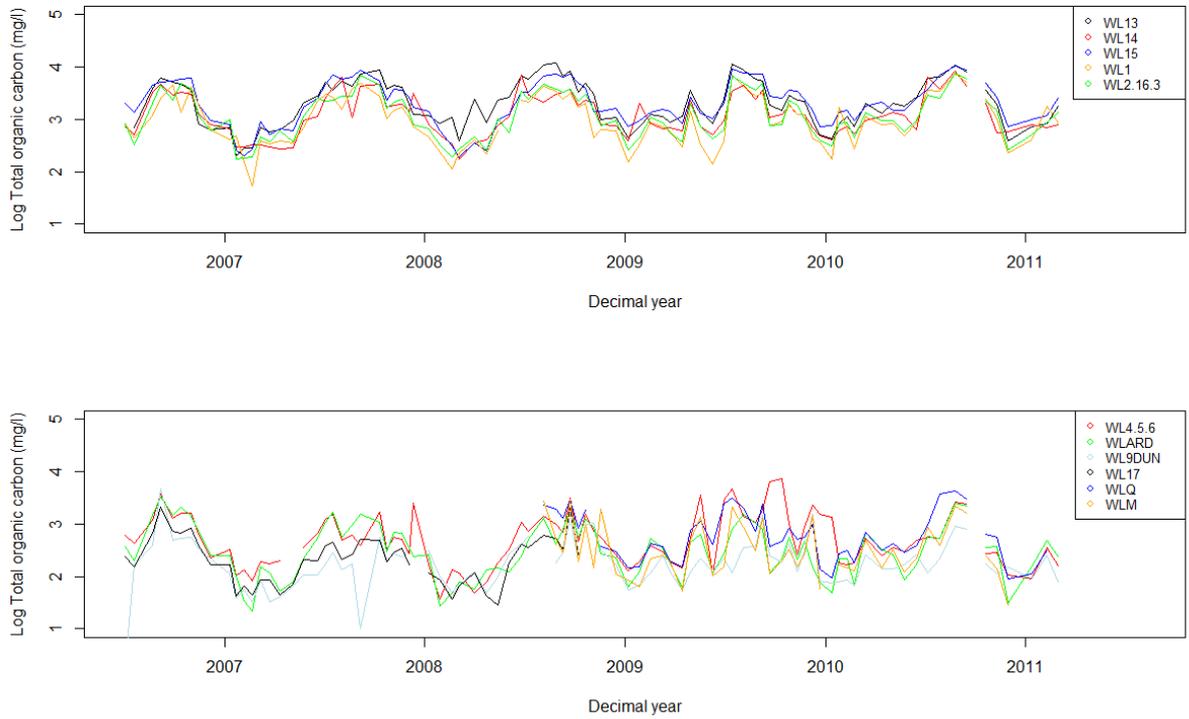
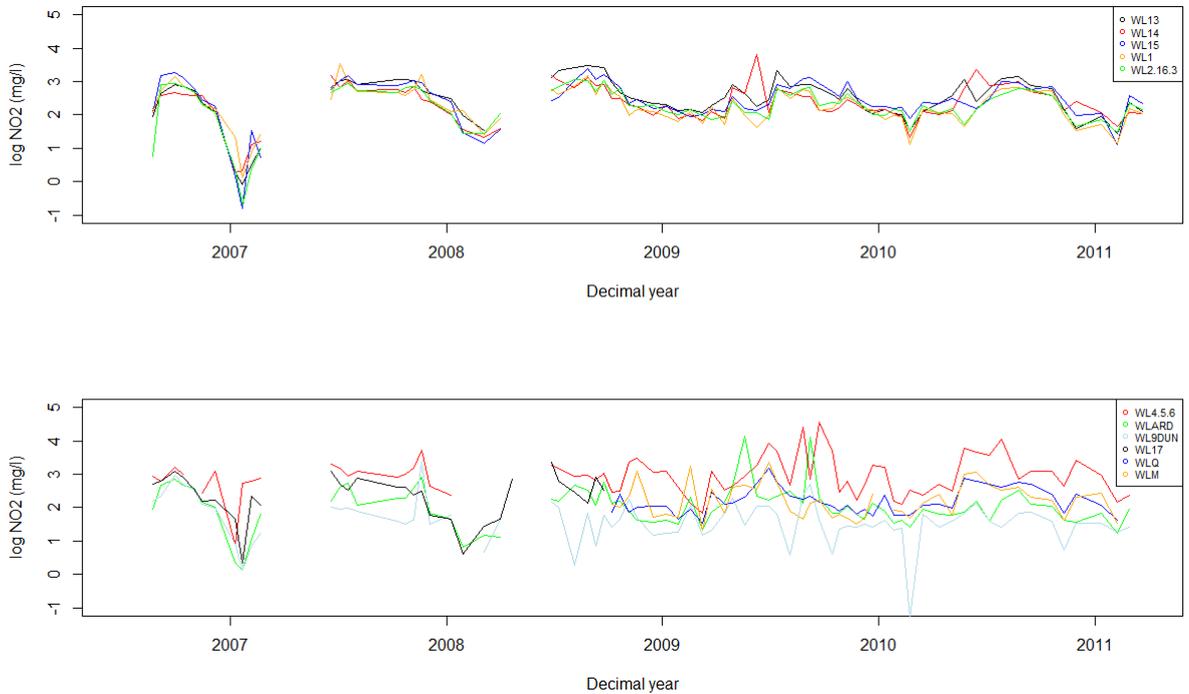


Figure 1.5 a, b: Time series of log Total Organic Carbon

Figure 1.5 shows log Total Organic Carbon over time, at the 11 sites, split into plots a and b. It is not clear from these plots whether any change has occurred around mid-2007 however there is a clear seasonal trend within the series and the variation between sites is less than that of Total Phosphorus.



Figures 1.6 a, b: Time series of log Total Nitrate Oxide

Figures 1.6a and 1.6b shows log Nitrate Oxide over time again at the 11 sites. Much of the data are missing within these series and therefore local linear regression will be used to estimate the missing values. Again, it is not clear whether there is a change in structure after mid-2007 however a seasonal cycle does seem present.

In all three variables data are missing however this can be easily estimated using local linear regression with weights.

1.4 Aim of this work

The aim of this thesis is to evaluate the statistical tools available for impact assessment and discontinuity detection, applied within an environmental setting. Current and past methods will be compared and contrasted through a literature review and a simulation study. The thesis will also consider a variety of statistical testing and modelling approaches and see how they can be used within an EIA framework.

Firstly within chapter 2, Before After Control Impact (BACI) will be applied. A literature review summarising the history and adaptations of BACI will be carried out and BACI analysis will be carried out on some of the data available.

Secondly, changepoint analysis techniques will then be reviewed and applied to the data within chapter 3. Following the identification of changepoints within the series, the methods which have been evaluated previously can then be applied to EMEP data and a number of variables from Whitelee wind farm.

In chapter 4, once these variables are assessed for discontinuities, we can then model them using Generalised Additive Modelling techniques.

Finally, chapter 5 will summarise the results and findings of the various approaches used throughout the thesis.

Chapter 2

Before After Control Impact (BACI) design

2.1 An introduction to BACI

A very simple and easy to interpret way of determining whether an event had an 'impact' on the environment is to carry out a Before After Control Impact (BACI) study.

Not only does BACI analysis allow us to evaluate whether or not an event has changed the environment, it also allows us to estimate the magnitude of its effects.

The theory behind BACI is that we have data before, during and after the 'event'. We also have control sites and 'impact sites'. These control sites are unaffected by the 'event' as they are either outside the area which has been affected or are under similar conditions to the impact sites but at a different location. The control sites allow us to compare an unaffected series with an affected one and would therefore allow us to attribute any change in the impact site that is not observed within the control site to the 'event'. Combined with the availability of data before during and after the event, the design allows us to fully understand whether the 'event' has made a substantial impact on the variables in question but the time at which the 'event' has occurred must be known.

Adaptions of BACI have also been used. Some examples of this include studies in which data are only available after the event, no control sites are available or only one impact site is available with no controls.

The following section discusses the possible designs that can arise, and a literature review discussing the different approaches that have been used.

2.2 Literature review on BACI analysis

The statistical model used within BACI is based entirely on the design. BACI analysis was firstly used by Green (1979) and has since then been further adapted by a number of authors to address various issues involving the number of control sites, number of samples and sampling times. Green's original design was basic, with one single sample before and one single sample after impact at both a control site and an impact site. Bernstein & Zalinski (1983) adapted this design to include multiple samples before and after the impact, sampling at paired times for both control and impact sites. Stewart-Oaten et al. (1986) also adopted this design but also mentioned that to avoid coincidences with natural cycles random sampling must be used (sometimes called BACIP, Osenberg et al. (1994)). Stewart-Oaten (1986) also considered how to analyse the data using t tests to compare means before and after the potential impact. Further adaptations including the same design as Bernstein & Zalinski's but unpaired samples have been suggested.

Dealing with spatial and temporal confounding has been addressed by Bernstein & Zalinski (1983) and Oaten et al (1986). Beyond BACI designs developed by Underwood (1991) has led to advances with association to human activities – designs which use multiple controls and are analysed with asymmetrical analysis of variance because of the presence of a single disturbed location. An example of the use of both univariate and multivariate asymmetrical analyses can be found in Terlizzi et al. (2005) with application to Mediterranean sub tidal sessile assemblages.

The number of impact sites and whether control sites are available will determine how the analysis is conducted. The sites in question may all have been affected equally or a proportion of these may have been affected by the intervention less than others, or not at all. These sites that are not affected are called control sites, and can be used to compare natural unaltered trends with those that have been impacted. The presence of these control sites determine which design we must consider and which methodology can be used. Within this section of the thesis, single series (impact) and multiple series (including only impact sites and both impact and control sites) will be assessed and the ways in which each are dealt with will be discussed.

The analyst may also want to consider the type of impact that the activity in question has. Disturbances in the environment can be categorised into two separate areas: Pulse and Press disturbances.

Pulses are short term episodes of disturbance which are then quickly removed, an example of this may include oil spills which are quickly removed but for the time that the spill is present, the impact is large. This is very different from a Press disturbance which is on-going and constant, for example, a daily discharge of a chemical from a power plant for example.

The two types of disturbances may be very different in terms of the potential effects on the variable in question (consult Bender et al. (1984) and Underwood (1989)) and therefore the resulting 'impact' may have to be treated differently since the magnitude of effects may be less apparent within Press disturbances and may be smoother over time.

In the following sections, each of these different settings are considered in greater detail.

2.2.1 Single series

The simplest method of impact assessment is where we have one single time series. This time series is an impact series and is not paired with a control site, and is simply a series of observations that contain the intervention. It is therefore a single 'impact series', one that contains an impact where the time of impact is known. We have data before, during and after this intervention hence the name Before-After design. This simple design does not account for any temporal variability or any other causes that could affect the series (Underwood (1992)) as it does not have a paired control site, therefore any difference after the intervention point t must be attributed to the intervention. Hurlbert (1984) states that a design which is used on single site with replicates before and after the impact are 'pseudoreplicated', such that any change in structure may be due to a variety of causes and not just the 'impact' in question and therefore to counter this several control and impact sites must be assessed.

An example of this design can be seen in figure 2.1 below, with an intervention at $t = 30$ (black line is the estimated model without variation).

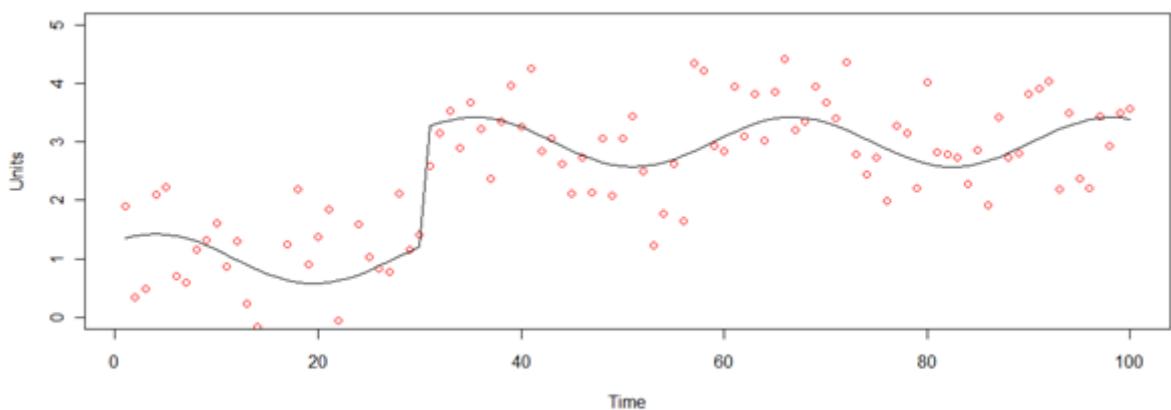


Figure 2.1: Single simulated series with a change in mean at $t = 30$

A problem with this design is that environmental trends are common in time series and an observed effect may not be due to the intervention. Furthermore, a lack of an observed effect may actually indicate an impact since a natural trend may have been interrupted by the intervention.

For a single series, simple tests can be carried out as a means of formal analysis. The simplest method given independent and normally distributed data is the t-test which simply calculates whether there is a significant difference in means. To use the 'two sample t-test', data must be independent, normally distributed and the groups must have equal variance (R.A Fisher (1925)). T-tests can be carried out where we have unequal sample sizes and variances, but formulas to calculate t are denoted differently because of assumption violation; for example the Welch two sample t-test (Welch (1938)) is a test used for unequal variances. Equal variance can be formally tested with tests such as Levene's test or the Brown–Forsythe test. Non-parametric tests such as the Mann Whitney U or Wilcoxon signed rank can also be used when assumptions are violated as mentioned by Ruxton (2006). Oaten (1986) states that to simply conduct these tests of before versus after is not appropriate since cyclical and long term variation is likely to occur.

To allow us to detect significant changes in mean and trend (omnibus test), a linear model can be fitted to the data and an ANOVA carried out to determine significant factors. This method is preferred where we want to determine whether there are significant differences between more than two groups, for example, we may have multiple intervention points within the analysis which would result in more than two separate groups of data. The reason this method is preferred is because by conducting a multiple comparisons procedure, we would need to carry out $\frac{n(n-1)}{2}$ t-tests (n=number of groups) to obtain the same results.

To allow for the intervention point, an indicator variable must be set up to indicate the changepoint. Additivity is an assumption of BACI modelling and Oaten (1986) suggests using Tukeys test (an approach used within a two way ANOVA to assess whether the factor variables are additively related to the expected value of the response variable) to formally test this assumption. This problem is also discussed in Oaten (2002), Oaten (2003) and Smith (1993).

A simple model for BA design as specified in the Encyclopedia of Environmetrics would be as follows in equation 2.1:

$$X_{ik} = \mu + \alpha_i + \tau_{k(i)} + e_{ik} \quad (2.1)$$

- i is a binary indicator where $i = B(\text{before}), A(\text{after})$
- μ is the overall mean
- α_i is the effect of the period
- $\tau_{k(i)}$ represents time within period ($k=1,2,3\dots t_A$ for $i = A, k = 1,2,3 \dots t_B$ for $i = B$)
- e_{ik} is the error term, normally distributed with constant variance.

This model could then be used within an ANOVA. From the ANOVA, the p-values in the output will allow us to determine whether each term is significant or not. These p-values were calculated by a series of F tests based on mean squares. Note that the errors are assumed to be normally distributed for this model.

It is important to note that neither t-tests nor ANOVA procedures account for temporal variability or serial correlation in the data. This does not apply for the ANOVA if the natural trends are taken into account within the model, however as we have stated before any change at time point t must be attributed to the intervention as we have no paired control series in BA design.

The standard ANOVA table for this model can be seen below within table 2.1 where mean squares are denoted as MS and defined as SS/df .

Source	SS	Df	F
Period: Before-After	SS_{BA}	1	MS_{BA}/MS_{times}
Sampling times	SS_{times}	$t_B + t_A - 2$	
Total	SS_{Total}	$t_B + t_A - 1$	

Table 2.1: ANOVA table for BA design (equation 2.1)

2.2.2 Multiple series

A variation on Before-After design is to sample from a number of sites rather than observe a single site over time. This is ultimately the same as Before-After apart from the analysis is carried out on M sites rather than 1 site, again each site is sampled in both periods of time so that we have $N \cdot (t_B + t_A)$ observations rather than $1 \cdot (t_B + t_A)$ observations. As noted in the Encyclopedia of Environmetrics a practical problem is whether to view the sites as replicates or subsamples – If the focus is on the extent of the impact then the sites are selected according to a preset scheme and viewed as replicates. If the sites are selected at random then they are viewed as subsamples.

There are a few differences with BA with multiple sites design compared with a BA design. Sites are not all necessarily impact sites; a proportion of the N sites could be control sites, allowing us to compare the control sites with the impact sites. If this is the case, the design is technically known as BACI (Before After Impact Control) design as we can compare controls with impact sites. Furthermore, the fact that there is more than one site lets us assume that the resulting change in the series is attributed to the impact rather than coincidence since each site may have been impacted to differing extents. A problem with this is as noted by Underwood (1992) is that similar changes at a reference site may actually be observed by chance, meaning the impact would not be detected while using a reference series. Underwood (1992, 1992, 1993, 1994) adapted BACI design further and called his analysis 'Beyond BACI', a design using multiple control sites to allow researchers to distinguish natural from impact induced variability at different temporal scales, however this design may be hard to implement due to the fact multiple controls may not always be available. The Beyond BACI approach using multiple controls applied in an environmental setting can be found within Musco et al. 2009, Knott et al. 2009 and Queiroz et al. 2006 applied to sewage dumping, water pollution and oil spills respectively. An example where two control sites and one impact series is present can be seen within figure 2.2 below.

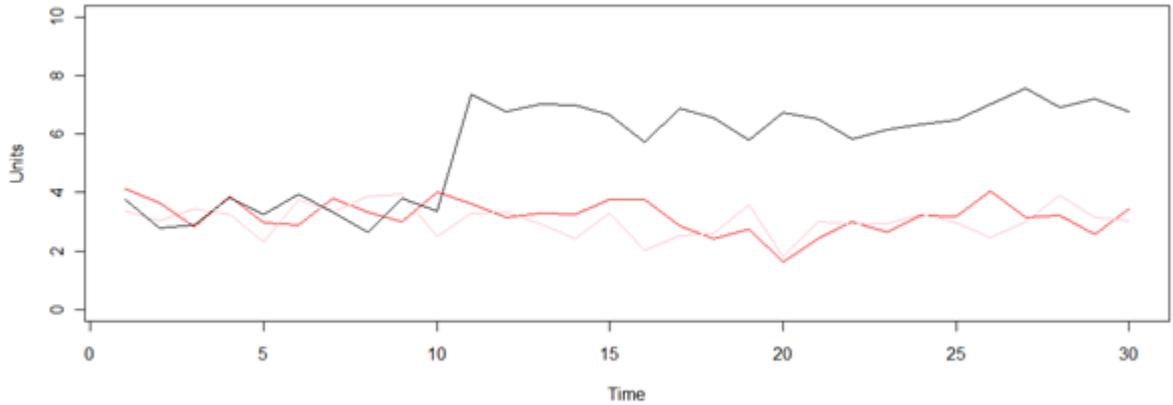


Figure 2.2: Multiple simulated series with a change in mean at $t = 30$

Red and pink lines represent the control series whereas the black line represents the impact series.

If no control sites are present and only impact sites are present with differing magnitudes of change, a plot of all series may look similar to figure 2.3 shown below:

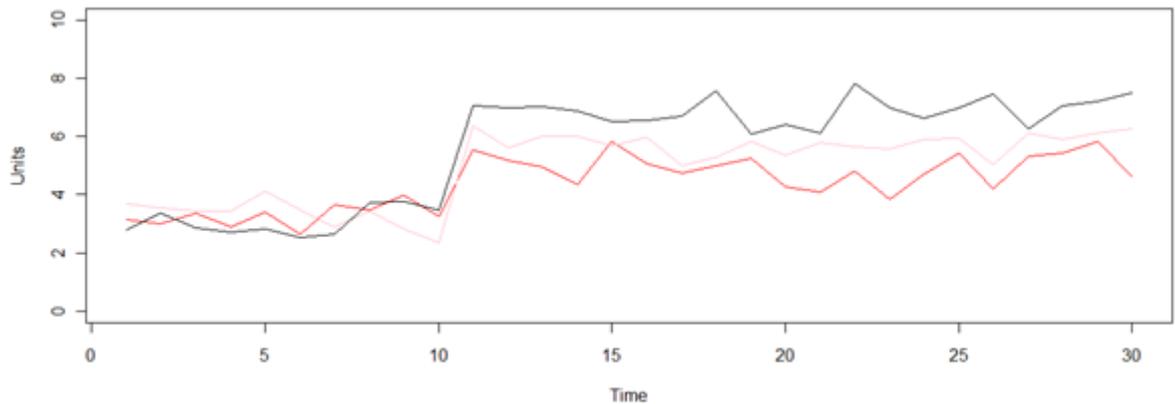


Figure 2.3: Multiple simulated series with a change in mean at $t = 30$

When considering the control sites which will be used, the design must consider the location of this site if variation is significantly different between locations. It must be located in an area with the same temporal and spatial variability as the impact sites but outside of the potential impact site as mentioned by Oaten (1986). This can prove a problem, as there is almost always natural variability in space as well as time (Underwood (1992)). Another problem may be that similar changes at a reference site may actually be observed by chance, meaning the impact would not be detected while using a reference series. These measurements are not necessarily paired in time either; if they are we can take the

difference series to ultimately remove the seasonal variation (assuming spatial variation is constant). Furthermore these sites are not necessarily randomly selected – this introduces potential analysis bias as the sites chosen as the impact sites (which are most likely chosen for a reason) could potentially harbour systematic differences in their environmental make up. Hurlbert (1984) criticises the analysis of designs without randomisation of impact and control sites because of this problem and as mentioned in the previous section, ‘pseudoreplication’ can become an issue if only one control and one impact site is available due to the fact that the two sites may have two different temporal trajectories in space and therefore more than one control site may be needed.

In theory truly randomising impact and control sites is hardly possible in a spatial context. In almost all cases the purpose in an EIA report is to assess the environmental impact of a man made structure or project – it is therefore not in the interests of those conducting the project to ‘randomly’ select an impact site as they will have probably chosen that site for a reason.

A variety of models can now be fitted to the data, each including the indicator parameter which identifies the point in time that an impact has been found from previous analysis.

As before, simple t-tests and non-parametric methods to determine whether population means differ can be used. However, using these methods for BACI and especially BACI paired designs are much better indicators of an overall effect since we can calculate t statistics for a difference series (impact-control) rather than only on the impact series.

As stated before, we cannot determine whether there is a presence of trends with these tests.

Linear models can again be fitted to the data and the appropriate ANOVA carried out however, if an impact site is available a difference series to remove seasonal trend can be calculated, a parameter to deduce whether there is significant trend in the data can be included in the model and assessed using the ANOVA. In theory by removing all seasonal variation, all that is left in the series is the effects of interaction which can be quantified and tested to check for significance of these trends.

The Encyclopedia of Environmetrics states that a model that can be used for BACI designs is as follows. Our response is predicted by an overall mean, the effect of the period (before and after), time within period, effect of location (impact of control), the interaction between period and location and finally an error term.

A model for the BACI design as specified would be as in equation 2.2:

$$X_{ijk} = \mu + \alpha_i + \tau_{k(i)} + \beta_j + (\alpha\beta)_{ij} + e_{ijk} \quad (2.2)$$

- i is a binary indicator where $i=B$ (before), A (after)
- j is a binary indicator where $j=C$ (control), I (impact)
- μ is the overall mean
- α_i is the effect of the period
- $\tau_{k(i)}$ represents time within period ($k=1,2,3\dots t_A$ for $i=A$, $k=1,2,3\dots t_B$ for $i=B$)
- β_j is the effect of location
- $(\alpha\beta)_{ij}$ is our interaction between period and location
- e_{ijk} is our error term, normally distributed with constant variance.

The appropriate ANOVA can then be used to determine whether there are differences in model trajectories between impact and control sites as well as before and after intervention. Again, p-values indicate whether each term is significant.

Within all of the above scenarios sampling times should be selected at random. This is so that natural cycles can be accounted for which may confound the analysis if the sampling times are fixed and far apart. In terms of including terms for seasonal patterns, harmonics can be used relating the response variable to time (Steward-Oaten et al. (1986) and Smith (2002)). The ANOVA table for BACI design can be seen in table 2.2.

Source	SS	df	F
Period: Before-After	SS_{BA}	1	
Location: Control- Impact	SS_{CI}	1	
Interaction BA x CI	SS_{BACI}	1	MS_{BACI}/MS_E
Error	SS_E	$N - 4$	
Total	SS_{Total}	$N - 1$	

Table 2.2: ANOVA table for BACI design (equation 2.2)

To summarise, if we have BA design with multiple sites we must consider:

- 1.) Do we have control sites, if so how many?
- 2.) Are the impact and control sites paired in time?
- 3.) Are the impact and control sites randomised?

These factors must be considered as the methods used to analyse the data series are dependent on them. Essentially methods for different designs will be similar however the way data are treated and the way results are interpreted will be different.

2.3 Application of BACI

2.3.1 Overview

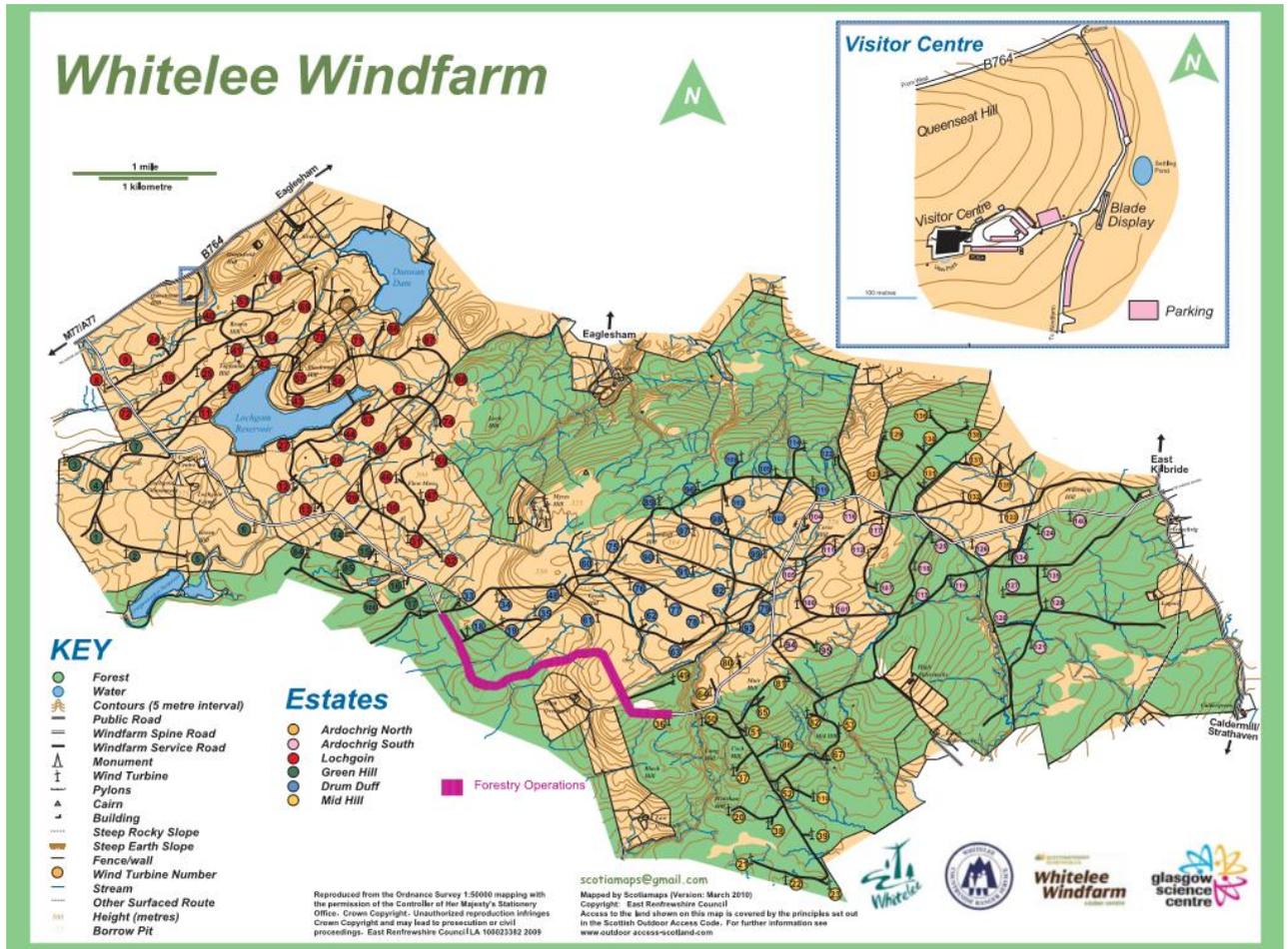


Figure 2.4: Map of Whitelee windfarm

Our investigation at Whitelee can be considered as an application of BACI design. The BACI framework here is used to evaluate changes in the determinants Total Phosphorus, Total Organic Carbon and Nitrate Oxide following the start of deforesting in the area, the insertion of wind turbines and inclusion of roads around Whitelee. A map of Whitelee wind farm can be seen in figure 2.4 which includes positions of wind turbines and the connecting roads. Determining whether a site should be regarded as an impact or control site is not simple. Control sites should be chosen as those which are similar to impact sites but are not likely to be affected by intervention. We will use 'percentage of deforested area' around

each catchment area as an indicator of whether we can include each site as an impact or control site. Table 2.3 below shows the percentage of deforested area at each site (note that no data are available for sites WLM or WLQ):

	Disturbance table: Percentage of catchment deforested									
Site	WL13	WL14	WL15	WL1	WL2.16.3	WL4.5.6	WL9A	WL9D	WL17	WL17U
% Deforested	12.2	10.3	3.5	10.7	13.2	0.8	0.5	2.3	2.0	3.9

Table 2.3: Disturbance table (Provided by Helen Murray, Glasgow University)

At each of the 11 sites, samples were taken roughly every three weeks. We will choose our control sites as those with “Percentage of catchment deforested” less than or equal to 2%, which means WL4.5.6, WL9A and WL17 will be chosen as our control sites.

The first analysis is based on differences between the control and impact sites paired by sampling date. The objective is to determine if the mean difference between impact and control sites have changed coincident with disturbance.

To obtain an initial impression, a simple before- after comparison of means will be carried out. This allows us to determine the overall mean difference at each site and determine whether the difference between before and after values at control sites are smaller however this does not take into account any change in trend or seasonal pattern at the change point.

To build a our model, we need to have a term for the trend, a term to account for seasonal pattern, a term to account for the changepoint, a term to distinguish between control and impact sites, an interaction between the type of site and the term accounting for the changepoint and finally an error term.

The general form of our statistical model which will be used within our BACI analysis will be in the following form:

Response = y_i Indicator (Factor) = I_{ik}
Time = t_i Control/Impact(Factor) = CI_{ir}
Month (Factor) = M_{ij} Error = e_{ijk_r}

$$y_{ijk_r} = \mu + \beta t_i + M_{ij} + I_{ik} + CI_{ir} + I_{ik}:CI_{ir} + e_{ijk_r} \quad (2.3)$$

Where our sites are indicated as $j = 1, \dots, 11$

Our observations are indicated as $i = 1, \dots, n$

Before or After intervention is indicated by $k = 0, 1$

Impact and Control sites are indicated by $r = 'Impact', 'Control'$

μ and β are our overall mean and slope respectively.

Our observed errors e_{ijk_r} are assumed to be normally distributed and independent with zero mean and constant variance.

Our parameter Indicator (I_{ik}) has two levels, 0 and 1 ($k = 0, 1$). The row vector is therefore in the following form:

$$I_{ik} = [0, \dots, 0, 1, \dots, 1]$$

Where the changepoint is indicated by the change in value from 0 to 1.

The parameter Control/Impact (CI_{ir}) is a row vector with two levels I (impact) and C (control) ($r = I, C$) and is in the following form:

$$CI_{ir} = [C, \dots, C, I, \dots, I]$$

Which indicates whether a site is a control site or an impact site.

The interaction $I_{ik}:CI_{ir}$ allows us to determine whether there is a difference in the change of mean before and after the changepoint between impact and control sites.

2.3.2 Total Phosphorus

2.3.2.1 Exploratory analysis (TP)

An initial impression of the data can be obtained by comparing mean values before and after the intervention in mid-2007.

Shown below in figure 2.3 are 11 boxplots, within each there is one box-and-whisker for before intervention and one after intervention. Table 2.4 also shows the mean before intervention and the mean after at each site, as well as the absolute difference between the two values. Absolute differences in bold are those sites that have a % deforested of less than 2%.

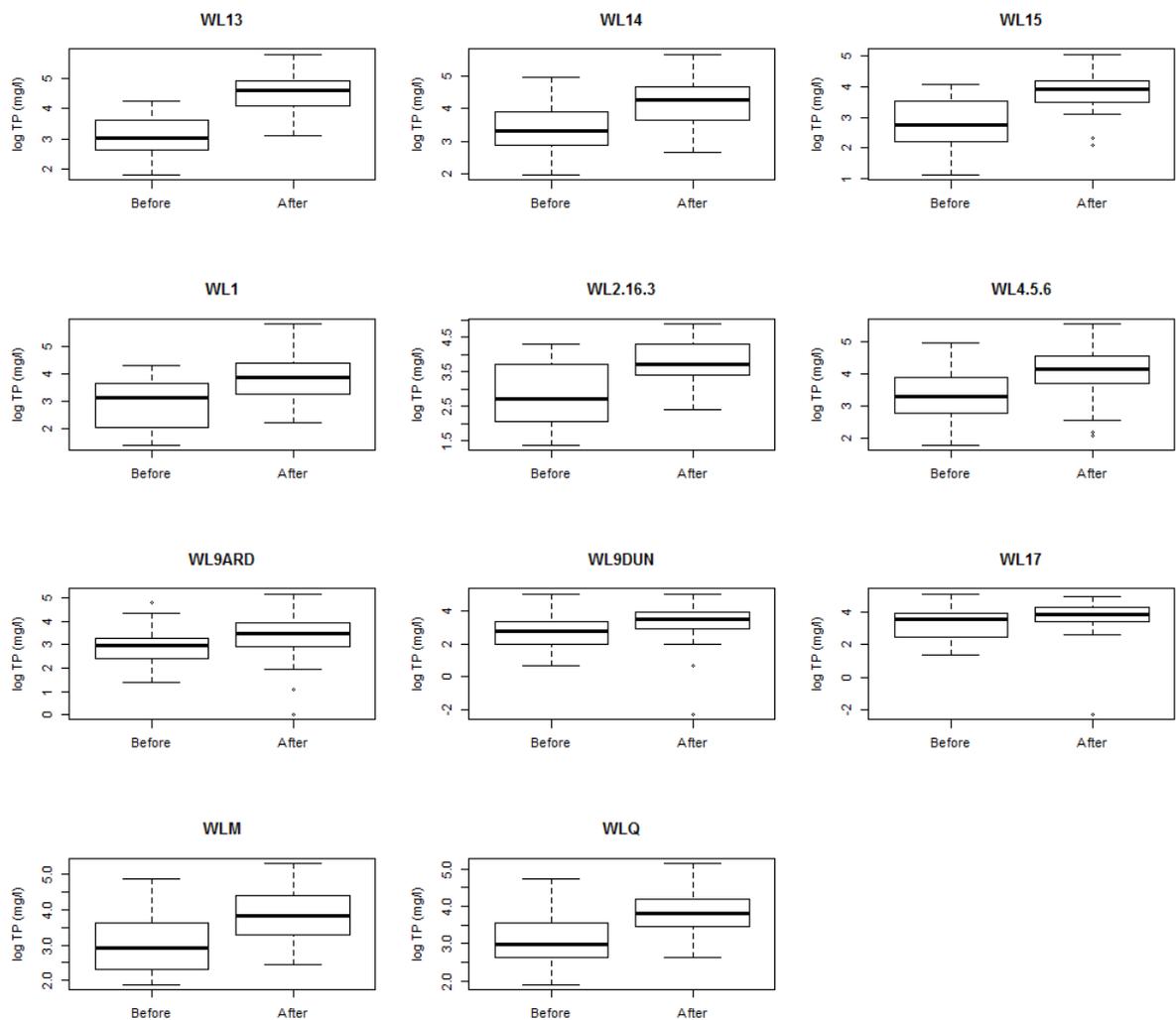


Figure 2.5: Boxplots of before-after at each site

	WL13	WL14	WL15	WL1	WL2.16.3	WL4.5.6	WL9ARD	WL9DUN	WL17	WLM	WLQ
Before	3.06	3.39	2.86	2.94	2.87	3.35	2.84	2.74	3.24	3.11	3.19
After	4.54	4.23	3.85	3.82	3.81	4.03	3.33	3.43	3.70	3.87	3.85
Absolute difference	1.48	0.84	0.99	0.88	0.94	0.68	0.49	0.69	0.46	0.76	0.66

Table 2.4: Table of means and mean differences

We can note that from both figure 2.3 and table 2.4, the three smallest differences between before and after mean values are at sites WL17, WL9ARD and WL4.5.6 in order of magnitude. Referring back to the disturbance table, these three sites have % of catchment deforested of 2.0, 0.5 and 0.8 respectively – the three smallest percentages.

The largest differences were found at sites WL13, WL15 and WL2.16.3, with respective % catchment deforested of 12.2, 3.5 and 13.2.

This may indicate that the sites with the smallest percentage of deforested area actually have the smallest absolute mean difference in terms of log TP, allowing us to use them as control sites.

2.3.2.2 BACI Analysis (TP)

The BACI analysis conducted below consisted of using a model including the terms Decimal year, Month, Indicator (Before or After impact), Control/Impact (whether the site is a control group or an impact group) and an interaction between Indicator and Control/Impact as shown in equation 2.3.

Analysis of Variance table					
Term	df	Sum Sq	Mean Sq	F-value	Pr(>F)
Time	1	21.02	21.020	40.7270	2.666e-10
Month	11	153.63	13.966	27.0590	< 2.2e-16
Indicator	1	91.82	91.824	177.9097	< 2.2e-16
Control/Impact	1	3.03	3.026	5.8626	0.01564
Interaction term	1	2.30	2.302	4.4600	0.03494

Table 2.5: BACI Analysis ANOVA (TP)

Firstly, linear model assumptions must be checked for the model above. We can assume constant variance as residual values are roughly equally distributed as shown in figure 2.6. Normality of residuals can also be assumed from figure 2.7 as the histogram shows that residuals do follow a rough normal distribution however there is a slight tail, this is because of two outlying values.

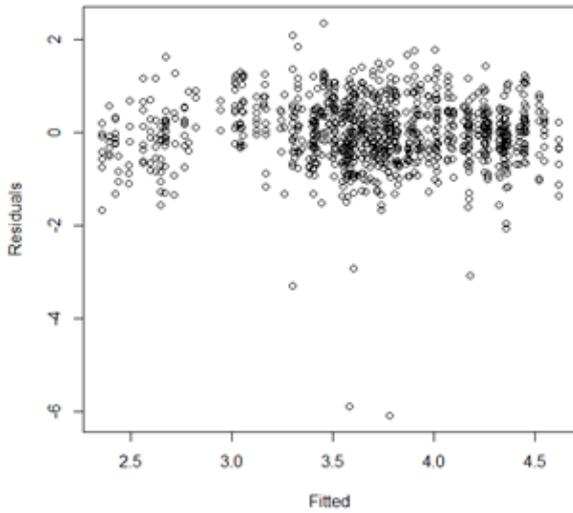


Figure 2.6: Residuals vs. Fitted values

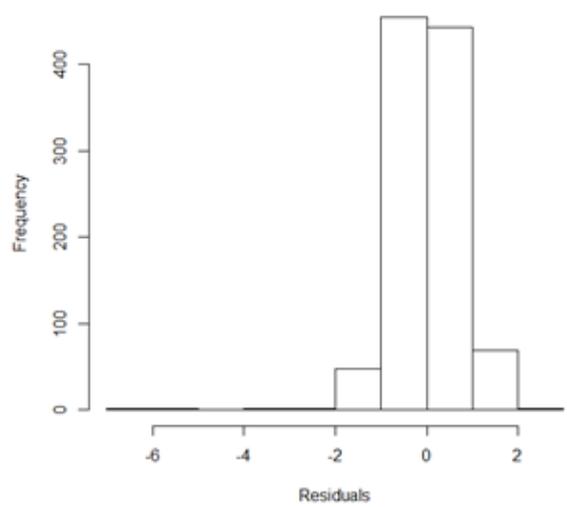


Figure 2.7: Histogram of residuals

From table 2.5, we can say that all terms within our model are significant with p-values less than 0.05. Decimal year is significant, indicating that there is an overall trend. Month is significant which means there is a significant seasonal pattern. Our indicator function is significant; this means there is an overall mean difference between before and after the intervention.

Our parameter Control/Impact is also significant which means that control and impact sites are structurally different (baseline = control). The interaction between Indicator and Control/Impact is also significant.

Intercept	Time	Indicator (after)	M2	M3	M4	M5	M6
295.193	-0.146	1.337	-0.029	-0.230	-0.632	0.141	0.699
M7	M8	M9	M10	M11	M12	Con/Imp (Impact)	Interaction
0.843	0.824	0.785	0.334	0.282	0.025	0.187	-0.473

Table 2.6: Table of coefficients (TP)

We can now comment on the magnitude of each parameters effects on log TP as shown in table 2.6. For each unit increase in Decimal year, log TP decreases by 0.14. If a site is an impact site, log TP increases by 0.19. After the changepoint, indicated by the Indicator function, log TP increases by 1.337 on average.

From table 2.5, we can conclude that there is structural difference between control groups and impact groups and there is also an overall mean difference before and after the proposed impact location in time.

2.3.3 Total Organic Carbon

2.3.3.1 Exploratory analysis (TOC)

Again by comparing boxplots below and the means within table 2.4 of before and after the intervention at mid-2007 we can obtain an initial impression to whether changes can be detected and whether larger changes in mean are observed within the impact sites.

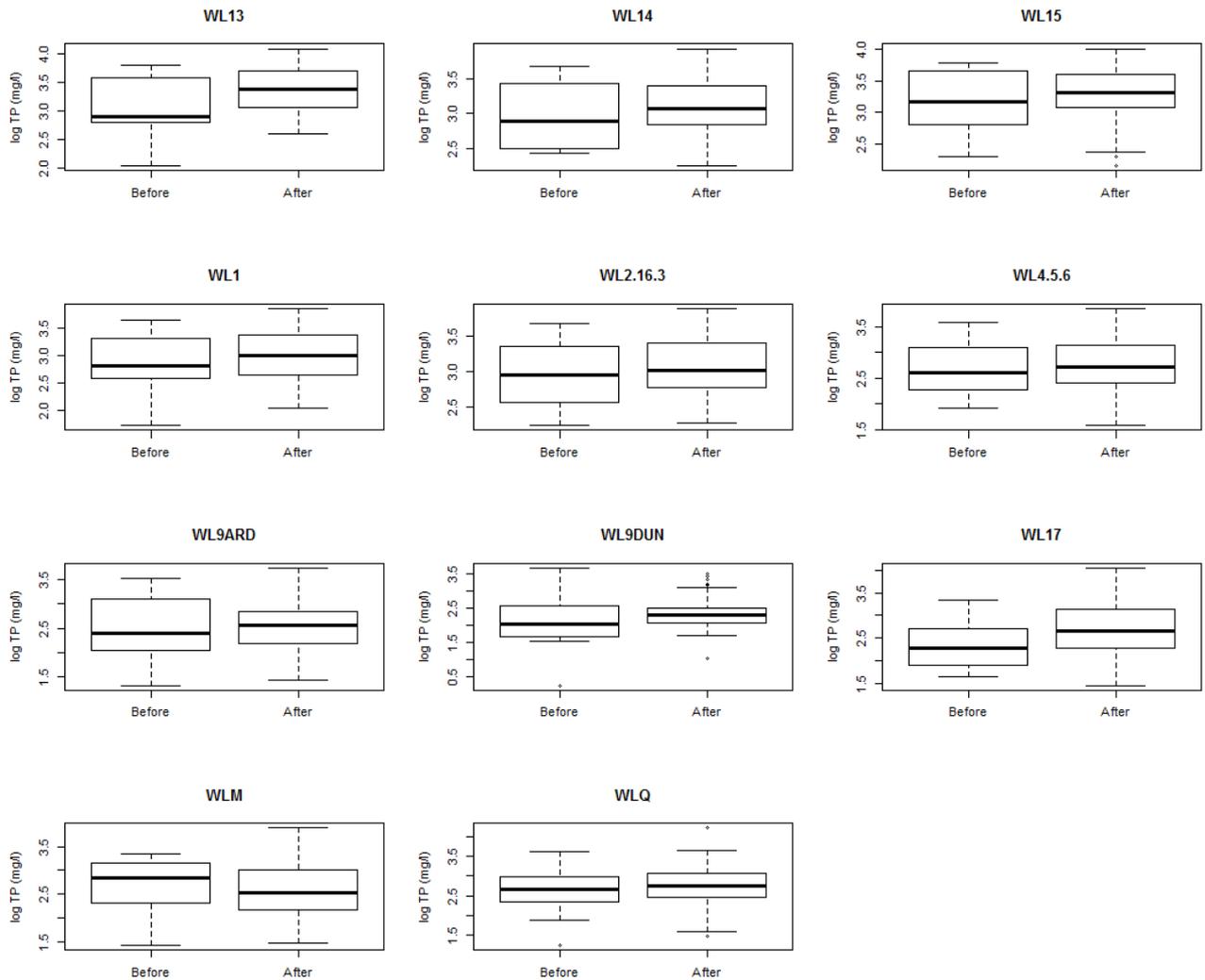


Figure 2.8: Boxplots of before-after at each site

	WL13	WL14	WL15	WL1	WL2.16.3	WL4.5.6	WL9ARD	WL9DUN	WL17	WLM	WLQ
Before	3.04	2.94	3.15	2.88	2.95	2.63	2.49	2.09	2.30	2.66	2.67
After	3.37	3.13	3.30	3.01	3.09	2.74	2.54	2.34	2.66	2.59	2.76
Absolute difference	0.33	0.19	0.15	0.13	0.14	0.11	0.05	0.25	0.36	0.07	0.09

Table 2.7: Table of means and mean differences

As specified before, site WL4.5.6, WL9ARD and WL17 had the lowest percentage of deforested catchment and it would be expected that these sites would be affected less than the others.

From table 2.7, we can see that WL4.5.6, WL9ARD and WL17 had absolute differences of 0.11, 0.05 and 0.36. The absolute difference observed within WLARD is the smallest absolute difference overall and comparatively to other sites the absolute difference observed at WL4.5.6 is also small however the difference observed at WL17 of 0.36 is the largest.

Overall, most absolute differences are small when compared to before and after mean values which indicate that there may not be a large jump in the overall mean which would make the 'Indicator' term within our statistical model insignificant.

From this exploratory analysis alone, it does not seem that there will be a mean difference before and after the changepoint for this variable.

2.3.3.2 BACI Analysis (TOC)

Analysis of Variance table					
Term	df	Sum Sq	Mean Sq	F-value	Pr(>F)
Time	1	1.392	1.392	6.291	0.01229
Month	11	8.692	8.692	39.280	<2e-16
Indicator	1	0.221	0.221	1.002	0.3170
Control/Impact	1	19.023	19.024	85.965	<2e-16
Interaction	1	0.0961	0.096	0.434	0.5100

Table 2.8: BACI Analysis ANOVA (TOC)

Of the 5 terms within the model, 3 are significant with p-values less than 0.05. These terms are Time, Month and Control/Impact. The two terms which are not significant are the Indicator function and the interaction between Indicator and Control/Impact.

Our factor Indicator is not significant which indicates that there is no mean change in log TOC around 2007 however our factor Control/Impact is significant indicating that there is structural difference between control sites and impact sites.

The interaction between the indicator function and Control/Impact is not significant either, which means there is no difference between the mean levels of log TOC for control and impact sites.

This model can now be reduced so that only significant terms are present.

Analysis of Variance table					
Term	df	Sum Sq	Mean Sq	F-value	Pr(>F)
Time	1	1.392	1.392	6.295	0.0122
Month	11	95.615	8.692	39.301	<2e-16
Control/Impact	1	19.024	19.024	86.014	<2e-16

Table 2.9: BACI Analysis ANOVA (TOC – reduced model)

Within table 2.9 is an ANOVA for the reduced model for log TOC. Only three terms are present all with p-values less than 0.05. The interaction term was removed first and the p-value for our indicator term reduced only slightly.

A table of coefficients for our model terms can be seen in table 2.10.

Intercept	Time	M2	M3	M4	M5	M6	M7	M8	M9	M10	M11	M12	Con/Imp (Impact)
-100.48	0.05	-0.3	-0.2	-0.2	-0.1	0.01	0.10	0.43	0.61	0.53	0.36	0.21	-0.29

Table 2.10: Table of coefficients (TOC)

We can now comment on the magnitude of each terms effects on log TOC. The term Month has 12 levels, representing each month. Therefore, log TOC varies by month. With each unit increase in Decimal year, log TOC increases by 0.05. If a site is an impact site, log TOC decreases by 0.29 (control sites are included within the intercept).

Assumptions for the model above are assessed below. Constant variance can be assumed as the spread within the residuals vs. fitted values plot is equal and normality can be assumed from the histogram in figure 2.10.

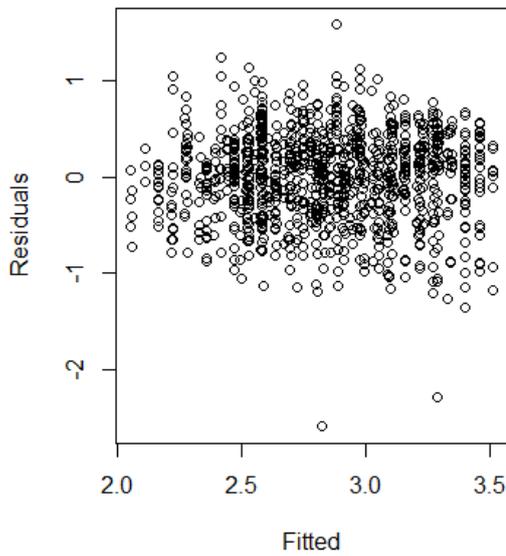


Figure 2.9: Residuals vs. Fitted values

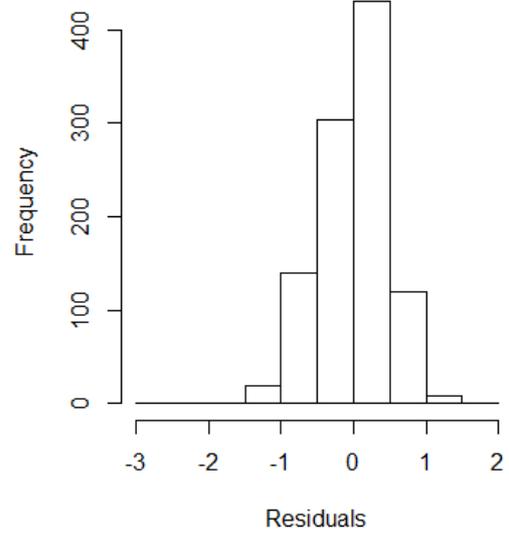


Figure 2.10: Histogram of residuals

2.3.4 Nitrate Oxide

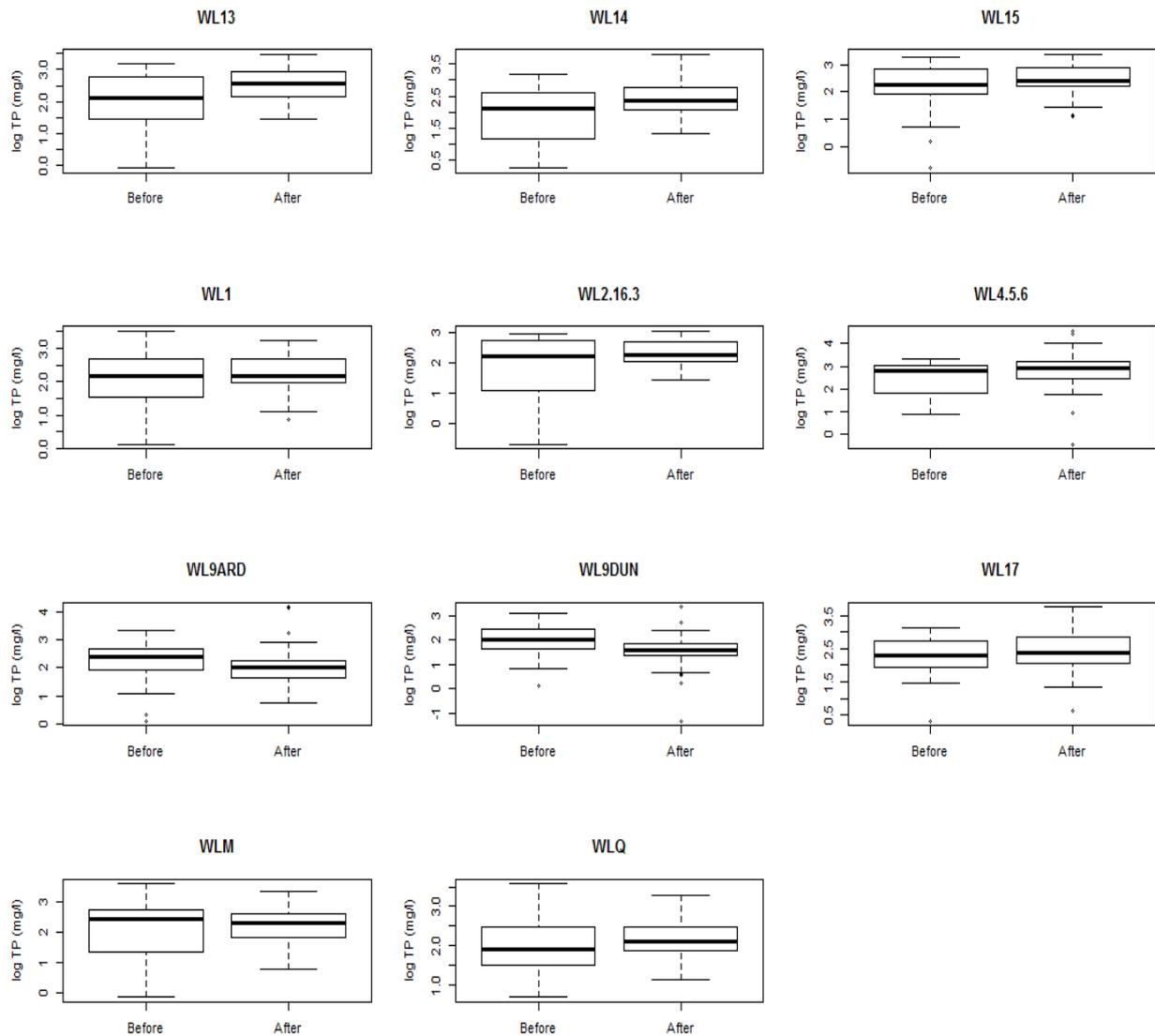


Figure 2.11: Boxplots of before-after at each site

Boxplots and a table of means can be seen within figures 2.11 and table 2.11 respectively. As before, mean changes can be analysed informally to give a subjective impression to whether any changes can be found before and after and between impact and control sites.

2.3.4.1 Exploratory analysis (NO2)

	WL13	WL14	WL15	WL1	WL2.16.3	WL4.5.6	WL9ARD	WL9DUN	WL17	WLM	WLQ
Before	1.97	1.94	2.14	2.09	1.90	2.44	2.21	1.95	2.27	2.14	1.94
After	2.56	2.40	2.47	2.25	2.33	2.83	2.02	1.56	2.37	2.23	2.17
Absolute difference	0.59	0.46	0.33	0.16	0.43	0.39	0.19	0.39	0.10	0.09	0.23

Table 2.11: Table of means and mean differences

From the table above, our three control sites WL4.5.6, WL9ARD and WL17 have absolute differences of 0.39 0.19 and 0.10 respectively.

Overall, the absolute mean difference for WL17 is second smallest after WLM however the absolute mean difference for WL9ARD is fourth smallest and WL4.5.6 is actually fourth largest out of 11.

Subjectively there may possibly be differences between impact and control sites, but overall there are some large differences between before and after means. The absolute difference for WL13 for example is 0.59 which is large in comparison to the mean value before and after.

Again, we will use WL4.5.6, WL9ARD and WL17 as our controls.

2.3.4.2 BACI Analysis (NO2)

Analysis of Variance table					
Term	Df	Sum Sq	Mean sq	F-value	Pr(>F)
Time	1	0.742	0.742	2.304	0.1293
Month	11	124.795	11.342	35.227	<2.2e-16
Indicator	1	5.795	5.792	17.986	2.24e-05
Control/Impact	1	4.565	4.564	14.1732	0.0017
Interaction	1	0.005	0.005	0.168	0.8977

Table 2.12: BACI Analysis (NO2)

From table 2.12, our ANOVA table, we can describe each parameters effect within the model. Of the 5 parameters, 3 are significant. Time indicates the presence of overall trend however this is not significant. Month is significant indicating that a seasonal trend is present. Indicator is significant which means that there is a significant jump in the mean at mid-2007. Control/Impact is significant, which indicates that there is structural difference between control and impact sites. The interaction between Indicator and Control/Impact is not significant which means that there is no mean difference between control and impact groups at mid-2007.

This model can now be reduced so that it only contains significant terms.

Analysis of Variance table					
Term	df	Sum Sq	Mean sq	F-value	Pr(>F)
Month	11	121.466	11.042	34.328	<2.2e-16
Indicator	1	9.589	9.589	29.811	6.11e-08
Control/Impact	1	4.565	4.565	14.190	0.000176

Table 2.13: BACI Analysis (NO₂ – reduced model)

Within table 2.13 we can see the ANOVA of our final model to predict log NO₂. All three terms within this model are significant with p-values less than 0.05. A table of coefficients for our final model can be seen in figure 2.14.

Intercept	M2	M3	M4	M5	M6	M7	M8	M9	M10	M11	M12	Indicator (After)	Con/Imp (Impact)
1.39	0.11	0.26	0.46	0.69	0.96	1.08	0.93	1.10	0.84	0.73	0.51	0.21	0.16

Table 2.14: Table of coefficients (NO₂)

We can interpret the model by assessing each estimate of the coefficient for each term. As Month has 12 levels, log NO₂ varies with each level of this term. If a site is an Impact site, log NO₂ increases by 0.16. If we wish to estimate after the changepoint, log NO₂ increases by 0.21 on average.

Again, linear model assumptions must be checked before any analysis can take place. Independence is again assumed by design and constant variance can be assumed from figure 2.10 as the datapoints do not fan. Normality of residuals can also be assumed as the histogram of residuals is bell shaped.

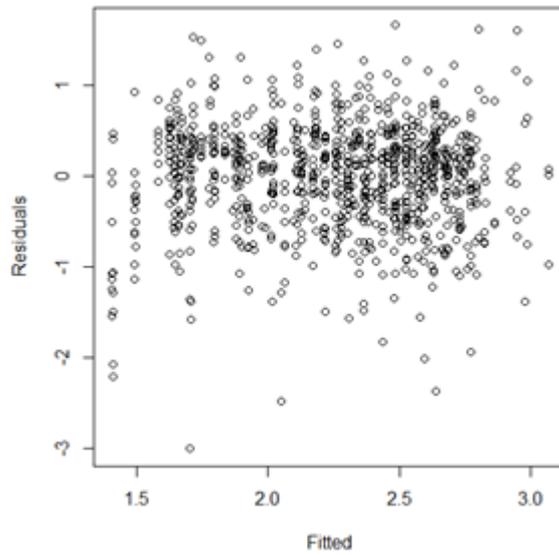


Figure 2.12: Residuals vs. Fitted values

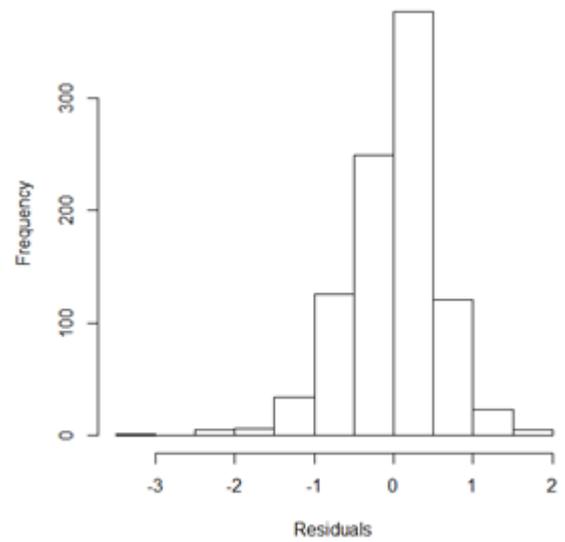


Figure 2.13: Histogram of residuals

2.4 Conclusions

To summarise, analysis was carried out on the three variables log Total Phosphorus, log Total Organic Carbon and log Nitrate Oxide.

Dummy variables were set up to indicate where the changepoint was proposed to be around mid-2007 and also a variable to indicate whether each site was a control site or an impact site. Control sites were chosen as those with the smallest percentages of deforested area (<2%). The control sites chosen were the same for all three variables.

If the variables Indicator and Control/Impact were significant on their own, this indicated that there were significant differences before and after the changepoint and significant differences between control and impact groups respectively.

The interaction between Indicator and Control/Impact indicates that there is a significant difference before and after the changepoint and the magnitude is dependent on whether the site is a control or impact site.

The table below shows each parameter for each variable and indicates whether each parameter within each model is significant or not. This allows us to determine whether there are structural differences before and after the changepoint, between control and impact groups and finally whether these variables interact.

	Dec.year	Month	Ind	Control/Impact	Ind:Control/Impact
TP	Yes	Yes	Yes	Yes	Yes
TOC	Yes	Yes	No	Yes	No
NO2	No	Yes	Yes	Yes	No

Table 2.15: Table of significant parameters

From table 2.15, we can see that all terms within our TP model are significant. Therefore there is evidence of a change in mean before and after the changepoint and the magnitude of the change depends on whether the site in question is an impact or control site. For our TOC model, the indicator function and the two way interaction are not significant. This

means that there is no overall change in mean before and after changepoint. Our final model, our model for NO₂, does not have a significant two way interaction or a significant overall trend. This indicates there is no difference between impact and control sites in terms of a change in mean at the changepoint at mid-2007.

In terms of model assumptions, all of our time series models had normality of errors. Within our TP residuals, two were outliers however we are still able to assume normality. To summarise, seasonal trend was present within all of our series as indicated by the significance of Month and there was also an overall structural difference between impact and control sites within all three variables as indicated by the Control/Impact term. For each variable, model coefficients suggested that variables were higher during the summer months than the winter months. There was overall trend present within both TP and TOC but not within NO₂, and there was an overall mean difference within both TP and NO₂ but not TOC before and after the changepoint.

To summarise, models for TP and NO₂ showed that there was a significant difference before and after the changepoint however there is no evidence to suggest a difference in the TOC series. All three series showed significant differences between the control and impact series.

BACI analysis only allows us to analyse whether a change occurs at a known point in time. If this point is not known, it must be estimated. Changepoint analysis allows us to statistically detect changes in structure within a time series and there are many methods available to do this. There are also problems with spatial aspects of this analysis. The control sites chosen here were simply those which were least affected by deforestation, and were chosen after the intervention had taken place. If control sites were chosen before the turbines had been set in place and deforestation had occurred, as well as been situated far enough away not to be affected by the intervention sites, the analysis would be much more precise.

Furthermore, temporal correlation is ignored by design which leads to the possibility of overstating evidence for associations of interventions.

The following chapter reviews changepoint analysis and the methods that have been used and improved within the past. Three of these techniques will be chosen and applied to a simulation study and then applied to the real data.

Chapter 3

Changepoint Analysis overview and Simulation Study

3.1 Overview of Changepoint Analysis Methods

The aim of this brief review is to assess current methodologies that are in use within environmental impact assessments. It will mainly focus on the statistical disciplines involved in determining whether an intervention adversely affects the environmental conditions. We will discuss methods used by others in determining the effects of various interventions and in the process compare and contrast the situations to which they apply these methods.

The least complicated analysis may be carried out with the use of regression techniques. For example Multiple Linear Regression is a linear modelling technique described by Vincent (1996) and slightly modified by Vincent and Gullet (1999). It is used to determine where regime shifts, if any, lie within a time series. The technique consists of the application of four increasingly complicated linear regression models, including terms for changes in trend and mean for example. After the application of each model, the residuals are analysed in order to assess fit. Consecutive significant autocorrelations in the residuals identified at low lags indicate the poor fit of the model, and in this case the fitted model is rejected and a different model is applied.

The use of likelihood criteria may be used to discriminate between a collection of regression models and determine which is the best fit to the data. For example, the Bayesian Information Criterion was used by Beaulieu (2010) to discriminate between models, as is the Akaike Information Criterion throughout statistics; however, Reeves and Chen (2006) state that Schwarz Information Criterion (SIC) penalizes more heavily than the AIC and tends to yield simpler models within time series.

The model that minimises the criterion is considered to be the most appropriate, taking into account both the number of parameters and the goodness of fit by residuals. The models used can incorporate shifts in the mean, variance, trends or combinations of these shifts.

In terms of using regression where we have access to a control series, a difference series may be calculated and analysis may be carried out on it. Developed by Hinkley (1969), two phase regression consisted of a model which accounted for a change point at time t . This method has been modified by Sollow (1987), Easterling and Peterson (1995) and later by Lund and Reeves (2002). A difference series between the impact and control series is calculated. A simple regression model is fitted to the entire difference series and we can then obtain the residual sum of squares denoted as RSS_1 . We then fit two models for each data point x_i : one before and one after x_i so that we have twice as many models as data points. RSS_2 can then be calculated simply as the minimum of the RSS values from the models we have built. A test statistic is then calculated and if the test statistic is above a critical value, then we can conclude there is a potential step at the data point corresponding to RSS_2 . One major drawback mentioned by Marchant (2008) is the models reliance on constant variance of the error term, and therefore this method can only be used to determine a shift in mean and trend but only if the variance is constant and therefore has not been affected by the intervention. Furthermore, it is suggested that any outliers can affect the efficiency of the least squares estimates.

The idea behind two phase regression can also be applied to non-parametric smoothing methods where at each point within the series smooths are calculated to the left and right of that point. If the difference in the two model trajectory's at point x_i is large, this indicates the presence of a discontinuity, this approach was first proposed by Hall and Titterington (1992) and Muller (1992) with the use of kernel estimates. Many others authors that used kernels were Speckman (1994) and Eubank and Speckman (1994) as well as many others. Local polynomial regression methods are also widely used such as those in Gregoire and Hamrouni (2002a) and Bowman et al. (2003).

Box and Tiao (1975) themselves developed a method to detect a change or 'intervention' at some time point. Intervention analysis introduced by Box and Tiao (1975) can be used to detect and model possible trends in a set of environmental time series, specifically to identify and model whether a statistically significant change in the time series occurs after intervention. A quote from Box and Tiao's 1975 original paper states the outline of the problem intervention analysis was intended to solve:

“Given a known intervention, is there evidence that change in the series of the kind expected actually occurred, and, if so, what can be said of the nature and magnitude of the change?”

Intervention analysis can be used to model both abrupt and gradual effects within a series and can accommodate both temporary and permanent changes. Examples of ‘abrupt’ and ‘gradual’ effects of intervention can be found in Timothy D Hogan’s paper on U.S fertility rates (Hogan (1984)). Further information on the above model can be found in Box & Tiao (1975) or work by Abraham (1980) who subsequently extended the results for use on multiple time series rather than a univariate case.

Other techniques have been developed such as the use of cumulative sums (CUSUM) to determine the location of changes. Page (1954) firstly used cumulative sums to sequentially detect discontinuities and laid the fundamental basis for its use and in 1955 he developed a retrospective method. Since then it has been used and developed for discontinuity detection for example by Pettitt (1980), Bagshaw and Johnson (1975) and Yashchin (1993) for both sequential and retrospective analysis. Inclan and Tiao developed a method in 1994 which uses the cumulative sum of squares at each data point to detect a change in variance and Rodionov (2004) developed STARS, a sequential method of using cumulative sums and t statistics to determine the presence of regime shifts.

CUSUM is widely used within financial analysis and can be used effectively within volatility modelling, as varying volatility can be treated as varying error. An example of CUSUM used in this context can be seen Inclan et al. (1999) where a GARCH (a model used to characterise and model observed time series where the terms within the model are believed to have variance) framework for errors was used and CUSUM was used to determine changes in variance. The changes in volatility could then be attributed to world events that affect various price indexes.

Recently, many authors have used and developed Bayesian techniques to determine whether a change has occurred. Bayesian techniques allow the user to calculate the probability of change at each data point within the series.

Zhao et al. (2005) used Bayesian analysis with the use of MCMC to model hurricane counts by a Poisson process where the rate parameter λ is treated as a random variable modelled by a gamma distribution.

There were three hypotheses formulated: no change in the rate, single change in the rate and double change in the rate. A hierarchical Bayesian approach was used to demonstrate posterior probabilities of the model parameters for each hypothesis through MCMC.

Wyse et al. (2010) presented a paper which also used Markov Chain Monte Carlo to perform retrospective inference on change point models which are collapsible. Wyse rephrased the problem as a stochastic model search over a large model space with the Bayes factors for competing models appearing in the acceptance probabilities for the MCMC sampling scheme.

Further application of MCMC to change point detection can be seen in Antock et al. (2008) where average year temperatures were analysed applying three models with random coefficients. The posterior distribution of the changepoint and other parameters were estimated from the random samples generated by the combination of the Metropolis-Hastings algorithm and the Gibbs sampler.

De-Lacy et al. (2008) explored the situation where the data could be modelled by multiple polynomial regression and by “exploiting” Bayesian theory De-Lacy proposed a method to detect discontinuities. The proposed method consisted of applying Bayesian theory to compute the marginal posterior distribution of the discontinuity and to detect it as maximum a posteriori (MAP).

Beaulieu et al. (2010) presented the Bayesian Normal Homogeneity Test (BNHT). Beaulieu states that BNHT may be applied to a series of ratios or differences between the base series and neighbour series as proposed in Alexanderson (1986) (SNHT). Beaulieu changed the prior probabilities of no change with p equal to 0.01, 0.05, 0.10, 0.25, 0.5, 0.75, 0.90, 0.95 and 0.99. He found; high prior probabilities of no change resulted in low false detection rates on homogeneous series, the test had high power on a series with a single shift, small shifts are detectable with a low prior of ‘no change’ and when the test was applied to series with more than one shift, as well as a high probability of ‘no change’, it performed well.

Other methods for multiple change points such as reversible jump Markov chain Monte Carlo (RJMCMC) was first proposed by Green 1995, and was used by Rotondi (2002) and more recently by Zhao 2010.

To approximate the full posterior distributions of change point characteristics, Nam et al. (2011) used Finite Markov Chain Imbedding in a Hidden Markov Model setting, and accounted for parameter uncertainty via Bayesian modelling and Sequential Monte Carlo. Nam states that the combination of the two is computationally efficient and does not require estimates of the underlying state sequence.

With application to socio-economic data, Western & Kleykamp (2003) wanted to model political relationships but by also taking into account shifts in institutions, ideas preferences or other social conditions. They used a Gibbs sampler and also stochastically sampled from the conditional posteriors to obtain regression coefficients in a Monte Carlo experiment. Perreault et al. (2000) also used a Gibbs sampler along with a Markovian updating scheme for both single change point analyses of a mean, and also of variance. A case study was introduced to demonstrate the suitability of the Gibbs sampler in an energy inflow setting managed by Hydro-Quebec.

In terms of application to time series data within this thesis, three methods have been chosen. All three are used on a single series of data and each is based on different framework.

- 1.) Local linear regression
- 2.) Binary Segmentation Method
- 3.) Barry and Hartigan algorithm, based on Bayesian theory

All three methods are explained in technical detail in the following section.

The remainder of this chapter will provide technical detail of the three chosen methods then apply each to a simulation study and their performance under various conditions can be analysed.

3.2 Applied Methods

3.2.1 Local linear regression

Bowman et al. (2006) developed a method where by using normal kernel smoothing, datum which have the largest difference between left and right model trajectories could be used to identify the possible presence of discontinuity. He used local linear regression with weights at each data point (potential step point t), one model that uses data from the left of point x_i and one that uses data from the right, where the weights are taken from a normal density function so that attention is focused only on the data lying near x_i . This is done for each data point in the whole series.

We let,

$$y_i = m(x_i) + \varepsilon_i \quad (3.1)$$

Where $m(x_i)$ is a regression function of unspecified smooth shape with a finite number of jump points and $\varepsilon_i \sim N(0, \sigma^2)$. The smoother used to estimate $m(x)$ is local linear regression where we need to solve

$$\min_{\alpha, \beta} \sum_{i=1}^n \{y_i - \alpha - \beta(x_i - x)\}^2 w(x_i - x; h) \quad (3.2)$$

and take the value of $\hat{\alpha}$ as this defines the position of the line at x . The local linear estimator can be given by calculating ordinary least squares OLS with weights which gives us the general form:

$$\hat{m}(x) = \frac{1}{n} \sum_{i=1}^n \frac{\{s_2(x; h) - s_1(x; h)(x_i - x)\} w(x_i - x; h) y_i}{s_2(x; h) s_0(x; h) - s_1(x; h)^2} \quad (3.3)$$

Where $w_i = w(\cdot; h)$ is defined by a normal distribution with mean 0 and standard deviation h and

$$s_r(x; h) = \{\sum (x_i - x)^r w(x_i - x; h)\} / n \quad (3.4)$$

If we have extra information on each data point, this can be incorporated into the least square function. For example, in Bowman et al. (2004) the precision of each data point in radio carbon data is known. Bowman used $\frac{1}{p_i^2}$ so that the problem now becomes:

$$\min_{\alpha, \beta} \sum_{i=1}^n \{y_i - \alpha - \beta(x_i - x)\}^2 w(x_i - x; h) \frac{1}{p_i^2} \quad (3.5)$$

which can be solved using weighted least squares to yield the linear smoother $m(x)$. At each point of interest x , we simply calculate local linear estimators to the left and right of x yielding $\widehat{m}_L(x)$ and $\widehat{m}_R(x)$. If there is a jump at x then we can then calculate $\widehat{m}(x^-)$ and $\widehat{m}(x^+)$ and the basic information on the presence of a discontinuity is contained in $\{\widehat{m}_L(x) - \widehat{m}_R(x)\}^2$. A better comparison is achieved when we standardise by the variance of the difference such that $\text{var}\{\widehat{m}_L(x) - \widehat{m}_R(x)\} = v(z_i)\sigma^2$ where $z_i = (x_i + x_{i+1})/2$ assuming the observed x_i are arranged in an increasing order. If we have $x_i = x_{i+1}$, these observations are omitted as well as any z_i which leaves less than 5 data points for the construction of the left or right estimate (i.e if the data point is so far left or right that there are less than 5 data points to construct a smooth we stop).

The test statistic is then:

$$T = \sum_{i=1}^g \frac{\{\widehat{m}_L(x) - \widehat{m}_R(x)\}^2}{v(z_i)\widehat{\sigma}^2} \quad (3.6)$$

Where the presence of a discontinuity is expressed in the p-value $P(T > T_{obs})$. The distribution for T under the null hypothesis of the discontinuity is a shifted and scaled χ^2 distribution. For further details see Bowman (2006). Our null hypothesis, H_0 , is that no discontinuities are present. Essentially the choice of h will determine the flexibility of the model. If h is too small, the kernels will only take a small number of data points surrounding the centre of the kernel into account which will result in a very flexible model. If h is too large then the kernel will take too many data points into account and will not be flexible enough. As with all methods which are based on smoothing, choosing the smoothing parameter is always a problem. We must choose a parameter which captures enough variability that we do not accidentally detect a changepoint, but we do not want to choose a parameter which does not detect a change because it tracks the data too closely. Because of this, a variety of values of h will be chosen and the effects on the methods ability to detect changepoints will be analysed.

3.2.2 Barry and Hartigan algorithm

Barry and Hartigan (2003) developed an algorithm for change point detection which assumes that observations are independent. We let the probability of change at each point i be given the probability of change p_i . Essentially the algorithm splits the data into blocks such that the mean is the same within each block. By applying the algorithm, we detect changepoint between these blocks. Barry and Hartigan state that independent assumptions can be weakened since ‘all that is required is that, given the partition and the parameters, observations in different blocks are mutually independent’.

The prior distribution μ_{ij} (the mean of the block beginning $i+1$ and ending j) is chosen as $\sim N(\mu_0, \frac{\sigma_0^2}{j-1})$. The algorithm uses a partition $p = (U_1, U_2, \dots, U_n)$ where $U_i = 1$ indicates a changepoint at position $i + 1$, we initialize $U_i = 0$ for all $i < n$.

The process starts with a Markov chain. In each step of the Markov chain, at each position i , a value U_i is drawn from the conditional distribution of U_i given the data and the current partition. Following Barry and Hartigan, we let b denote the number of blocks obtained if $U_i = 0$ conditional on U_j . The transition probability p for the conditional probability of a change point at the position $i+1$ can be obtained from the ratio:

$$\frac{p_i}{1-p_i} = \frac{P(U_i=1 | X, U_j, j \neq i)}{P(U_i=0 | X, U_j, j \neq i)} \quad (3.7)$$

Where W_0, B_0, W_1 and B_1 are the within and between block sums of squares obtained when $U_i = 0$ and $U_i = 1$ respectively and X is the data. The tuning parameters γ and λ allow us to place restrictions on the priors and may take values between 0 and 1, chosen so that the method is effective in situations where there are not many changes (γ small) and where the changes that do occur are of reasonable size (λ small). After each iteration the dataset is updated conditional on the current partition.

No null or alternative hypothesis is specified within this framework but rather a probability of a change point being present. Therefore, we must chose cut off points for probabilities of change, which will be discussed in the next chapter.

3.2.3 Binary Segmentation Method

Early applications of Binary Segmentation Method include Scott and Knott (1974) and Sen and Srivastava (1975). This method uses two separate test statistics to determine if a change point has occurred for both single and multiple changepoints.

Killick (2011) developed an R package which allows users to apply BSM to time series data. For the single changepoint series, we consider a changepoint τ_1 . We will use a likelihood ratio test statistic to determine whether there is a change in the series. The use of likelihood ratio tests within changepoint analysis was first proposed by Hinkley (1970) and the test statistic was firstly used to determine a change in the mean within normally distributed data, however Gupta and Tang further developed the test to include changes in variance.

This method requires us to maximise the log likelihood value under both the null (no change detected) and alternative hypotheses (change detected) where the maximum log-likelihood value under the null hypothesis is $\log p(y_{1:n}|\hat{\theta})$ where $p(\cdot)$ is the PDF and $\hat{\theta}$ is the MLE of the parameters.

Under the alternative hypothesis we consider a model with changepoint at τ_1 where τ_1 can take any value in the closed set $(1, 2, \dots, n - 1)$ then the maximum log likelihood for a given τ_1 is:

$$\lambda = 2[\max ML(\tau_1) - \log p(y_{1:n}|\hat{\theta})] \quad (3.8)$$

Such that we would reject H_0 if $\lambda > c$ and then estimate the position of the changepoint ($\hat{\tau}_1$) as the value of τ_1 that maximises $ML(\tau_1)$.

We can modify the single test to maximise $ML(\tau_1)$ over m segments allowing us to determine the location of multiple changepoints. The method requires us to minimise the function:

$$\sum_{i=1}^{m+1} [C(y(\tau_{i-1} + 1) : \tau_i)] + \beta f(m) \quad (3.9)$$

Here C is a cost function for a segment and $\beta f(m)$ is a penalty to guard against over fitting. Using notation from the single changepoint section, C may be taken as the negative log likelihood and $\beta f(m)$ may be cm .

The algorithm shown below applies the single changepoint test and upon identifying a change, iteratively implements the test statistic on the sub-segments of the data developed by Killick et al (2011).

Input: A set of data of the form (y_1, y_2, \dots, y_n)

A test statistic $\lambda(\cdot)$ dependent on the data

An estimator of changepoint position $\tau(\cdot)$

A rejection threshold (penalty), c

Initialise: Let $C = \phi$ and $S = \{[1, n]\}$

Iterate while $S \neq \phi$

1. Choose an element of S ; denote this element as $[s, t]$.
2. If $\lambda(y_{s:t}) < c$ remove $[s, t]$ from S .
3. If $\lambda(y_{s:t}) \geq c$ then;
 - (a) remove $[s, t]$ from S
 - (b) calculate $\tau = \hat{\tau}(y_{s:t}) + s - 1$, and add τ to C ;
 - (c) if $\tau \neq s$ add $[s, \tau]$ to S
 - (d) if $\tau \neq t - 1$ add $[\tau + 1, t]$ to S

Output the set of changepoints recorded C

(Taken from Killick (2011))

In essence the method extends a single changepoint method to multiple changepoints by repeating the method on varying subsets of the series iteratively:

The penalty can take any value, however Killick (2011) uses $\lambda * \log(n)$ where n is the number of observations within the series and λ can be arbitrarily changed depending on the expected size of the change.

3.3 Simulation study

In this section we will consider how various tests perform at detecting a discontinuity within a time series. We wish to determine the effects of changing shape and trend in the series change the chances of detecting a discontinuity when it is present within a time series, as well as the ability to detect a discontinuity when various model parameters are changed such as variance and serial correlation. Of the 6 cases, 3 will not include seasonal trend and 3 will include seasonal trend which reflect the application areas.

3.3.1 Statistical model

Our statistical model which we will use to assess the various approaches can be described as:

$$y_i = f_1(x_i) + \epsilon_i \quad \text{where } x < c$$
$$y_i = f_2(x_i) + \epsilon_i \quad \text{where } x \geq c$$

Where the functions $f_1(x_i)$ and $f_2(x_i)$ can be used to describe the trend of the data before and after the change point which is denoted as c . These functions will take a variety of forms such as straight lines and trigonometric curves to simulate seasonal trends. It is reasonable to adopt a simple AR(1) model for the correlation structure for the error, which can be described as:

$$\epsilon_i = \alpha\epsilon_{i-1} + Z_i \tag{3.10}$$

Where Z_i is purely random process such that $E(Z_i) = 0$ and $Var(Z_i) = \sigma^2$ and α is our chosen correlation coefficient. Our error term can therefore be fully described as $\epsilon \sim N(0, \sigma^2 \Sigma)$ where Σ is our correlation matrix.

In real life situations, the correlation coefficient may have an impact on our ability to determine if a discontinuity is present, and therefore it may be necessary to estimate this coefficient if the test does not take autocorrelation into account. To do this, we must firstly remove the trend that may be present within the data and perform analysis on the leftover residual series – this can be done by using both parametric and nonparametric modelling procedures. In this distinct case, we wish to determine the effects of changing α .

The simulation study will let us analyse the effects of changing this parameter and determine whether we must consider estimating this coefficient before performing subsequent analysis. In this case we did not estimate α from any of the series, we simply vary it to determine its effects on each tests size and power.

For each distinct case, the chosen test will be run 200 times on uniquely simulated, equally spaced data. This allows us to determine the proportion of times the test has correctly identified the discontinuity. Probability plots will be constructed to determine the areas where the test identified where discontinuities as present.

Only one parameter will be changed at a time, leaving the rest unchanged. This allows us to determine the effects of the change of that single parameter – and subsequently allow us to determine which parameters are most suitable to use for our real data.

3.3.2 Size and power

The performance of the various tests can be determined by calculating the size and the power, given various parameters of the test. Size and power can be described in terms of type I (the odds of saying there is not a difference when there is) and type II error (the odds of saying there is a changepoint when there is not).

Size can be calculated as the proportion of times that the test concludes that there is a discontinuity when there is in fact not. This can be computed simply by performing the test on a simulated series where there is no changepoint. Power can be calculated as the proportion of times that the test concludes that there is *at least one* discontinuity when there actually *at least one* present, and can be calculated by performing the test on series with at least one discontinuity.

In hypothesis terms, we could say:

$$H_0: \text{no changepoint} \quad \text{size} = P(\text{reject } H_0 | H_0 \text{ true})$$

$$H_1: \text{changepoint} \quad \text{power} = P(\text{reject } H_0 | H_1 \text{ true})$$

3.3.3 Scenarios

We will perform the various tests on 6 distinct scenarios, listed below. The plots shown are without random variation to show the underlying true curves.

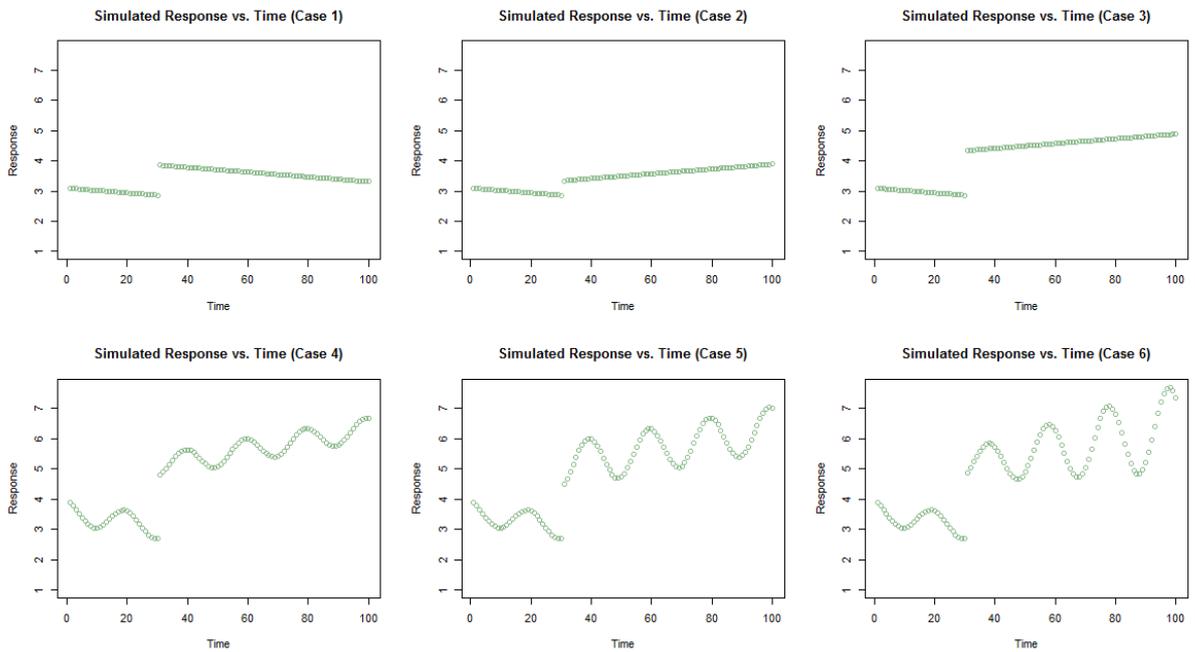


Figure 3.1: Simulation study scenarios

Model	Model description
Scenario 1: Change in intercept	$y_i = 3.1 + 0.008x_i, \quad i = 1, \dots, 50$ $y_i = 4.1 + 0.008x_i, \quad i = 50, \dots, 100$
Scenario 2: Change in slope	$y_i = 3.1 - 0.008x_i, \quad i = 1, \dots, 50$ $y_i = 3.1 + 0.008x_i, \quad i = 50, \dots, 100$
Scenario 3: Change in slope & intercept	$y_i = 3.1 - 0.008x_i, \quad i = 1, \dots, 50$ $y_i = 4.1 + 0.008x_i, \quad i = 50, \dots, 100$
Scenario 4: Change in slope & intercept with seasonal trend	$y_i = 3.6 - 0.17x_i - 0.1 \sin\left(\frac{2\pi x_i}{20}\right) + 0.37 \cos\left(\frac{2\pi x_i}{20}\right), \quad i = 1, \dots, 50$ $y_i = 4.6 + 0.17x_i - 0.1 \sin\left(\frac{2\pi x_i}{20}\right) + 0.37 \cos\left(\frac{2\pi x_i}{20}\right), \quad i = 50, \dots, 100$
Scenario 5: Change in slope & mean with seasonal trend (change in seasonal pattern)	$y_i = 3.6 - 0.17x_i - 0.1 \sin\left(\frac{2\pi x_i}{20}\right) + 0.37 \cos\left(\frac{2\pi x_i}{20}\right), \quad i = 1, \dots, 50$ $y_i = 4.6 + 0.17x_i - 0.2 \sin\left(\frac{2\pi x_i}{20}\right) + 0.70 \cos\left(\frac{2\pi x_i}{20}\right), \quad i = 50, \dots, 100$
Scenario 6: Change in slope & mean with seasonal trend (change in seasonal pattern - varying coefficient)	$y_i = 3.6 - 0.17x_i - 0.1 \sin\left(\frac{2\pi x_i}{20}\right) + 0.37 \cos\left(\frac{2\pi x_i}{20}\right), \quad i = 1, \dots, 50$ $y_i = 4.6 + 0.17x_i - r_i \sin\left(\frac{2\pi x_i}{20}\right) + r_i \cos\left(\frac{2\pi x_i}{20}\right), \quad i = 50, \dots, 100$ $(r_i \text{ increases from 1 to 2, equally spaced over the 50 datapoints})$

Table 3.1: Simulation study scenarios

The error is normally distributed with constant variance, where σ_i was estimated from site WL13 at Whitelee by simply fitting a local linear regression model to the series and calculating the standard deviation from the residuals. Scenarios 1-3 are simple straight lines with either a jump in the mean value or change in trend. Scenarios 4, 5 and 6 incorporate trigonometric functions to represent seasonality over time, within each case 4 years' worth of data are represented by 4 continuous cycles.

Scenario 4 has a simple shift in the mean at the change point, with seasonal patterns which are identical before and after this point. Scenario 5 has both a shift in the mean and also a change in seasonal pattern, with the amplitude changing from 3 to 5 units.

Scenario 6 is the most complicated of the cases, with a varying coefficient. This coefficient allows us to change the amplitude of the trigonometric curves where the coefficients of the sin and cosine functions increase or decrease with time. Technically, cases 4 and 5 can be explained by varying coefficient models, however their coefficients are constant.

3.4 Test Conditions

3.4.1 Test Conditions: Local linear regression (LLR)

Described in section 3.1, Bowman (2006) developed a test based on local linear regression with weights to determine the presence of a discontinuity. We will allow our smoothing parameter, the bandwidth of the kernel to take a number of values – this will allow us to determine the effect of the smoothing parameter.

Bowman’s algorithm does not take serial correlation into account when assessing the series and by default assumes that data points are independent. The size test will therefore be around 5% where there is no serial correlation, however if serial correlation is present this may not be the case. A modified version of the test does allow the user to input an estimate of the error covariance matrix into the function however since we wish to test the effects of correlation on the test we will not use the modified version (preliminary analysis of the Whitelee series shows no significant autocorrelation).

In real life situations, if we wish to use the modified test accounting for serial correlation, Σ will have to be estimated, which we will call $\hat{\Sigma}$. As mentioned previously, this can be estimated by modelling the general trend of the data and assessing the residuals. One problem with this method is that, especially with non-parametric smoothing techniques where the kernel bandwidth must be chosen, the data may actually be over smoothed resulting in a lower estimate of the correlation. The opposite is also true in that the data can also be under smoothed leaving a hidden function within the data, resulting in a higher estimate of the correlation.

From preliminary analysis we know that we do not have any significant autocorrelation, the covariance matrix is simply an $AR(1)$ as shown below:

$$\Sigma = \begin{bmatrix} 1 & \alpha & \alpha^2 & \dots & \alpha^{n-1} \\ \alpha & 1 & \alpha & \dots & \alpha^{n-2} \\ \alpha^2 & \alpha & 1 & \dots & \vdots \\ \vdots & \vdots & \vdots & \ddots & \vdots \\ \alpha^{n-1} & \dots & \dots & \dots & 1 \end{bmatrix}$$

If the true value of α is not known, we must obtain $\hat{\alpha}$ - our estimate of the correlation, to obtain $\hat{\Sigma}$.

The smoothing parameter, which in this case is the kernel bandwidth of the weights must also be chosen when carrying out this test. This is denoted as h . This parameter determines the standard deviation of the kernel, which is a normal probability distribution. This means that points that lie close to the point of interest will have more of an influence than points lying farther away.

For scenarios 4-6, smaller values of h will be used since the data has a seasonal cycle. The smaller bandwidth allows the smoothed curve to follow the data more closely since points that are farther away, and therefore less likely to have a relationship with the point of interest, will have less influence on the smoothed curve.

To determine each test's power we have chosen a range of bandwidths to assess. The choices represent small moderate and large amounts of smoothing for the regression.

The simulated data will have the following characteristics:

Model and data conditions	
Data spacing	Equally spaced
Number of simulations	100
Length of time series	100
Varying parameters	
Error	$\epsilon \sim N(0, \sigma^2 \Sigma)$, $\sigma = 0.35, 0.50, 0.65$
Correlation	$\rho = 0.1, 0.15, 0.2$
Kernel Bandwidth	$h = 10, 12, 14, 16, 18$

Table 3.2: Simulation conditions

Where within each case, various correlation coefficients, errors and smoothing parameters will be used.

Scenario 1: Change in mean

Correlation = 0			
h	σ_i^2	Size	Power
10	σ_1^2	0.07	0.74
10	σ_2^2	0.01	0.40
10	σ_3^2	0.10	0.31
12	σ_1^2	0.06	0.87
12	σ_2^2	0.01	0.48
12	σ_3^2	0.11	0.34
14	σ_1^2	0.04	0.94
14	σ_2^2	0.01	0.60
14	σ_3^2	0.11	0.39
16	σ_1^2	0.05	0.97
16	σ_2^2	0.03	0.62
16	σ_3^2	0.11	0.45
18	σ_1^2	0.05	0.99
18	σ_2^2	0.04	0.71
18	σ_3^2	0.09	0.48

Correlation = 0.1			
h	σ_i^2	Size	Power
10	σ_1^2	0.17	0.91
10	σ_2^2	0.13	0.53
10	σ_3^2	0.16	0.47
12	σ_1^2	0.16	0.95
12	σ_2^2	0.14	0.63
12	σ_3^2	0.10	0.55
14	σ_1^2	0.18	0.95
14	σ_2^2	0.13	0.71
14	σ_3^2	0.08	0.60
16	σ_1^2	0.17	0.97
16	σ_2^2	0.13	0.76
16	σ_3^2	0.07	0.66
18	σ_1^2	0.15	0.97
18	σ_2^2	0.12	0.77
18	σ_3^2	0.10	0.67

Correlation = 0.15			
h	σ_i^2	Size	Power
10	σ_1^2	0.18	0.89
10	σ_2^2	0.21	0.68
10	σ_3^2	0.27	0.58
12	σ_1^2	0.17	0.95
12	σ_2^2	0.24	0.72
12	σ_3^2	0.25	0.65
14	σ_1^2	0.15	0.99
14	σ_2^2	0.25	0.79
14	σ_3^2	0.25	0.71
16	σ_1^2	0.17	0.99
16	σ_2^2	0.26	0.80
16	σ_3^2	0.21	0.72
18	σ_1^2	0.14	0.99
18	σ_2^2	0.24	0.83
18	σ_3^2	0.19	0.71

Correlation = 0.2			
h	σ_i^2	Size	Power
10	σ_1^2	0.31	0.92
10	σ_2^2	0.36	0.78
10	σ_3^2	0.41	0.63
12	σ_1^2	0.36	0.95
12	σ_2^2	0.33	0.83
12	σ_3^2	0.38	0.68
14	σ_1^2	0.29	1.00
14	σ_2^2	0.32	0.85
14	σ_3^2	0.38	0.69
16	σ_1^2	0.29	0.99
16	σ_2^2	0.29	0.89
16	σ_3^2	0.34	0.72
18	σ_1^2	0.26	0.99
18	σ_2^2	0.27	0.88
18	σ_3^2	0.36	0.74

Table 3.3: Power and size results (LLR, Scenario 1)

Scenario 2: Change in slope

Correlation = 0			
h	σ_i^2	Size	Power
10	σ_1^2	0.06	0.21
10	σ_2^2	0.09	0.15
10	σ_3^2	0.05	0.11
12	σ_1^2	0.06	0.23
12	σ_2^2	0.08	0.16
12	σ_3^2	0.05	0.14
14	σ_1^2	0.06	0.29
14	σ_2^2	0.05	0.17
14	σ_3^2	0.03	0.18
16	σ_1^2	0.05	0.31
16	σ_2^2	0.05	0.18
16	σ_3^2	0.04	0.21
18	σ_1^2	0.05	0.35
18	σ_2^2	0.06	0.21
18	σ_3^2	0.0.	0.21

Correlation = 0.1			
h	σ_i^2	Size	Power
10	σ_1^2	0.17	0.47
10	σ_2^2	0.13	0.26
10	σ_3^2	0.16	0.24
12	σ_1^2	0.16	0.44
12	σ_2^2	0.14	0.25
12	σ_3^2	0.10	0.18
14	σ_1^2	0.18	0.52
14	σ_2^2	0.13	0.26
14	σ_3^2	0.08	0.18
16	σ_1^2	0.17	0.57
16	σ_2^2	0.13	0.27
16	σ_3^2	0.07	0.18
18	σ_1^2	0.15	0.65
18	σ_2^2	0.12	0.28
18	σ_3^2	0.10	0.17

Correlation = 0.15			
h	σ_i^2	Size	Power
10	σ_1^2	0.18	0.49
10	σ_2^2	0.21	0.32
10	σ_3^2	0.27	0.36
12	σ_1^2	0.17	0.50
12	σ_2^2	0.24	0.34
12	σ_3^2	0.25	0.35
14	σ_1^2	0.15	0.52
14	σ_2^2	0.25	0.34
14	σ_3^2	0.25	0.37
16	σ_1^2	0.17	0.59
16	σ_2^2	0.26	0.35
16	σ_3^2	0.21	0.34
18	σ_1^2	0.14	0.66
18	σ_2^2	0.24	0.35
18	σ_3^2	0.19	0.36

Correlation = 0.2			
h	σ_i^2	Size	Power
10	σ_1^2	0.31	0.54
10	σ_2^2	0.36	0.34
10	σ_3^2	0.41	0.47
12	σ_1^2	0.36	0.60
12	σ_2^2	0.33	0.39
12	σ_3^2	0.38	0.42
14	σ_1^2	0.29	0.65
14	σ_2^2	0.32	0.41
14	σ_3^2	0.38	0.40
16	σ_1^2	0.29	0.69
16	σ_2^2	0.29	0.40
16	σ_3^2	0.34	0.41
18	σ_1^2	0.26	0.71
18	σ_2^2	0.27	0.45
18	σ_3^2	0.36	0.37

Table 3.4: Power and size results (LLR, Scenario 2)

Scenario 3: Change in slope and mean

Correlation = 0			
h	σ_i^2	Size	Power
10	σ_1^2	0.06	0.97
10	σ_2^2	0.09	0.81
10	σ_3^2	0.05	0.61
12	σ_1^2	0.06	0.98
12	σ_2^2	0.08	0.89
12	σ_3^2	0.05	0.77
14	σ_1^2	0.06	1.00
14	σ_2^2	0.05	0.93
14	σ_3^2	0.03	0.78
16	σ_1^2	0.05	1.00
16	σ_2^2	0.05	0.97
16	σ_3^2	0.04	0.81
18	σ_1^2	0.05	1.00
18	σ_2^2	0.06	0.97
18	σ_3^2	0.0.	0.82

Correlation = 0.1			
h	σ_i^2	Size	Power
10	σ_1^2	0.17	1.00
10	σ_2^2	0.13	0.91
10	σ_3^2	0.16	0.67
12	σ_1^2	0.16	1.00
12	σ_2^2	0.14	0.98
12	σ_3^2	0.10	0.74
14	σ_1^2	0.18	1.00
14	σ_2^2	0.13	0.99
14	σ_3^2	0.08	0.79
16	σ_1^2	0.17	1.00
16	σ_2^2	0.13	0.99
16	σ_3^2	0.07	0.84
18	σ_1^2	0.15	1.00
18	σ_2^2	0.12	0.99
18	σ_3^2	0.10	0.87

Correlation = 0.15			
h	σ_i^2	Size	Power
10	σ_1^2	0.18	1.00
10	σ_2^2	0.21	0.96
10	σ_3^2	0.27	0.76
12	σ_1^2	0.17	1.00
12	σ_2^2	0.24	0.98
12	σ_3^2	0.25	0.82
14	σ_1^2	0.15	1.00
14	σ_2^2	0.25	0.99
14	σ_3^2	0.25	0.88
16	σ_1^2	0.17	1.00
16	σ_2^2	0.26	0.99
16	σ_3^2	0.21	0.89
18	σ_1^2	0.14	1.00
18	σ_2^2	0.24	0.99
18	σ_3^2	0.19	0.92

Correlation = 0.2			
h	σ_i^2	Size	Power
10	σ_1^2	0.31	1.00
10	σ_2^2	0.36	0.98
10	σ_3^2	0.41	0.86
12	σ_1^2	0.36	1.00
12	σ_2^2	0.33	0.99
12	σ_3^2	0.38	0.91
14	σ_1^2	0.29	1.00
14	σ_2^2	0.32	1.00
14	σ_3^2	0.38	0.93
16	σ_1^2	0.29	1.00
16	σ_2^2	0.29	1.00
16	σ_3^2	0.34	0.92
18	σ_1^2	0.26	1.00
18	σ_2^2	0.27	1.00
18	σ_3^2	0.36	0.93

Table 3.5: Power and size results (LLR, Scenario 3)

Scenario 4: Change in slope and mean (Seasonal trend)

Correlation = 0			
h	σ_i^2	Size	Power
2.2	σ_1^2	0.06	0.20
2.2	σ_2^2	0.08	0.20
2.2	σ_3^2	0.02	0.11
2.4	σ_1^2	0.06	0.31
2.4	σ_2^2	0.10	0.26
2.4	σ_3^2	0.3	0.12
2.6	σ_1^2	0.09	0.46
2.6	σ_2^2	0.12	0.31
2.6	σ_3^2	0.04	0.14
2.8	σ_1^2	0.10	0.63
2.8	σ_2^2	0.11	0.37
2.8	σ_3^2	0.06	0.17
3	σ_1^2	0.13	0.76
3	σ_2^2	0.13	0.47
3	σ_3^2	0.06	0.22

Correlation = 0.1			
h	σ_i^2	Size	Power
2.2	σ_1^2	0.11	0.53
2.2	σ_2^2	0.15	0.32
2.2	σ_3^2	0.19	0.31
2.4	σ_1^2	0.13	0.63
2.4	σ_2^2	0.14	0.38
2.4	σ_3^2	0.20	0.31
2.6	σ_1^2	0.17	0.74
2.6	σ_2^2	0.19	0.45
2.6	σ_3^2	0.26	0.32
2.8	σ_1^2	0.28	0.89
2.8	σ_2^2	0.23	0.58
2.8	σ_3^2	0.29	0.38
3	σ_1^2	0.42	0.94
3	σ_2^2	0.25	0.70
3	σ_3^2	0.30	0.42

Correlation = 0.15			
h	σ_i^2	Size	Power
10	σ_1^2	0.32	0.69
10	σ_2^2	0.29	0.39
10	σ_3^2	0.31	0.38
12	σ_1^2	0.33	0.75
12	σ_2^2	0.29	0.45
12	σ_3^2	0.30	0.45
14	σ_1^2	0.38	0.84
14	σ_2^2	0.31	0.50
14	σ_3^2	0.31	0.48
16	σ_1^2	0.46	0.87
16	σ_2^2	0.38	0.65
16	σ_3^2	0.36	0.52
18	σ_1^2	0.57	0.96
18	σ_2^2	0.44	0.72
18	σ_3^2	0.45	0.59

Correlation = 0.2			
h	σ_i^2	Size	Power
10	σ_1^2	0.47	0.82
10	σ_2^2	0.41	0.53
10	σ_3^2	0.30	0.48
12	σ_1^2	0.51	0.86
12	σ_2^2	0.46	0.60
12	σ_3^2	0.30	0.52
14	σ_1^2	0.61	0.92
14	σ_2^2	0.47	0.71
14	σ_3^2	0.37	0.56
16	σ_1^2	0.73	0.96
16	σ_2^2	0.54	0.75
16	σ_3^2	0.38	0.64
18	σ_1^2	0.79	0.97
18	σ_2^2	0.62	0.81
18	σ_3^2	0.45	0.65

Table 3.6: Power and size results (LLR, Scenario 4)

Scenario 5: Change in slope and mean (Seasonal trend – increase in amplitude)

Correlation = 0			
h	σ_i^2	Size	Power
2.2	σ_1^2	0.06	0.21
2.2	σ_2^2	0.08	0.14
2.2	σ_3^2	0.02	0.06
2.4	σ_1^2	0.06	0.37
2.4	σ_2^2	0.10	0.17
2.4	σ_3^2	0.3	0.12
2.6	σ_1^2	0.09	0.56
2.6	σ_2^2	0.12	0.22
2.6	σ_3^2	0.04	0.15
2.8	σ_1^2	0.10	0.77
2.8	σ_2^2	0.11	0.34
2.8	σ_3^2	0.06	0.19
3	σ_1^2	0.13	0.92
3	σ_2^2	0.13	0.57
3	σ_3^2	0.06	0.26

Correlation = 0.1			
h	σ_i^2	Size	Power
2.2	σ_1^2	0.11	0.49
2.2	σ_2^2	0.15	0.25
2.2	σ_3^2	0.19	0.17
2.4	σ_1^2	0.13	0.65
2.4	σ_2^2	0.14	0.34
2.4	σ_3^2	0.20	0.27
2.6	σ_1^2	0.17	0.81
2.6	σ_2^2	0.19	0.46
2.6	σ_3^2	0.26	0.38
2.8	σ_1^2	0.28	0.91
2.8	σ_2^2	0.23	0.60
2.8	σ_3^2	0.29	0.50
3	σ_1^2	0.42	0.99
3	σ_2^2	0.25	0.75
3	σ_3^2	0.30	0.62

Correlation = 0.15			
h	σ_i^2	Size	Power
10	σ_1^2	0.32	0.66
10	σ_2^2	0.29	0.39
10	σ_3^2	0.31	0.42
12	σ_1^2	0.33	0.80
12	σ_2^2	0.29	0.47
12	σ_3^2	0.30	0.47
14	σ_1^2	0.38	0.92
14	σ_2^2	0.31	0.61
14	σ_3^2	0.31	0.52
16	σ_1^2	0.46	0.98
16	σ_2^2	0.38	0.76
16	σ_3^2	0.36	0.61
18	σ_1^2	0.57	0.98
18	σ_2^2	0.44	0.85
18	σ_3^2	0.45	0.69

Correlation = 0.2			
h	σ_i^2	Size	Power
10	σ_1^2	0.47	0.74
10	σ_2^2	0.41	0.65
10	σ_3^2	0.30	0.45
12	σ_1^2	0.51	0.86
12	σ_2^2	0.46	0.73
12	σ_3^2	0.30	0.50
14	σ_1^2	0.61	0.96
14	σ_2^2	0.47	0.82
14	σ_3^2	0.37	0.59
16	σ_1^2	0.73	1.00
16	σ_2^2	0.54	0.86
16	σ_3^2	0.38	0.68
18	σ_1^2	0.79	1.00
18	σ_2^2	0.62	0.95
18	σ_3^2	0.45	0.82

Table 3.7: Power and size results (LLR, Scenario 5)

Scenario 6: Change in slope and mean (Seasonal trend – varying coefficient)

Correlation = 0			
h	σ_i^2	Size	Power
2.2	σ_1^2	0.06	0.25
2.2	σ_2^2	0.08	0.09
2.2	σ_3^2	0.02	0.05
2.4	σ_1^2	0.06	0.37
2.4	σ_2^2	0.10	0.17
2.4	σ_3^2	0.3	0.08
2.6	σ_1^2	0.09	0.63
2.6	σ_2^2	0.12	0.30
2.6	σ_3^2	0.04	0.11
2.8	σ_1^2	0.10	0.90
2.8	σ_2^2	0.11	0.45
2.8	σ_3^2	0.06	0.20
3	σ_1^2	0.13	0.99
3	σ_2^2	0.13	0.72
3	σ_3^2	0.06	0.29

Correlation = 0.1			
h	σ_i^2	Size	Power
2.2	σ_1^2	0.11	0.47
2.2	σ_2^2	0.15	0.28
2.2	σ_3^2	0.19	0.24
2.4	σ_1^2	0.13	0.70
2.4	σ_2^2	0.14	0.40
2.4	σ_3^2	0.20	0.39
2.6	σ_1^2	0.17	0.92
2.6	σ_2^2	0.19	0.60
2.6	σ_3^2	0.26	0.39
2.8	σ_1^2	0.28	0.99
2.8	σ_2^2	0.23	0.75
2.8	σ_3^2	0.29	0.52
3	σ_1^2	0.42	1.00
3	σ_2^2	0.25	0.91
3	σ_3^2	0.30	0.70

Correlation = 0.15			
h	σ_i^2	Size	Power
10	σ_1^2	0.32	0.61
10	σ_2^2	0.29	0.52
10	σ_3^2	0.31	0.38
12	σ_1^2	0.33	0.83
12	σ_2^2	0.29	0.62
12	σ_3^2	0.30	0.40
14	σ_1^2	0.38	0.96
14	σ_2^2	0.31	0.80
14	σ_3^2	0.31	0.53
16	σ_1^2	0.46	0.99
16	σ_2^2	0.38	0.88
16	σ_3^2	0.36	0.60
18	σ_1^2	0.57	1.00
18	σ_2^2	0.44	0.96
18	σ_3^2	0.45	0.75

Correlation = 0.2			
h	σ_i^2	Size	Power
10	σ_1^2	0.47	0.75
10	σ_2^2	0.41	0.58
10	σ_3^2	0.30	0.48
12	σ_1^2	0.51	0.84
12	σ_2^2	0.46	0.71
12	σ_3^2	0.30	0.56
14	σ_1^2	0.61	0.96
14	σ_2^2	0.47	0.78
14	σ_3^2	0.37	0.65
16	σ_1^2	0.73	0.99
16	σ_2^2	0.54	0.92
16	σ_3^2	0.38	0.77
18	σ_1^2	0.79	1.00
18	σ_2^2	0.62	0.98
18	σ_3^2	0.45	0.84

Table 3.8: Power and size results (LLR, Scenario 6)

3.4.2 Test Conditions: Bayesian Change point Analysis (Barry and Hartigan algorithm)

Described in detail in section 3.2, Barry and Hartigan developed a method based on Bayesian analysis to evaluate the probability of discontinuity at each data point within the series.

The hyper parameters λ and γ are variable and can be changed and tuned assuming that we know the number of changes and the magnitude of these changes, however, the hyperparameters do have default values in which both are equal to 0.2 (which were found to work well by Barry and Hartigan). Since the default value was found to work well for all cases, we will keep them constant throughout the simulated tests.

Furthermore, the number of ‘burn-in’ (number of iterations we throw away at the start of an MCMC run) iterations and number of iterations used in the estimation of the posterior means are optional with a default number of burn-in iterations at 50 and default number of iterations used for the estimation of the posterior means at 500. We will keep these constant throughout the tests.

The conditions under which the cases were simulated are as described below:

Model and data conditions	
Data spacing	Equally spaced
Number of simulations	100
Length of time series	100
Varying parameters	
Error	$\epsilon \sim N(0, \sigma^2 \Sigma)$, $\sigma = 0.35, 0.50, 0.65$
Correlation	$\rho = 0.1, 0.15, 0.2$
Threshold	$h = 0.45, 0.55, 0.65, 0.75, 0.85$

Table 3.9: Simulation conditions

A fixed threshold value of 0.85 will be used with variable standard deviations and correlation coefficients.

3.4.3 Test conditions: Binary Segmentation Method

For BSM, Rebecca Killick (2011) developed a method which uses two separate test statistics to determine if a change point has occurred for both single and multiple changepoints.

The model structures will be the exact same as those tested by LLR & B&H with the same variances and correlation structures.

We will vary the penalty variable where $pen = \gamma * \log(n)$, allowing the parameter γ to take the values 5, 5.5, 6, 6.5, 7. The choice of these values are arbitrary, however they allow the test to return reasonable power while the significance lies below our threshold.

3.5 Conclusions and Results

3.5.1 Local Linear Regression

Overall this test performs very well, especially when a change in mean is present. The test perform especially well when a seasonal trend is present as the smooths are able to account for the pattern, however in a large number of cases the number of false positives may pose a problem especially when correlation is present. As the correlation increases, the test is effectively useless since as the number of false positives is so high. The tests size and power increases rapidly as the correlation coefficient increases. For example in case 1, size is around 0.00-0.10 with a correlation coefficient of 0 however with a correlation coefficient of 0.2 this increases to around 0.30. The presence of seasonal trend also affects the power of the test, however, this increase in power does not mean the test performs better under these conditions as the size also increases . Size remains under 0.10 with a correlation coefficient of 0 however this increases with a seasonal trend to around 0.60-0.70 in some cases.

When a change in trend is present, the power of the test is not very high. However when both a change in trend and mean are present the power within each distinct case is higher than the respective power for a change in mean only.

With seasonal trend, a much smaller kernel bandwidth is needed so that the smooth lines can follow the data closer, however, the presence of this trend effects the size so much with correlation present that to apply this to real data there must be no correlation or the seasonal trend must be removed.

3.5.2 Barry & Hartigan algorithm

A summary of results can be found in table 3.10. Barry & Hartigan's algorithm works very well when seasonal trend is not present. With a simple change in mean, the power is very high (80-90%) with standard deviation equal to 0.35. With larger standard deviation, the power rapidly decreases, for example with a simple change in mean with a standard deviation equal to 0.65 the power reduces to 0.50; however, the size always remains around 0 even with larger correlation coefficients.

With both a change in mean and change in slope and mean, as the threshold increases the power decreases and the size stays constant at 0. For example, with both a change in mean and slope the power reduces from 0.86 at a threshold value of 0.45 down to 0.27 at a threshold of 0.85 (standard deviation equal to 0.65 in both cases).

With seasonal trend present, size increases under all conditions to the point that the test cannot determine whether there is a difference between non present and present change points as the power and size are almost identical, especially when seasonal amplitude is increased. When seasonal trend is present and the threshold decreases, the power and size both decrease at roughly the same rate.

To use this test effectively, seasonal trend must be removed so that only the change point is present along with overall slope and random variation. Under these conditions, the test works very well.

Change in mean			
Threshold	σ_i^2	Size	Power
0.45	σ_1^2	0	0.95
0.45	σ_2^2	0	0.81
0.45	σ_3^2	0	0.50
0.55	σ_1^2	0	0.97
0.55	σ_2^2	0	0.70
0.55	σ_3^2	0	0.36
0.65	σ_1^2	0	0.80
0.65	σ_2^2	0	0.52
0.65	σ_3^2	0	0.23
0.75	σ_1^2	0	0.64
0.75	σ_2^2	0	0.37
0.75	σ_3^2	0	0.13
0.85	σ_1^2	0	0.53
0.85	σ_2^2	0	0.25
0.85	σ_3^2	0	0.09

Change in slope and mean			
Threshold	σ_i^2	Size	Power
0.45	σ_1^2	0	1.00
0.45	σ_2^2	0	0.98
0.45	σ_3^2	0	0.86
0.55	σ_1^2	0	1.00
0.55	σ_2^2	0	0.86
0.55	σ_3^2	0	0.67
0.65	σ_1^2	0	0.95
0.65	σ_2^2	0	0.79
0.65	σ_3^2	0	0.55
0.75	σ_1^2	0	0.94
0.75	σ_2^2	0	0.70
0.75	σ_3^2	0	0.41
0.85	σ_1^2	0	0.88
0.85	σ_2^2	0	0.53
0.85	σ_3^2	0	0.27

Change in slope and mean (seasonal trend)			
Threshold	σ_i^2	Size	Power
0.45	σ_1^2	1.00	1.00
0.45	σ_2^2	1.00	1.00
0.45	σ_3^2	1.00	1.00
0.55	σ_1^2	1.00	1.00
0.55	σ_2^2	1.00	1.00
0.55	σ_3^2	1.00	0.98
0.65	σ_1^2	0.99	0.98
0.65	σ_2^2	0.99	0.98
0.65	σ_3^2	0.98	0.96
0.75	σ_1^2	0.95	0.93
0.75	σ_2^2	0.91	0.97
0.75	σ_3^2	0.95	0.92
0.85	σ_1^2	0.80	0.75
0.85	σ_2^2	0.75	0.81
0.85	σ_3^2	0.79	0.73

Change in slope and mean (seasonal trend – varying coefficient)			
Threshold	σ_i^2	Size	Power
0.45	σ_1^2	1.00	1.00
0.45	σ_2^2	1.00	1.00
0.45	σ_3^2	1.00	1.00
0.55	σ_1^2	1.00	1.00
0.55	σ_2^2	1.00	1.00
0.55	σ_3^2	0.98	1.00
0.65	σ_1^2	1.00	1.00
0.65	σ_2^2	0.98	1.00
0.65	σ_3^2	0.98	1.00
0.75	σ_1^2	1.00	1.00
0.75	σ_2^2	0.96	1.00
0.75	σ_3^2	0.90	1.00
0.85	σ_1^2	0.90	1.00
0.85	σ_2^2	0.83	1.00
0.85	σ_3^2	0.66	1.00

Table 3.10: Summary of Barry and Hartigan algorithm results (all correlations set to 0.1)

3.5.3 Binary Segmentation Method

A summary of results for BSM with correlations set to 0.1 can be seen in table 3.11. This method works extremely well under all conditions. Variance, correlation and the presence of seasonal trend do not affect the power or size in any of the cases. Size stayed at 0 throughout each simulation and power stayed at 1 in all cases apart from one. The only case that this method did not perform well was within case 2 where no change was detected when in fact there was a change in the slope parameter.

However, as this study simply identifies whether at least one change point has been detected and not where the change point has been detected. Therefore there is a possibility that the test is quite sensitive and a larger penalty parameter may need to be used.

Another problem is that the penalty parameter chosen is arbitrary. Therefore, the chances of detecting a changepoint when it is actually present is all down to the penalty chosen. If it is too small, it will pick up the smallest changes. If it is too large, it will not pick up even large changes. These results are therefore good based on the penalty chosen.

Change in mean			
Penalty	σ_i^2	Size	Power
5	σ_1^2	0.00	1.00
5	σ_2^2	0.00	1.00
5	σ_3^2	0.00	1.00
5.5	σ_1^2	0.00	1.00
5.5	σ_2^2	0.00	1.00
5.5	σ_3^2	0.00	1.00
6	σ_1^2	0.00	1.00
6	σ_2^2	0.00	1.00
6	σ_3^2	0.00	1.00
6.5	σ_1^2	0.00	1.00
6.5	σ_2^2	0.00	1.00
6.5	σ_3^2	0.00	1.00
7	σ_1^2	0.00	1.00
7	σ_2^2	0.00	1.00
7	σ_3^2	0.00	1.00

Change in slope and mean			
Penalty	σ_i^2	Size	Power
5	σ_1^2	0.00	1.00
5	σ_2^2	0.00	1.00
5	σ_3^2	0.00	1.00
5.5	σ_1^2	0.00	1.00
5.5	σ_2^2	0.00	1.00
5.5	σ_3^2	0.00	1.00
6	σ_1^2	0.00	1.00
6	σ_2^2	0.00	1.00
6	σ_3^2	0.00	1.00
6.5	σ_1^2	0.00	1.00
6.5	σ_2^2	0.00	1.00
6.5	σ_3^2	0.00	1.00
7	σ_1^2	0.00	1.00
7	σ_2^2	0.00	1.00
7	σ_3^2	0.00	1.00

Change in slope and mean (seasonal trend)			
Penalty	σ_i^2	Size	Power
5	σ_1^2	0.00	1.00
5	σ_2^2	0.00	1.00
5	σ_3^2	0.00	1.00
5.5	σ_1^2	0.00	1.00
5.5	σ_2^2	0.00	1.00
5.5	σ_3^2	0.00	1.00
6	σ_1^2	0.00	1.00
6	σ_2^2	0.00	1.00
6	σ_3^2	0.00	1.00
6.5	σ_1^2	0.00	1.00
6.5	σ_2^2	0.00	1.00
6.5	σ_3^2	0.00	1.00
7	σ_1^2	0.00	1.00
7	σ_2^2	0.00	1.00
7	σ_3^2	0.00	1.00

Change in slope and mean (seasonal trend – varying coefficient)			
Penalty	σ_i^2	Size	Power
5	σ_1^2	0.00	1.00
5	σ_2^2	0.00	1.00
5	σ_3^2	0.00	1.00
5.5	σ_1^2	0.00	1.00
5.5	σ_2^2	0.00	1.00
5.5	σ_3^2	0.00	1.00
6	σ_1^2	0.00	1.00
6	σ_2^2	0.00	1.00
6	σ_3^2	0.00	1.00
6.5	σ_1^2	0.00	1.00
6.5	σ_2^2	0.00	1.00
6.5	σ_3^2	0.00	1.00
7	σ_1^2	0.00	1.00
7	σ_2^2	0.00	1.00
7	σ_3^2	0.00	1.00

Table 3.11: Summary of BSM algorithm results (all correlations set to 0.1)

3.5.4 Summary of results

To conclude, all three tests work well under different conditions. LLR works well when a mean change is present and is affected by both correlation and by seasonal trend, especially when the amplitude of the trend is increased. When correlation is even slightly higher than 0, the size increases drastically. Given a series with correlation, the modified test which can account for the correlation coefficient should be used.

B&H works extremely well when seasonal trend is not present however the size increases to almost 1 when seasonal trend is present. It would therefore make sense to remove seasonal trend before applying this test to the data.

BSM works very well even when seasonal trend is present; however when there is no mean change then impacts cannot be detected. A simple change in slope parameter was not picked up at all (case 2) and therefore this test should only be used when there is a change in mean.

As a recommendation, B&H should not be used when seasonal trend is present. When there is a simple change of mean within series, LLR should be used and this test also works reasonably well when seasonal trend is present (when seasonal trend is present a smaller kernel bandwidth should be used). BSM seems to work well under all conditions however as noted before it is possible that the test is sensitive with the penalty parameter used in this simulation study.

Referring back to our simulation results, it makes logical sense to remove the seasonal pattern from each series as both LLR and B&H are seriously affected by the presence of seasonal trend.

To remove seasonal trend, a difference series will be calculated (raw series minus the seasonal trend modelled by harmonics) by taking the original series and taking away a harmonic function of decimal year.

3.6 Applications

The three approaches will now be applied to real life data. Firstly, we will apply the three tests in full to the EMEP data. The sites AT02 in Austria and GB02 in Great Britain will be used. Each method will be applied in turn and then comparing the results by assessing if changepoints that have been detected by each method are at a similar position.

A changepoint analysis will then be carried out on the Whitelee data, however the full analysis will not be shown. A summary of results will be produced and each method can be compared.

3.6.1 EMEP Network

The EMEP network data was taken from EMEP's (<http://www.emep.int/>) and can be accessed publically. For both AT02 and GB02, each data point is the log weekly SO₂ which was calculated by taking the log of the mean of each week. This smoothed out the seasonal pattern within the data.

Autocorrelation was checked by producing ACF plots. No significant autocorrelation was found within either series.

3.6.1.1 AT02: Illmitz

AT02 is a site located in Illmitz which is an area at the south eastern tip of Austria, 117 meters above sea level. Log SO₂ data has been taken here and is plotted within figure 3.2, both with super imposed estimated values (replaced NA values) and a superimposed local linear regression line. A small amount of variability has been added to each data point, drawn from a normal distribution with standard deviation representative of the rest of the series. The data are daily.

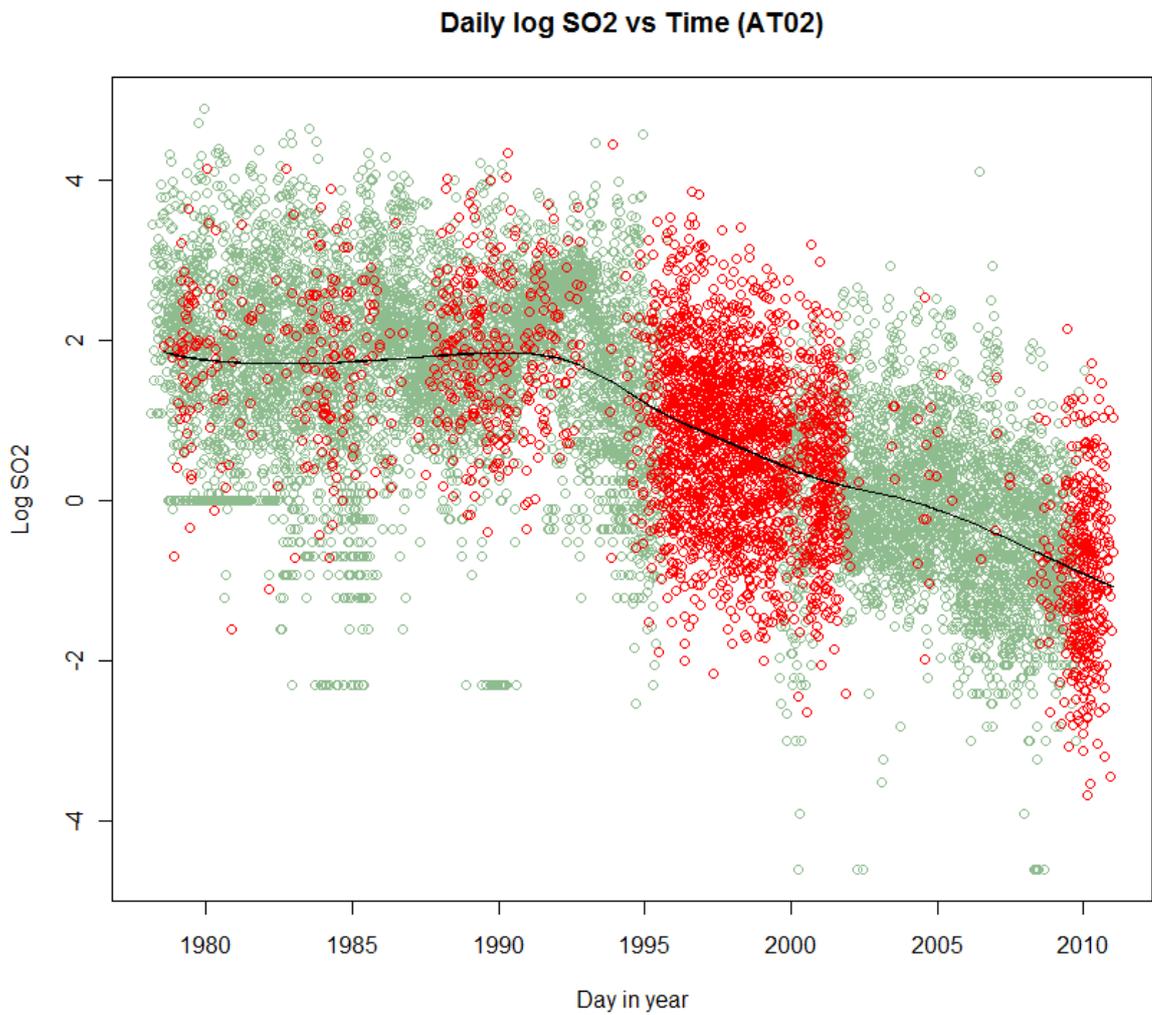


Figure 3.2: Log SO₂ versus Time at station AT02, superimposed local linear regression line and superimposed estimated value (red).

To decrease day to day variability, the log mean weekly average will be taken of data allowing seasonal trend to still be visible but variance to decrease. Shown in figure 3.3 is the log average SO₂ data versus weeks in year.

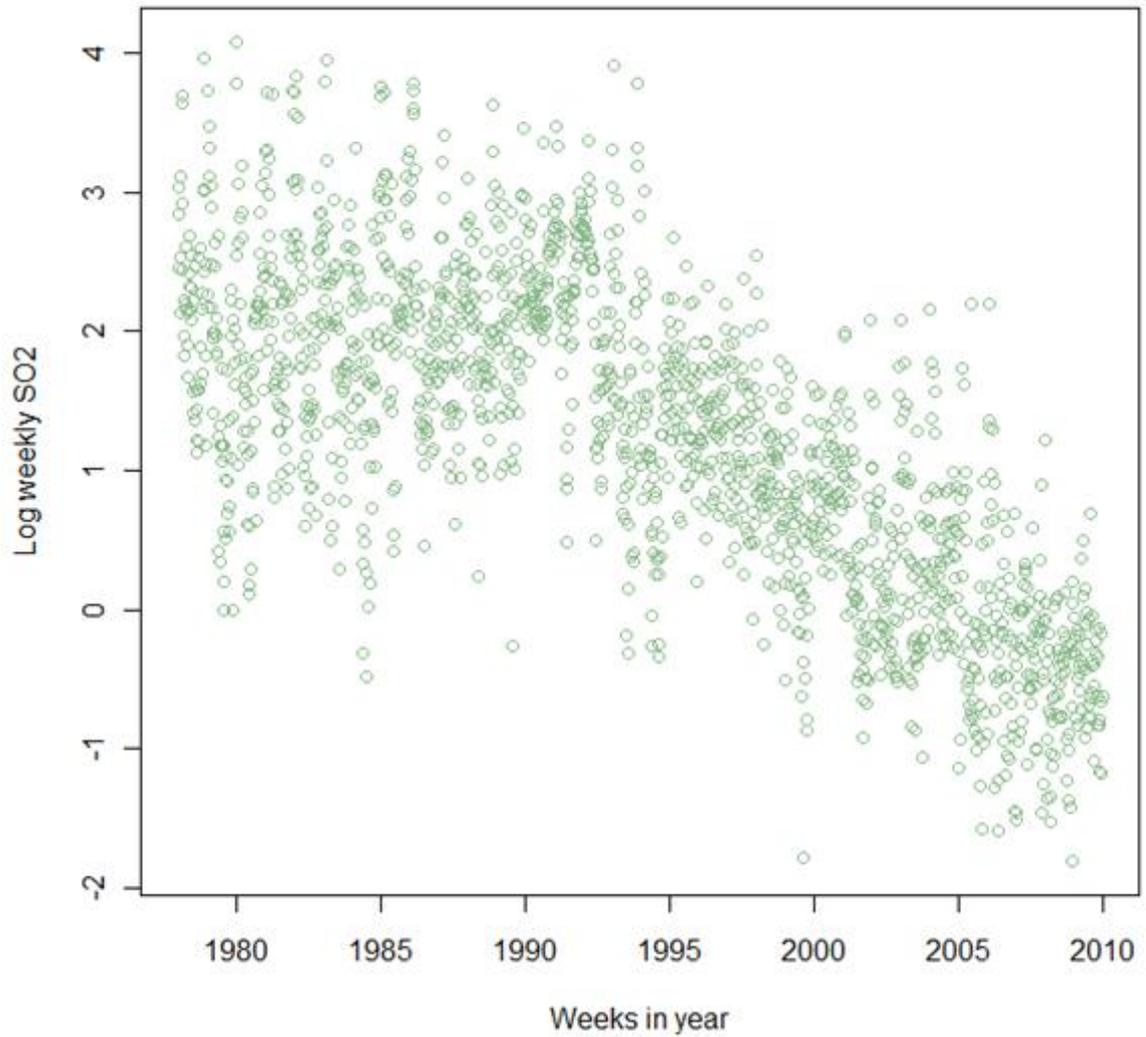


Figure 3.3: Log weekly average SO2 versus Weeks in year

We can now apply the three techniques to the data shown above.

For local linear regression, a number of bandwidths will be chosen to observe the effects on the number of flagged discontinuities. For the Bayesian method a number of threshold probabilities will also be chosen to observe the effects on the number of flagged discontinuities.

Local linear regression with weights: Kernel bandwidth $h = 2$

Firstly, we will assess the time series using local linear regression with kernel bandwidth (standard deviation) equal to 2, as determined by the degrees of freedom.

Decimal year	1980.772	1989.578	1992.974	1996.082	1999.247	2003.015	2005.850	2005.885
Jump	-7.00	3.07	-10.05	12.19	-5.56	2.67	-4.80	3.92

Table 3.12: Location and size of jumps

Where the left and right smooths leave the shaded area in figure 3.4 a, suggests the possible presence of a discontinuity. Red vertical lines indicate where discontinuities were detected and the precise location of these change points can be found in table 3.12. The analysis was on a logarithmic scale, and therefore the jumps shown within figure 3.4 are given on a \log_e

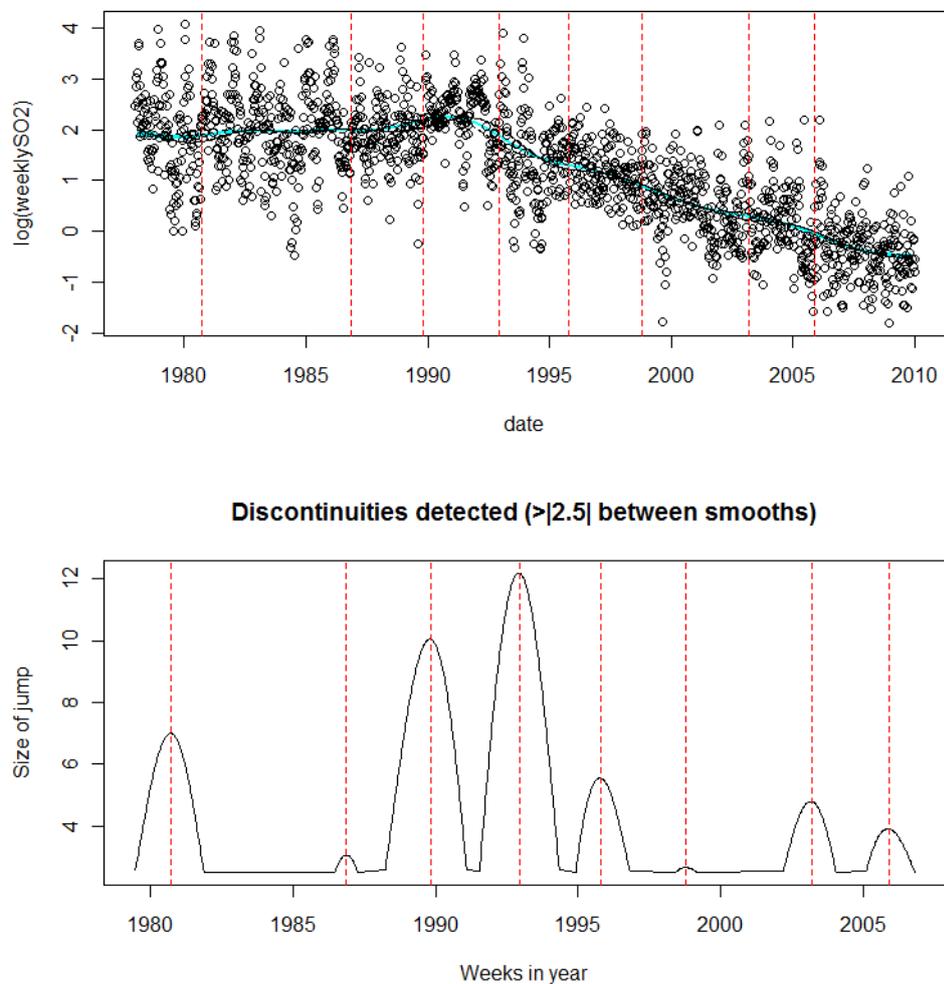


Figure 3.4 a,b: Log weekly SO₂ versus Weeks in year including location of jumps

scale. The magnitude each of the points can also be found in table 3.12 but are shown visually within figure 3.4b (absolute magnitude has been taken and plotted). Left and right smooths may leave the shaded confidence area for large periods of time however discontinuities are taken to be the maximum difference between the two smooths, or the turning points within figure 3.4b.

Figure 3.5 helps us to visually interpret the change points and furthermore, to estimate the individual non-parametric trend between discontinuities. Local linear regression with weights was used for this however the kernel bandwidth was increased to stop the trend picking up seasonal variation.

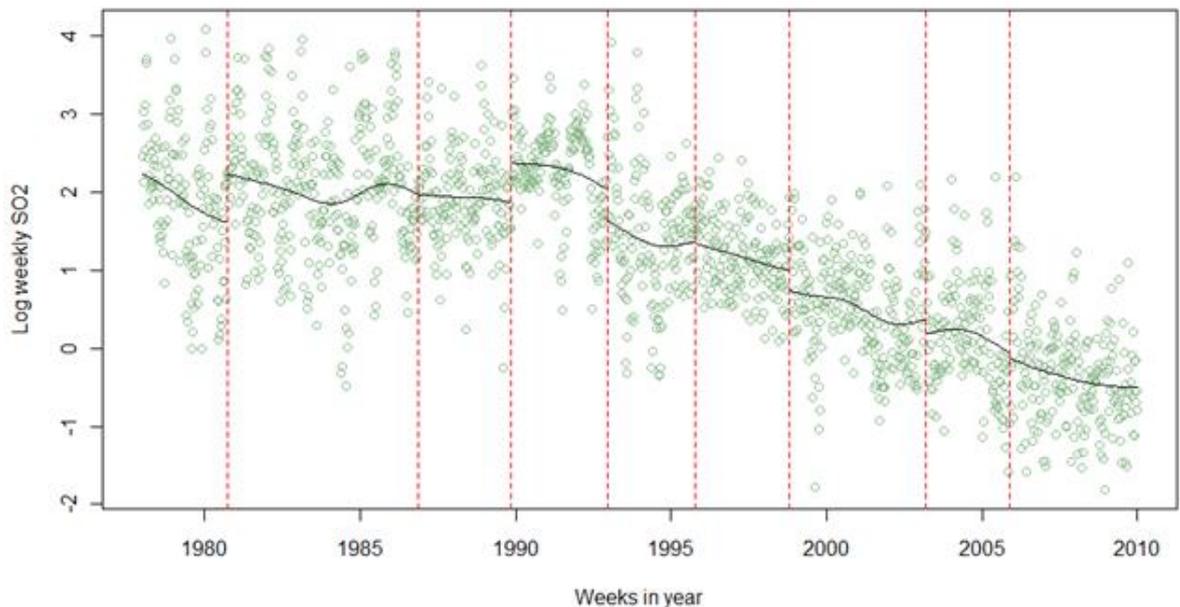


Figure 3.5: Log weekly SO₂ versus Weeks in year including change point locations and non-parametric curves.

Table 3.12 shows the detected change points along with the respective size of jump between smooths. Jumps 1, 3 and 4 at 1992.974 (week 50) and 1996.082 (week 4) have the largest jumps of 7, 10.05 and 12.19 respectively. All other jumps are less than 5 in absolute magnitude.

Since the size of the kernel bandwidth is relatively low in this case, many discontinuities that have been flagged may not actually be change points and seasonal variation may actually affect the testing procedure. Increasing the size of the bandwidth h , will allow our smoothing function to fit a straighter curve and will not flag a point as a discontinuity when in fact it is seasonal variation.

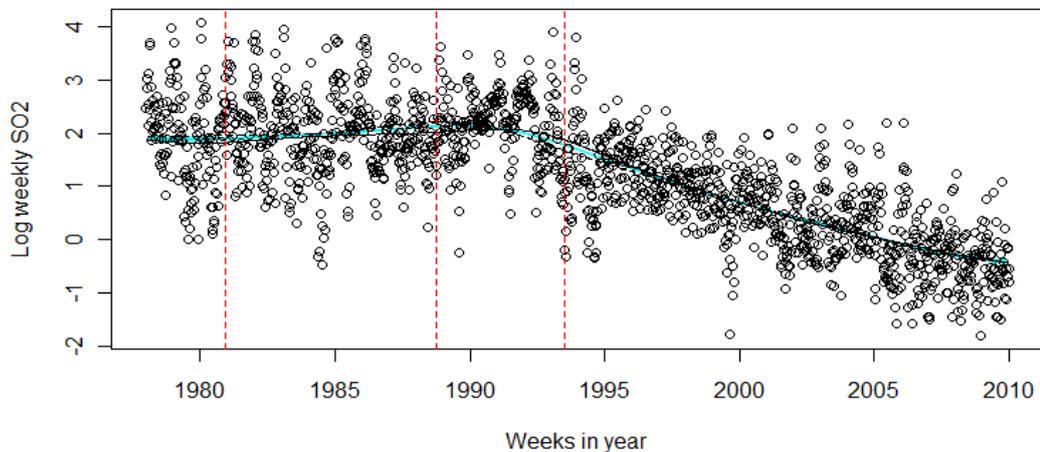
Local linear regression with weights: Kernel bandwidth $h = 4$

By increasing the kernel bandwidth such that more data points are taken into consideration when determining the smoothing function, the number of detected change points has

Decimal year	1980.926	1988.753	1993.530
Jump	-4.75	-9.65	10.29

Table 3.13: Location and size of jumps

decreased from 8 to 3. The 3 change points roughly correspond with the first three change points within the first analysis where $h = 2$. Figure 3.6a shows our plotted Log weekly SO₂ with confidence bands and detected change points and figure 3.6b allows us to visually



Discontinuities detected ($>|2.5|$ between smooths)

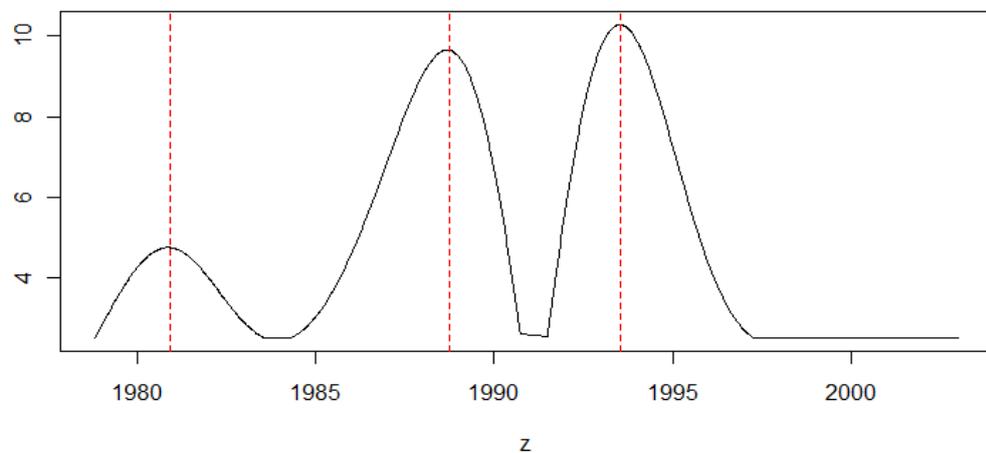


Figure 3.6a, b: Log weekly SO₂ versus Weeks in year including location of jumps. Absolute value of differences between left and right smooths.

interpret the size of these changes. Table 3.13 shows the location of these jumps along with their size. Figure 3.7 is a plot of log weekly SO₂ versus weeks in year with superimposed change point locations as well as superimposed curves between change points.

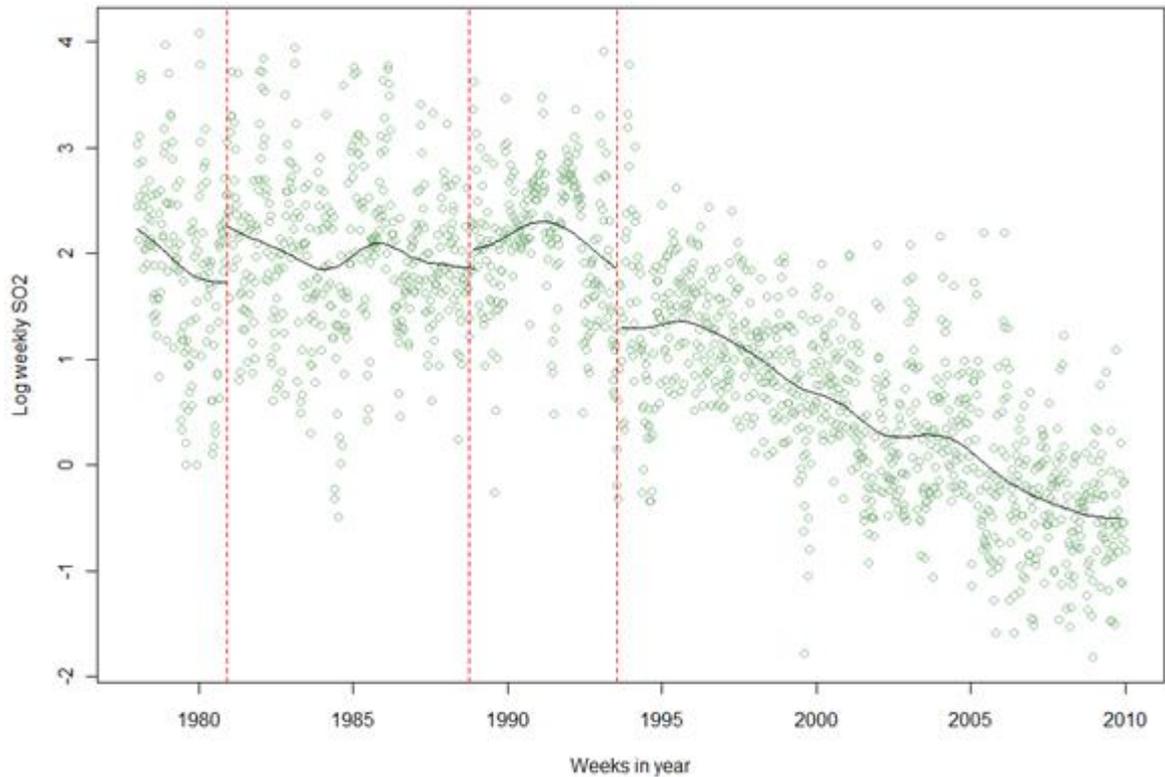


Figure 3.7: Log weekly average SO₂ versus Weeks in year including change point locations and non-parametric curves.

From figure 3.7 we can see the three change points situated at the 48th week in 1980, the 39th week in 1988 and the 28th week in 1993. The mean increases from the end of the first parametric line through to the second and then increases again from the second to the third. After change point 3 at week 28 in 1993 the mean decreases and the non-parametric curve shows a negative slope.

These estimated change points can now be compared with estimated change points as detected by Barry and Hartigans algorithm.

Bayesian change point analysis: Barry and Hartigan algorithm

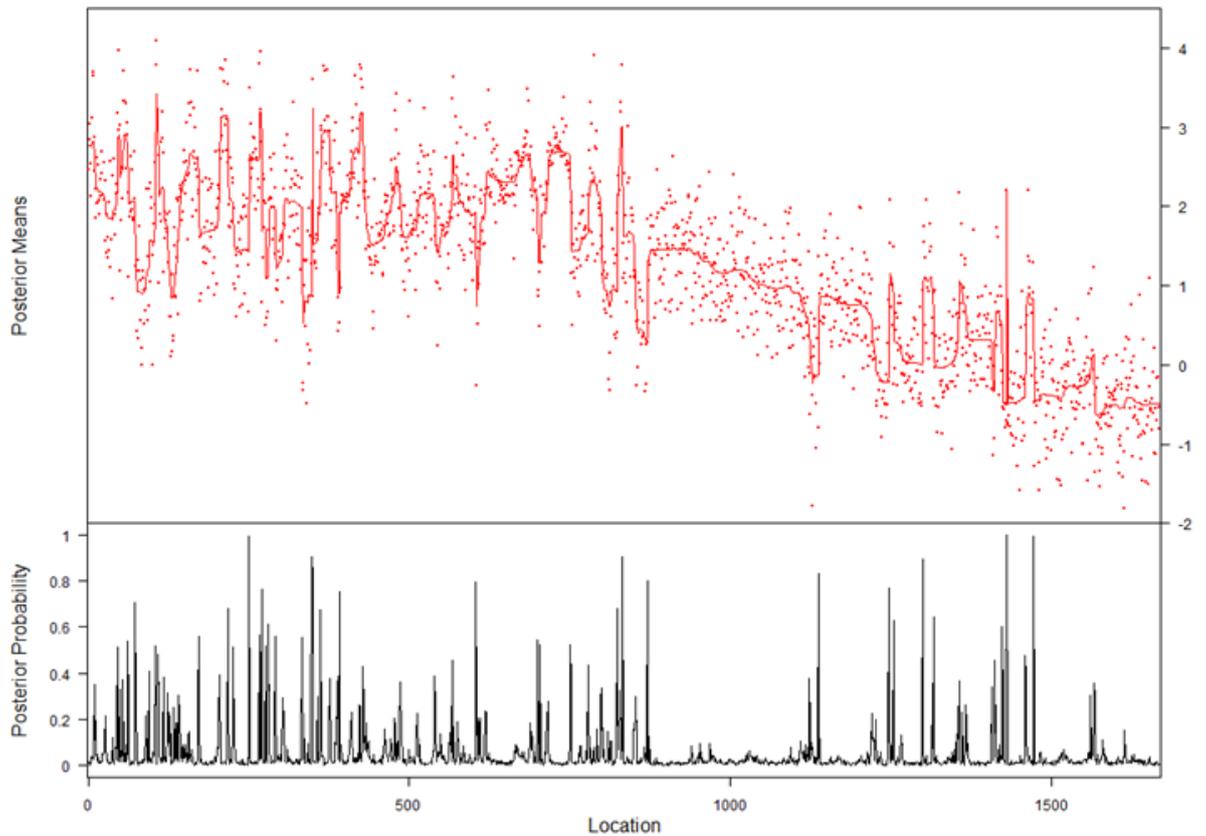


Figure 3.8: Posterior means and Probabilities of change as calculated by Barry and Hartigan's algorithm

Under the exact same conditions in the previous analysis, Barry and Hartigan's algorithm was used on the log weekly mean SO₂ from the EMEP website. The hyper parameters γ and λ are both set to their default value of 0.2 (chosen by Barry and Hartigan). The number of burn in iterations is set to 50 and the number of iterations after burn in is set to 500 (default values).

Figure 3.8 allows a visual interpretation of the posterior means plotted over time as well as the locations respective probabilities of a change point occurring at that exact location. A number of threshold probabilities are shown below in table 3.14 which allows us to determine the number of change points, given an arbitrary threshold value.

Table 3.14: Probabilities of change with corresponding locations at various thresholds

Threshold		CP1	CP2	CP3	CP4	CP5	CP6	CP7
0.95	Probability	0.992	0.964	0.978				
	Location	1982.777	1984.657	2006.240				
0.90	Probability	0.992	0.964	0.900	0.912	0.914	0.978	
	Location	1982.777	1984.657	1993.942	2005.415	2005.434	2006.240	
0.80	Probability	0.992	0.964	0.866	0.900	0.912	0.914	0.978
	Location	1982.777	1984.657	1984.676	1993.942	2005.415	2005.434	2006.240

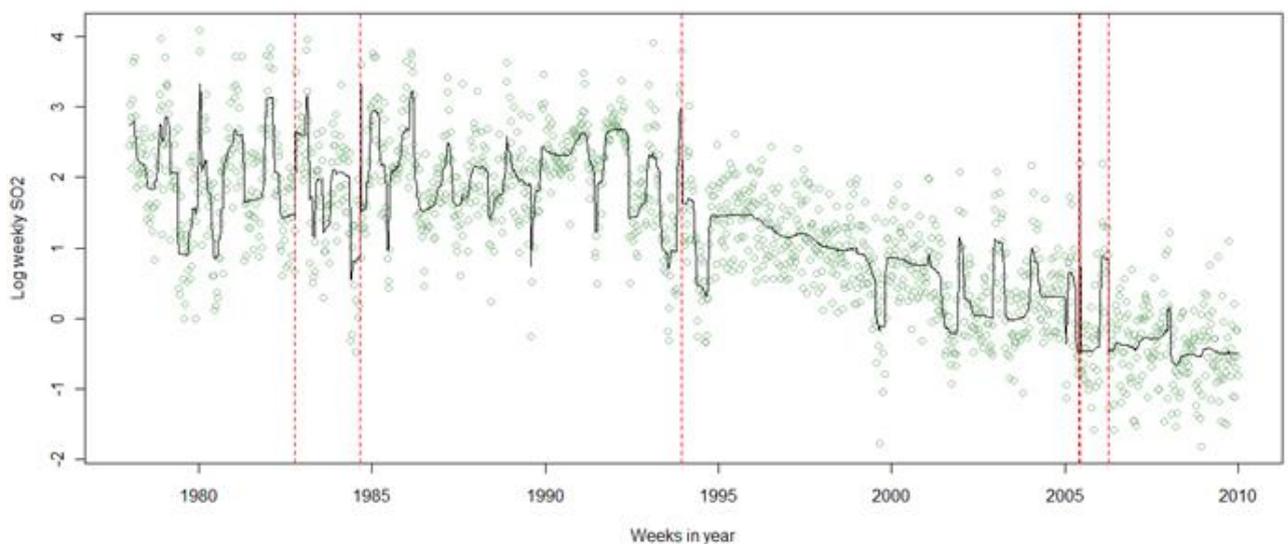


Figure 3.9: Log weekly SO₂ versus weeks in year with superimposed change point locations ($p > 0.90$).

It is clear from table 3.14 that as the threshold probability drops the number of change points increases. The largest 3 change points are located at week 12 in 2006, week 40 in 1982 and week 34 in 1984. Change points may also occurred at weeks 21 and 22 in 2005, week 12 in 2006 and week 49 in 1993.

Binary Segmentation Method

BSM can now be applied to the time series, which can then be compared with the two previous methods.

Binary Segmentation Method					
Location (datapoint number)	1054.0000	1424.0000	798.0000	1221.0000	393.0000
Test statistic	635.0151	112.1629	102.8227	25.75492	23.67609
Optimal Changepoints	3				
Penalty	92.74975				

Table 3.15: Changepoints detected by BSM

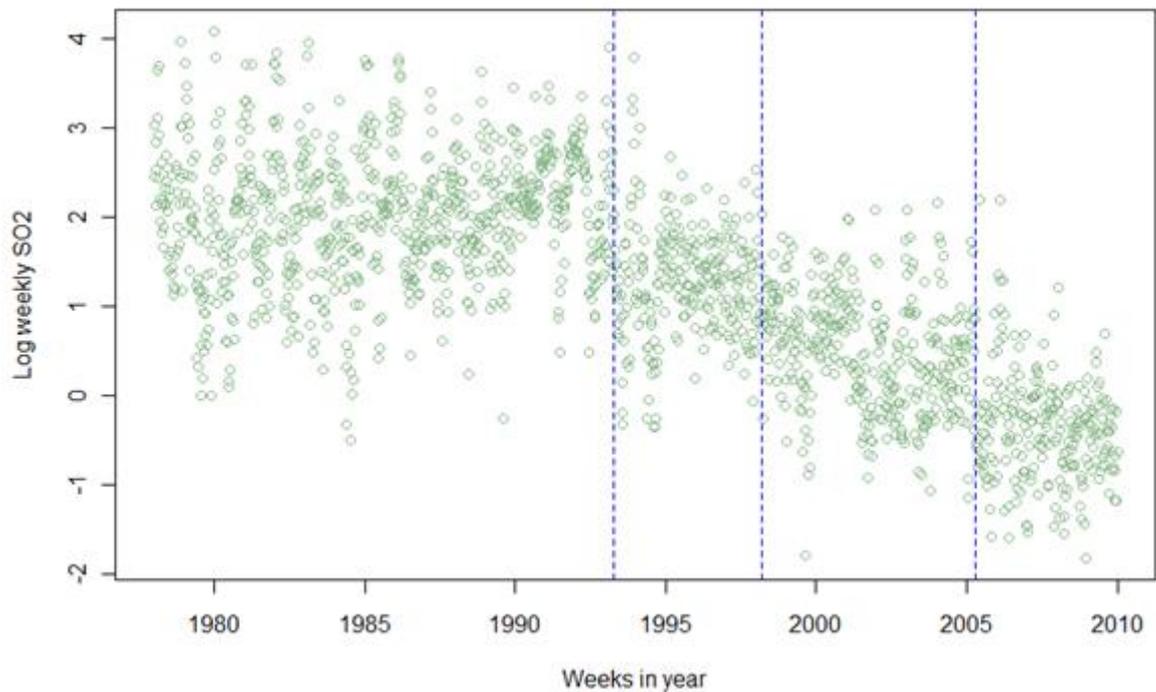


Figure 3.10: Daily log average SO₂ versus weeks in year with super imposed changepoint locations.

Three changepoints were detected while using the arbitrary penalty equal to $\lambda * \log(n)$ where λ is equal to 10.5. One was located at 1998.201 (week 10), one at 2005.3 (week 15) and 1993.29 (week 15).

Comparing LLR, B&H and BSM: AT02

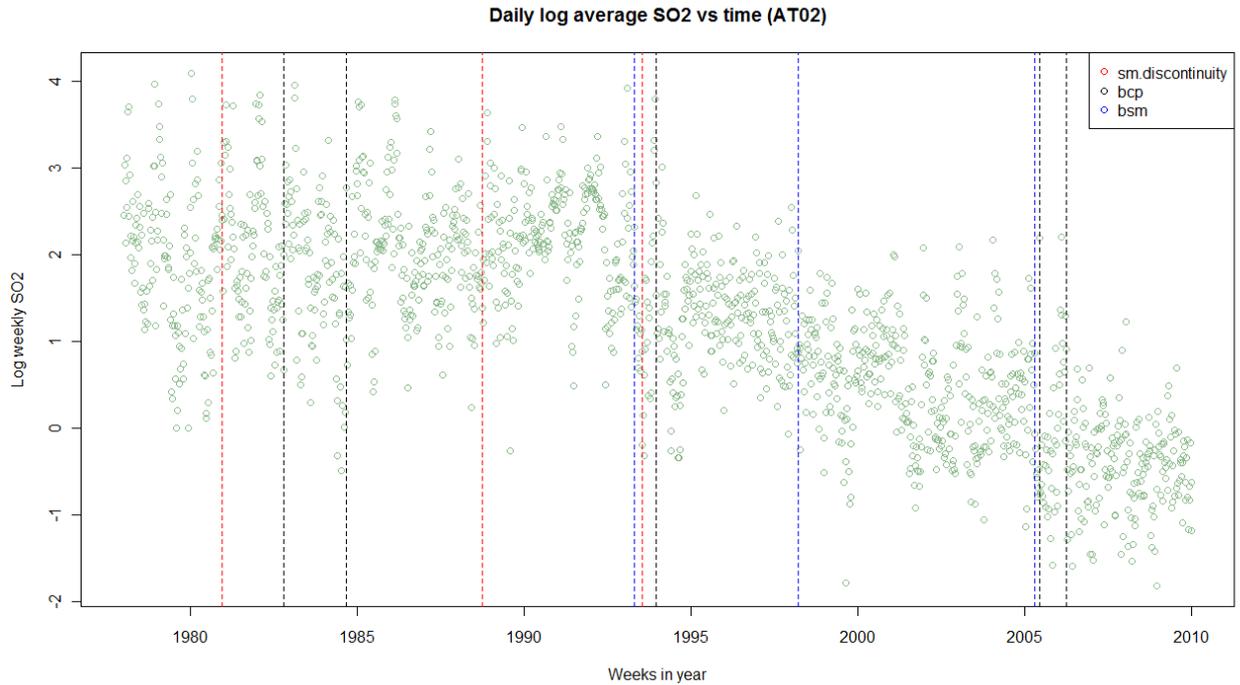


Figure 3.11: Log weekly SO₂ versus weeks in year with superimposed change points

Our final figures and plots, figure 3.11 and table 3.16 allow us to interpret the discontinuities detected by each method side by side. Vertical lines in black are discontinuities indicated by Barry and Hartigans algorithm (bcp), vertical lines in red are those as indicated by local linear regression with weights (sm.discontinuity) and blue by binary segmentation method (bsm). Of the 11 discontinuities shown in total (two discontinuities as indicated by B&H have been merged since they were one week apart) five are from B&H, 3 are from local linear regression and 3 by BSM.

Table 3.16: Location of change points

LLR	1980.926	1988.753	1993.530		
B&H	1982.777	1984.657	1993.942	2005.424	2006.240
BSM			1993.29	1998.201	2005.3

LLR estimates a change at 1980.926 (week 48) whereas B&H estimates a change 1.85 years later at 1982.777 (week 40). This could possibly be the same point as LLR's change point is estimated to be the maximum difference between left and right smooths but this may not

necessarily be the exact location - figure 3.6 indicates a difference of over |2.5| overlapping the same region as B&H's indication. Another very close comparison can be made in 1993. BSM indicates a change at 1993.29 (week 15), LLR indicates a change point at 1993.530 (week 27) and B&H indicates a change 0.41 years later at 1993.942 (week 48). Again, this could possibly be the same change point as figure 10 indicates a difference in smooths within this region, overlapping B&H's proposed value.

A noticeable difference between methods is that B&H and BSM indicate changes at 2005 and 2006, whereas LLR does not where the kernel bandwidth is equal to 4. However where the kernel bandwidth is equal to 2, LLR indicates changes at 2003.015 (week 1), 2005.850 (week 44) and 2005.885 (week 46) (refer to table 3.23).

By comparing the change point indicated by B&H at 2005.424 (week 21) with the change point as indicated by LLR ($h=2$) at 2005.850 (week 44) the difference is small at 0.426 years (22 weeks). Referring to table 3.23 is it possible that this is the same point as the plot of absolute differences overlaps B&H's proposed value.

3.6.1.2 GB02: Eskdalemuir

Eskdalemuir is an area located within Dumfries and Galloway, Scotland near the Scottish/English border. As within analysis for AT02, the data plotted below are log daily SO₂ where the red data is predicted data from a local linear regression model with added normal variability representative of the rest of the data. We can therefore define the model used to predict the data as shown below as $\hat{y} = m(x) + \varepsilon_i$ where $\varepsilon_i \sim N(0, 0.788)$.

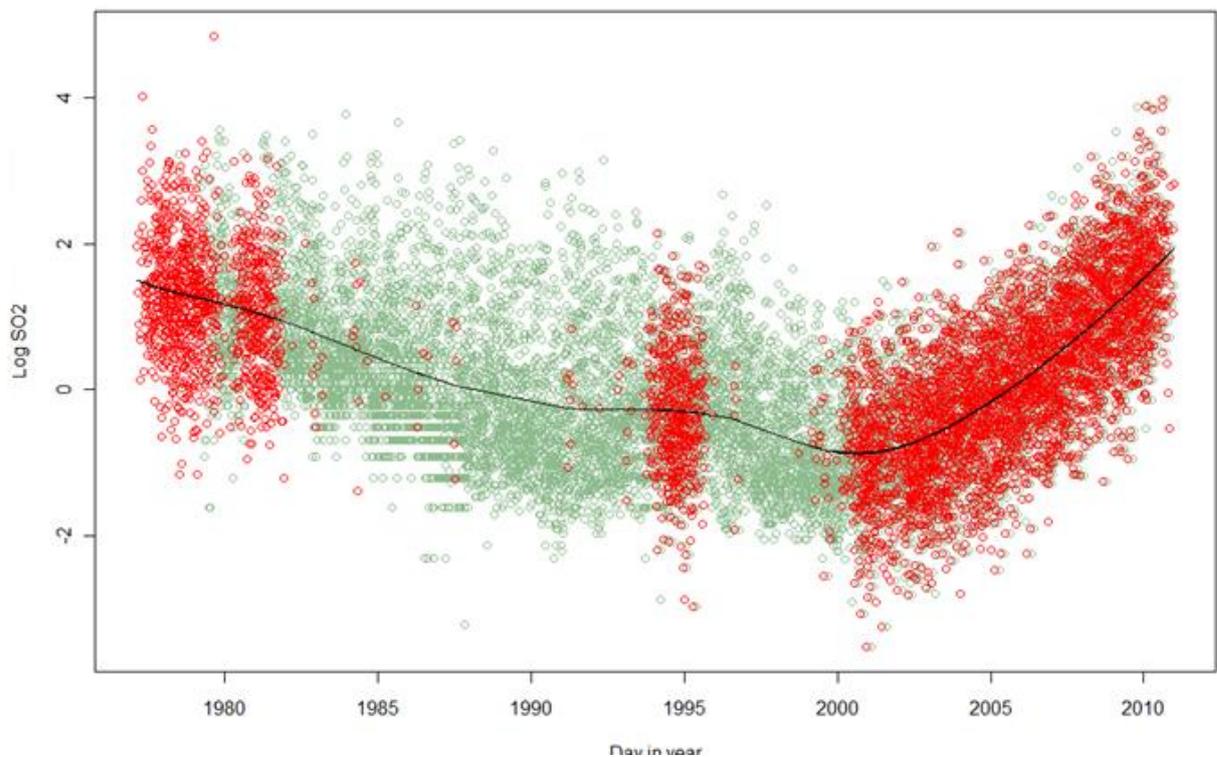


Figure 3.12: Log SO₂ versus Time at station GB02, superimposed local linear regression line and superimposed estimated values (red).

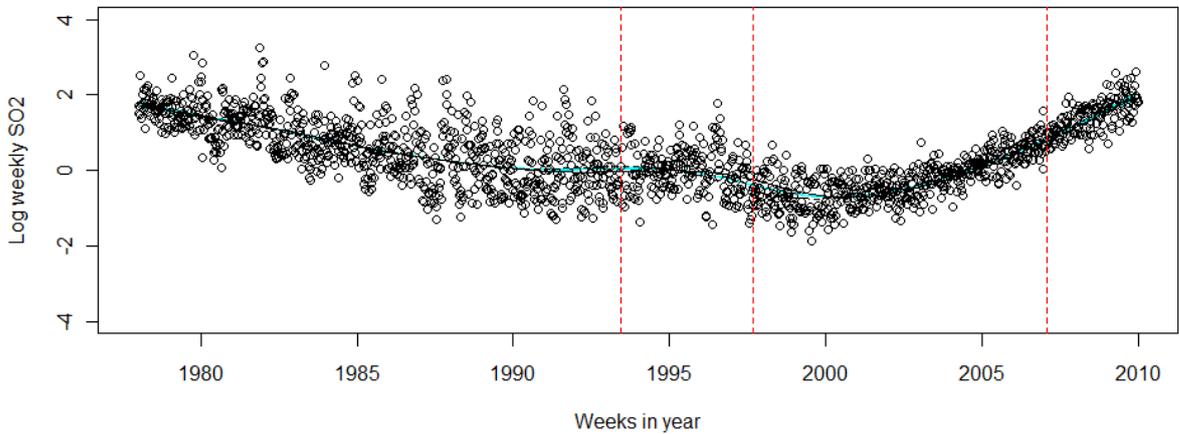
As before, predicted values were calculated on the series within figure 3.12 then weekly averages were calculated afterwards. Both local linear regression, BSM and B&H will be used on the series to detect discontinuities.

Local linear regression with weights: Kernel bandwidth $h = 2$

Table 3.17 shows the location and the size of jump from one stable state to another, this can be seen visually within figure 3.13. Discontinuities were detected at 1993.423 (week 21), 1997.686 (week 35) and 2007.068 (week 3), one other discontinuity was detected at 2003.457 (week 24) however the jump was close to 2.5 and was left out. The kernel bandwidth here is equal to 2, equivalent to two standard deviations.

Table 3.17: Location and size of jumps

Decimal year	1993.423	1997.686	2007.068
Jump	-6.24	13.25	-3.42



Discontinuities detected ($>|2.5|$ between smooths)

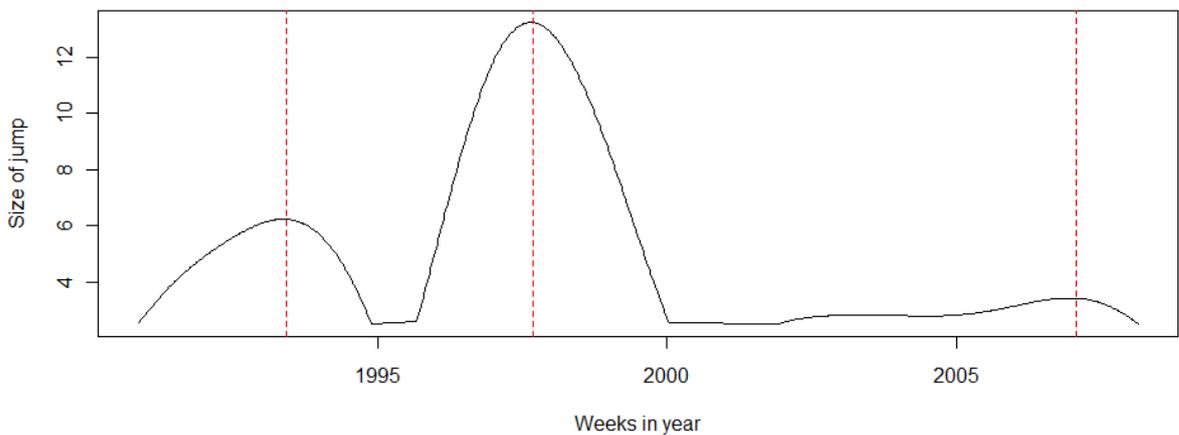


Figure 3.13a, b: Log weekly SO₂ versus Weeks in year including location of jumps. Absolute value of differences between left and right smooths.

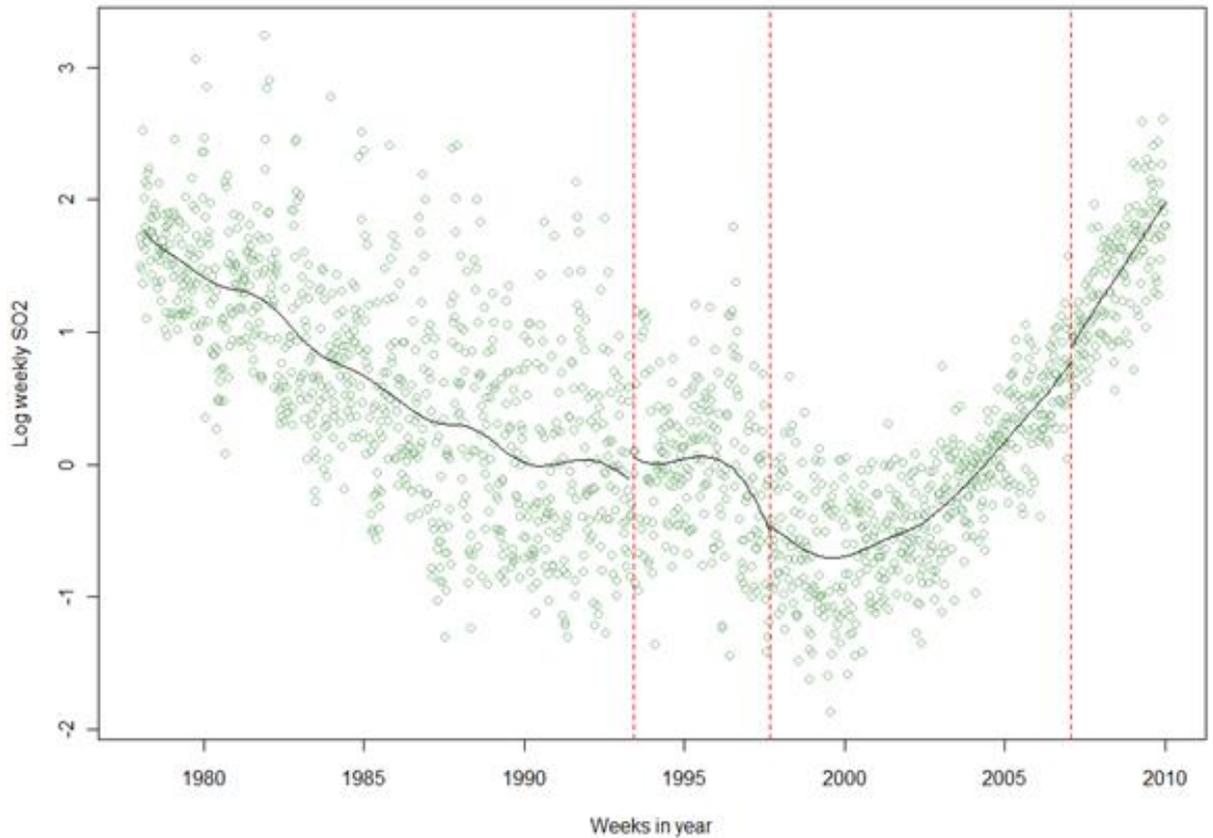


Figure 3.14: Log weekly SO2 versus Weeks in year including change point locations and non-parametric curves.

Log weekly SO2 versus weeks in year can be seen in figure 3.14 along with superimposed change point locations and non-parametric curves between change points (local linear regression, $h=0.8$). The first discontinuity is located at 1993.423, the second at 1997.868 (week 45) and the third at 2007.068 (week 3) although this discontinuity is small in comparison to the first and second.

Until the first change point there is a clear negative trend within the data. Between change points 1 and 2 there is an upward curved trend and there is then a positive trend after the second change point.

Bayesian change point analysis: Barry and Hartigan algorithm

For B&H we will use three threshold probabilities to determine the location of change points within the GB02 series. The location and probabilities of discontinuities can be found in figure 3.15. As with AT02, three threshold probabilities will be taken to determine the number of change points, all of which can be found in table 3.25.

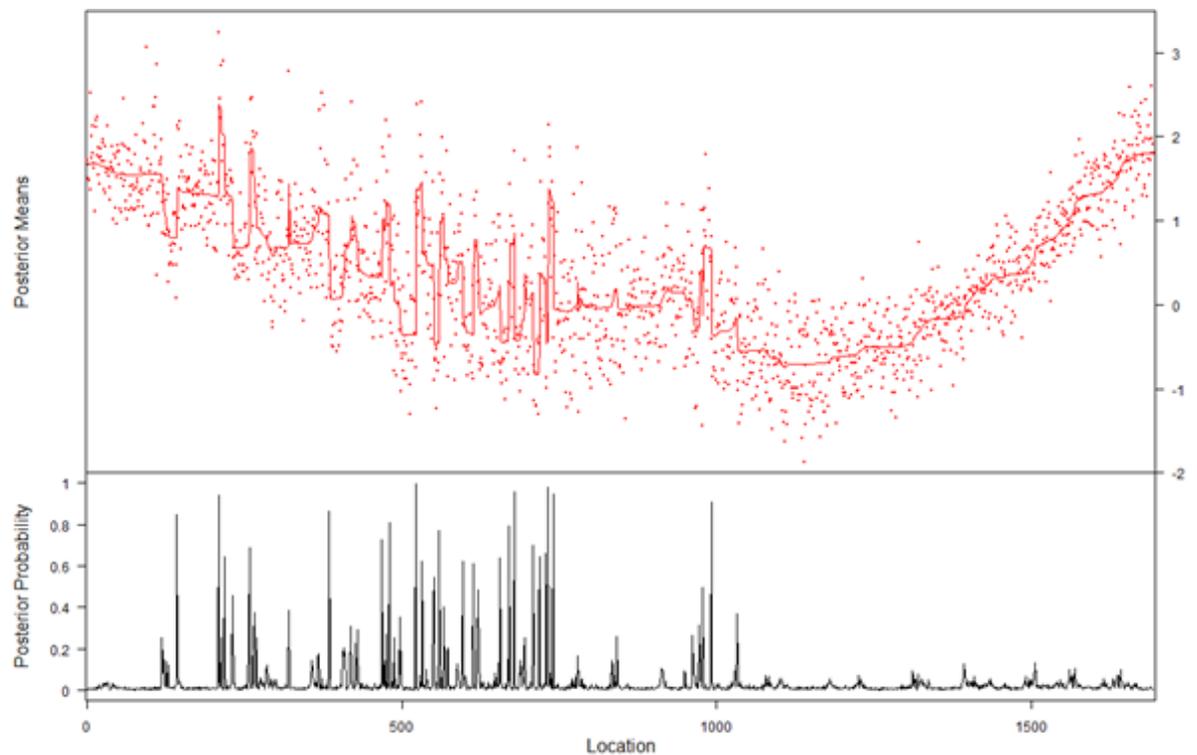


Figure 3.15: Posterior means and Probabilities of change as calculated by Barry and Hartigans algorithm

The majority of large probabilities lie before the first change point as detected by local linear regression and also around where the second change point was detected.

Threshold		CP1	CP2	CP3	CP4
0.95	Probability	0.998	0.996		
	Location	1987.699	1991.608		
0.90	Probability	0.998	0.904	0.996 0	0.924
	Location	1987.699	1991.180	1991.608	1996.448
0.85	Probability	0.998	0.904	0.996 0	0.924
	Location	1987.699	1991.180	1991.608	1996.448

Table 3.18: Probabilities of change with corresponding locations at various thresholds

Within table 3.18 probabilities of discontinuities and their respective locations are contained. Above 0.9, 4 locations were detected. One in 1987, two in 1991 and one in 1996. Above 0.95 only 2 locations were detected which were in 1987 and 1991.

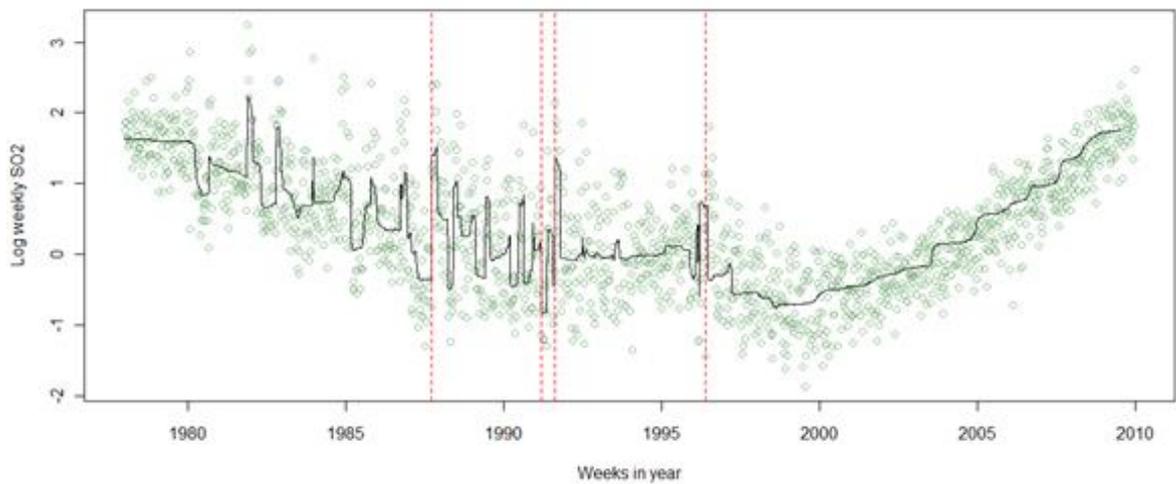


Figure 3.16: Log weekly SO₂ versus weeks in year with superimposed change point locations ($p > 0.90$).

Binary Segmentation Method

The penalty parameter is set to $3.5 * \log(n)$ here and the maximum number of changepoints we can detect will be set to 5.

Binary Segmentation Method					
Location (datapoint number)	385.0000	1504.0000	1592.0000	992.0000	1374.0000
Test statistic	270.2248	270.2248	68.83.177	68.12605	68.1205
Optimal Changepoints	3				
Penalty	92.94298				

Table 3.19: Changepoints detected by BSM

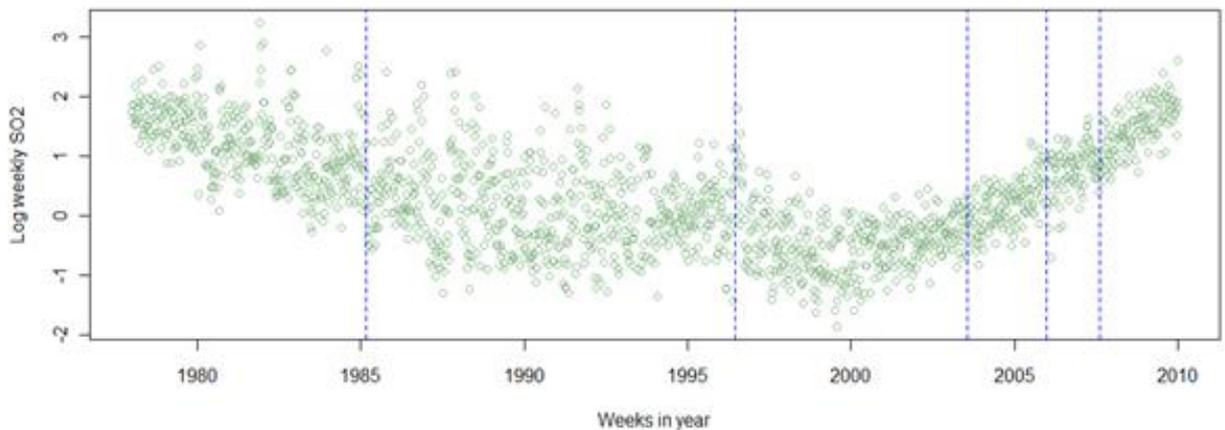


Figure 3.17: Log weekly SO2 vs. weeks in year with superimposed changepoint locations

Three changepoints were detected while using the penalty equal to $3.5 * \log(n)$. One was located at the same point as located by B&H at 2007.395 (week 20), one was located slightly later at 2007.542 (week 28) and one was located at 2009.378 (week 19).

Comparing LLR, B&H & BSM: GB02

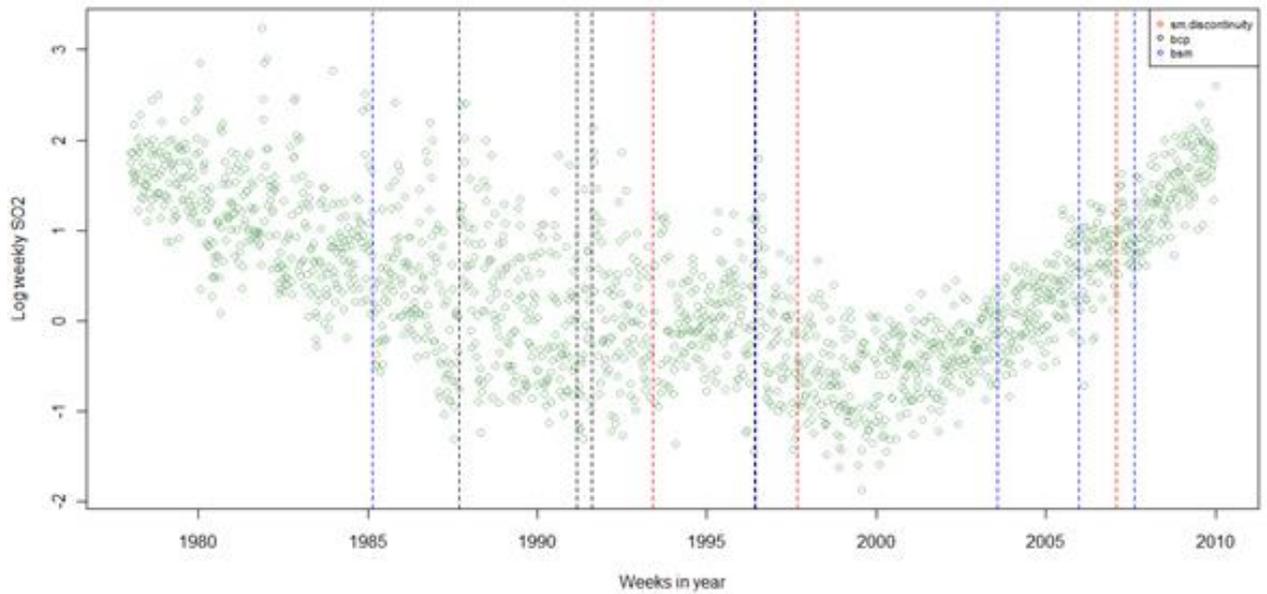


Figure 3.18: Log weekly SO₂ versus weeks in year with superimposed change points

We can now compare positions of change points flagged by LLR, B&H ($p > 0.9$) and BSM. Of the 12 change points, 4 were detected by local Barry and Hartigan's algorithm, 3 were detected by local linear regression with weights and 5 by BSM, however the discontinuity detected in 2007 which is that very last discontinuity shown in figure 3.18, is small.

Table 3.20: Location of change points

LLR	1993.423	1997.686	2007.068		
B&H	1987.699	1991.180	1991.608	1996.448	
BSM	1985.148	1996.448	2003.559	2005.979	2007.617

Of the change points detected by LLR and B&H, three sets are relatively close. B&H detects a change at 1996.448 whereas LLR detects a change at 1997.686 (week 35), only just over a year apart, however BSM detects the same change as B&H at 1996.448 (week 23). Referring to figure 3.13 we can see that the curve's maximum is at 1997.686 (week 35) however it does overlap 1996.448.

Another two change points which are relatively close are those situated at 1991.608 (week 31) and 1991.180 (week 9) detected by B&H and another at 1993.423 (week 21) detected by LLR. Again, referring to figure 3.13 the absolute value of the changes detected show that the turning point is located at 1993.423 (week 21) however the curve does overlap the value detected at 1991.608 (week 31). The change point detected by B&H at 1991.180 (week 9) is also within the limits of this curve.

Three changes are relatively close around 2006, two detected by BSM lie at 2005.979 (week 50) and 2007.617 (week 32) and one detected by LLR lies at 2007.068 (week 3).

3.6.2 Whitelee Windfarm

Discussed in detail in section 1.3.4, Whitelee windfarm is a windfarm situated south of Glasgow and is one of the largest windfarms in Europe.

As with the EMEP analysis shown in the previous section, analysis was carried out on the Whitelee data. Of the 11 sites, 3 sites were chosen for analysis (WL13, WL14, WL1) such that the three changepoints techniques was applied each site. Plots of each variable can be found with a corresponding table of changepoints.

3.6.2.1 Total phosphorus: Comparing WL13, WL14 and WL1

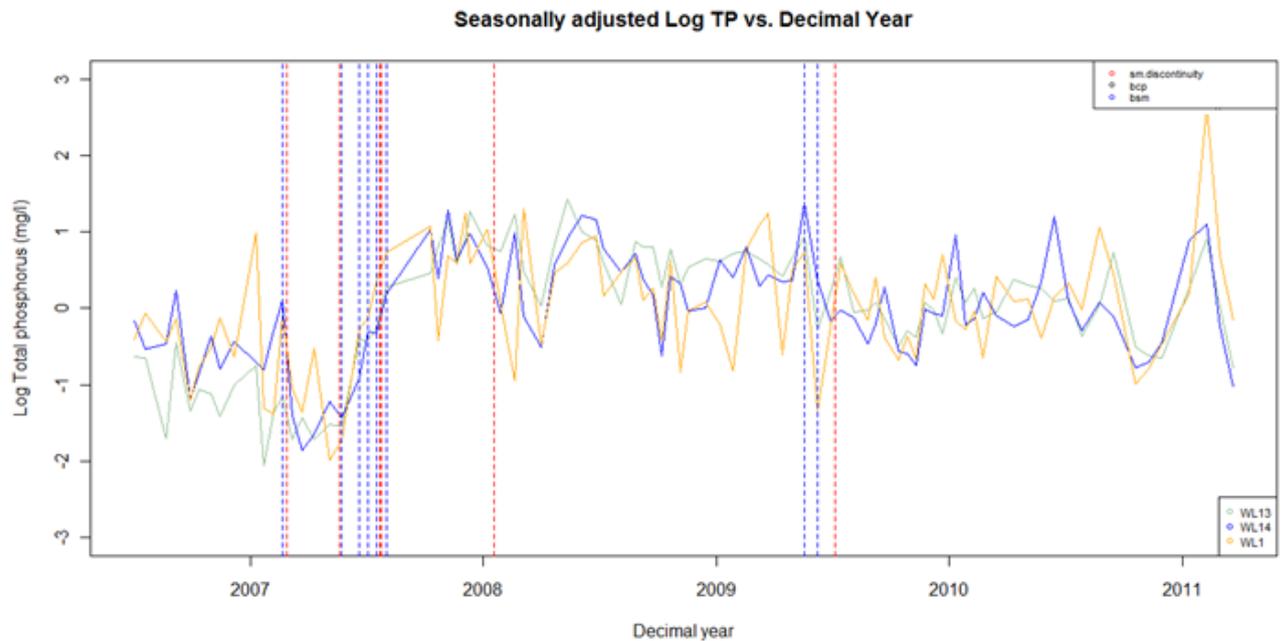


Figure 3.19: Seasonally adjusted log TP vs. decimal year for stations WL13, WL14 and WL1 with superimposed changepoint locations.

Within figure 3.19, seasonally adjusted TP versus decimal year is plotted for the three stations at Whitelee. We also have superimposed changepoint locations which can be found within table 3.21. For WL13, only one change point was detected using LLR which was located at 2007.564 (week 29) and only one changepoint was detected using B&H located at 2007.395 (week 20), which are relatively close. Three changepoints were detected at this station with the use of BSM at 2007.395 (week 20), 2007.542 (week 28) and 2009.378 (week 19). For WL14, 5 changepoints were found with LLR but no changepoints were found with B&H. Four changes were detected using BSM. For WL1, one change point was found at 2007.564 (week 29) with the use of LLR, no change points were detected with B&H and one was detected by BSM.

The most notable result is that at 2007.564 (week 29), changepoints were detected at all three stations with the use of LLR, where the changepoint is described as the maximum absolute difference between smooths. Furthermore, BSM detected a change very close to this value at stations WL13 and WL14, located at 2007.542 (week 28). As this was consistent, it is quite clear that there is a shift in the series at this point in mid-2007, at the time where

the new turbines were being put into place – this has had an effect on the phosphorus levels within the area.

Another notable result is that within WL13, B&H and BSM detected changes at 2007.395 (week 20) and within WL14 LLR detected a change at 2007.385. These two values are relatively close, and by referring to figure 3.19 (week 20) we can deduce that this is around the time where there is an apparent jump in the series.

To summarise, all three stations have jumps around mid-2007, around the time where new turbines were put in place and new roads were being built. It is highly likely that the installation of these turbines affected the phosphorus levels within the area.

Table 3.21: Location of changepoints for stations WL13, WL14 and WL1

WL13	LLR			2007.564		
	B&H		2007.395			
	BSM		2007.395	2007.542		2009.378
WL14	LLR	2007.159	2007.385	2007.564	2008.048	2009.512
	B&H					
	BSM	2007.137	2007.466	2007.542		2009.436
WL1	LLR			2007.564		
	B&H					
	BSM		2007.446			

3.6.2.2 Total Organic Carbon: Comparing WL13, WL14 and WL1

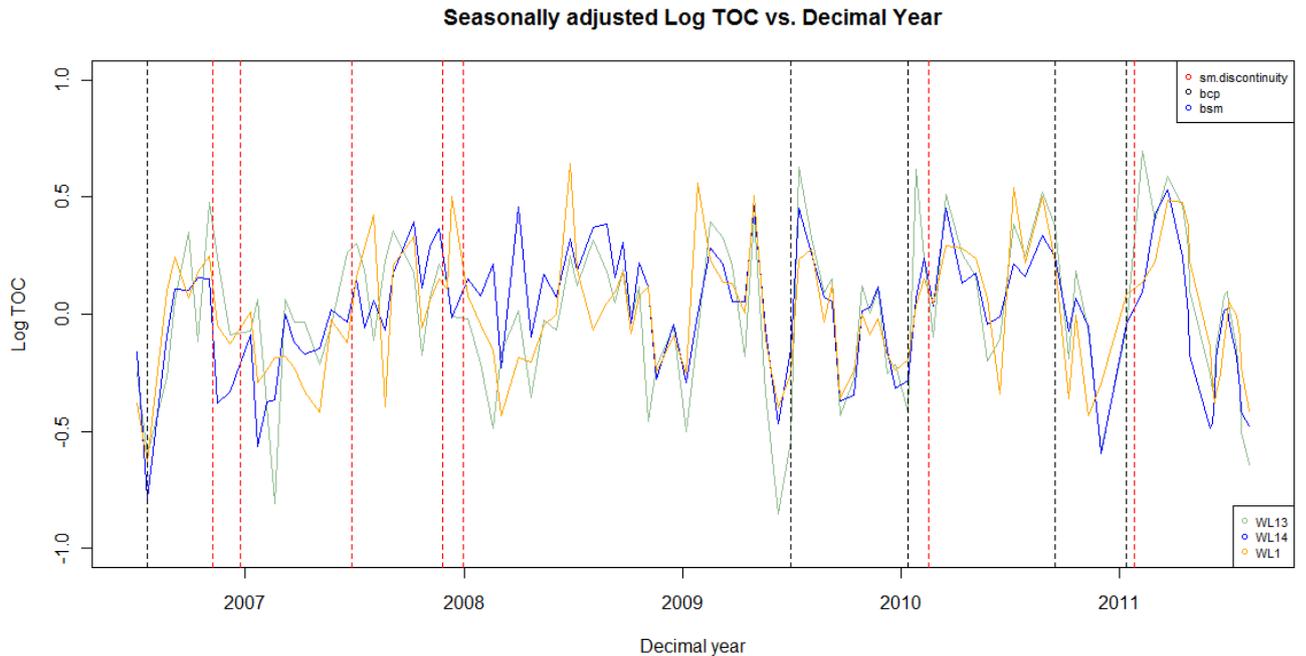


Figure 3.20: Seasonally adjusted log TOC vs. decimal year for stations WL13, WL14 and WL1 with superimposed changepoint locations.

Table 3.22: Location of changepoints for stations WL13, WL14 and WL1

WL13	LLR	2006.852	2007.430	2010.126	2010.126		
	B&H						
	BSM						
WL14	LLR		2007.486	2007.997	2010.115		
	B&H	2006.551				2010.704	
	BSM						
WL1	LLR	2006.977		2007.904			2011.067
	B&H			2009.493	2010.030		2011.030
	BSM						

Overall, 15 changepoints were detected over the three series. 4 were detected within WL13, all by LLR where one was detected in 2006, one in 2007 and two in 2010. 5 (week 26) were detected within WL14, 3 by LLR and 2 by B&H where LLR detected two in 2007 and one in 2010 and B&H detected one in 2006 and one in 2010. Finally 6 were detected within WL1

where 3 were detected by LLR and 3 by B&H, where LLR detected one in 2006, one in 2007 and one in 2011 and B&H detected one in 2009, one in 2010 and one in 2011.

Before 2008, 7 discontinuities were detected, three in 2006 where B&H detected one at 2006.551 (week 28) (WL14) and two at the end of 2006 at 2006.852 (week 44) (WL13) and 2006.977 (WL1). Two discontinuities were detected relatively close to each other at 2007.430 (WL13) and 2007.486 (WL14) – both by LLR which may indicate a change around this time.

Another area which has two close detected change points is at 2007.997 (week 51) (WL14) and 2007.904 (week 47) (WL1), again indicating the possibility of change.

From 2008 to mid-way through 2009, no discontinuities were detected by either LLR or by B&H.

At 2010.030 (WL1) and 2010.115 (WL14) changes were detected by LLR and B&H respectively, which are relatively close. Two other close changes were detected at 2011.030 (week 1)(WL1) and 2011.067 (week 3) (WL1) by B&H and LLR respectively, which were detected within the same series.

To conclude, there are no large clusters of changes within the series which may indicate no changes may be present however a number of changepoints were relatively close to each other, namely two in mid-2007, two at the end of 2007, two at the start of 2010 and two at the start of 2010, indicating possibility of change at these times. However, no changes are expected within 2010 and 2011 so it is probable these are due to random variation.

3.6.2.3 NO2: Comparing WL13, WL14 and WL1

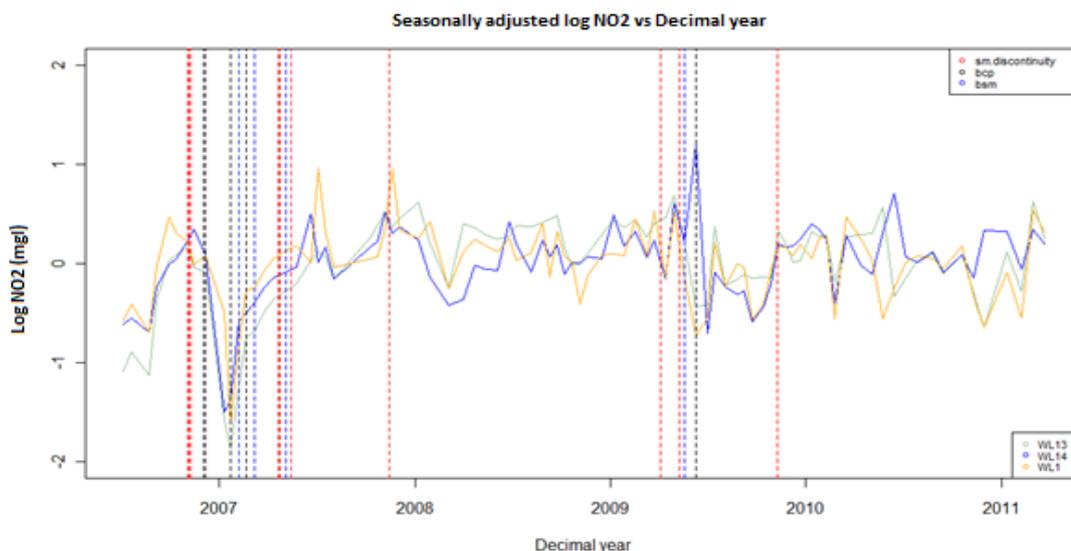


Figure 3.21: Seasonally adjusted log NO2 vs. Decimal year for stations WL13, WL14 and WL1 with superimposed changepoint locations.

Overall 19 discontinuities were detected within the three time series (WL13, WL14, WL1) for NO2, detected by the three methods LLR, B&H and BSM. Ten discontinuities were detected by LLR, 5 by B&H and 4 by BSM. The majority of discontinuities were detected within late 2006 and mid-2007, 13 discontinuities were detected within 0.65 years around 2007. Another cluster of discontinuities lie around mid-2009 where 4 discontinuities were detected within series WL13 and WL14.

Table 3.23: Location of changepoints for stations WL13, WL14 and WL1

WL13	LLR	2006.854	2007.307		2009.352	2009.838	
	B&H	2006.929	2007.137				
	BSM		2007.34		2009.378		
WL14	LLR	2006.852	2007.367	2007.866	2009.255		
	B&H	2006.929	2007.058		2009.436		
	BSM		2007.181				
WL1	LLR	2006.852	2007.307				
	B&H						
	BSM		2007.099				

From both figures 3.21 and table 3.23, there is evidence to suggest that a discontinuity occurred around 2007, and possibly another in mid-2009.

3.7 Conclusions

The simulation study carried out allowed us to determine the strength of each of the three tests under certain conditions. Various scenarios were constructed with differing changes in mean, trend, seasonal patterns, correlation and variance. The size and power of each test was then calculated by simulating data 100 times under the same unique conditions and determining the number of times that a changepoint was detected when no change point was present and when a changepoint was present respectively.

Once the three tests were analysed we could determine how they would perform given the data we have. Since two of the three tests were affected by seasonal trend, seasonal trend was removed within the Whitelee data by modelling the data with harmonic functions. The tests were then used on the difference series (original minus trend). The EMEP data did not have seasonal trend and therefore tests were used on the raw series.

Local linear regression was used to estimate missing values within both the EMEP and the Whitelee series using a small kernel bandwidth allowing us to follow the trend of each series quite closely. Random variation (Gaussian) was added onto the trend since all the data is normally distributed. The standard deviation of the random component was calculated by taking the standard deviation of the difference series between the estimates for local linear regression and the raw data.

For the EMEP data, the three tests assessed sites AT02 and GB02. For AT02, 11 discontinuities were found in total. Of the 11, there were two clusters of detected changepoints (both containing 3 changepoints each), one cluster in 1993 and one from mid-2005 to 2006. For site GB02 there are 3 clusters of detected changepoints. One cluster is around mid-2006 to mid-2007 (three detected within this range). Another cluster is located in 1991 (one at 1991.18 (week 9) and one at 1991.6 (week 31)) with another changepoint relatively close at 1993. One more cluster lies around 2006 with one change located at 2005.9 (week 46), one at 2007.1 (week 5) and another at 2007.6 (week 31).

For the Total Phosphorus data at Whitelee, two clusters of changepoints are present. One large cluster of 12 changepoints were detected in 2007, two at the start of 2007 and ten in the middle of 2007. Another cluster of 3 changepoints are in mid-2009.

Total organic carbon at Whitelee did not seem to have any clusters of changepoints but rather a spread of changepoints throughout the series. The only two close changepoints were located within 2011, one at 2011.067 (week 3) and one at 2011.030 (week 1).

Nitrate oxide at Whitelee had two large clusters of detected changepoints within the series, one around 2007 and one around mid-2009. 13 detected changepoints were found around 2007, 5 late 2006 and 8 early 2007. 4 changepoints were located around mid-2009 and a further 1 in late 2009.

To conclude, for the EMEP data detected changepoints were relatively spread out with a few clusters of detected changes in the series. For the Whitelee series, there were clusters of changepoints in both the Total Phosphorus series and the Nitrate Oxide series, but not for the Total Organic Carbon series, however from preliminary analysis this was to be expected. This similarity in the location of the changepoints indicates that it is likely that the detected changes at these clusters are very likely to truly be changepoints rather than the methods variation.

Chapter 4

Modelling the trends and changes in determinands at Whitelee

4.1 Introduction

The purpose of this chapter is to model our time series data with the use of a number of modelling techniques. Many modelling techniques are available to do this, parametric, non-parametric or a combination of both.

Information gained from previous chapters can be used within our modelling process to help build our models, making them a better fit to the data. The use of an *indicator* function for example to indicate a changepoint, used within chapter 2 for our BACI modelling and adjusted within chapter 3 using changepoint analysis, can be used to help our models move more freely.

Models built within this chapter will be compared and contrasted with the use of various criteria, allowing us to adapt and refine them.

These models will be built logically with the use of this previously gained information, including parameters that we know are likely to add value to our models, rather than building from the ground up.

Several modelling techniques may be considered while analysing the various time series of the three variables TP, TOC and NO₂. One of the techniques which will be considered is GAM modelling. GAM's are made of parametric terms and non-parametric smoothing terms. This allows for parametric characteristics with the benefit of smooth terms – GAM's are flexible and effective for conducting nonlinear regression.

4.1.1 Generalised Additive Models

For generalised additive models, we assume $\mu_i = E(Y_i)$ and $Y_i \sim$ some exponential family distribution (in this case, they are normally distributed). Y_i can be regarded as our response variable, X is our model matrix for any parametric components of our model and β the corresponding parameter vector. f_i' s are regarded as our smooth functions of covariates x_k . Various smoothing functions can be used and the *basis* can be chosen, where the *basis* is defined as the type of smoothing function we use. Within our analysis, we will use regression splines and cubic regression splines, the latter when representing seasonal variation.

Representing a smooth function via regression splines requires that f be represented in such a way that it becomes a linear model, whereas representing a smooth via a cubic spline is made up of sections of cubic polynomial to make a curve. These sections that the sections meet are called *knots* and must be chosen, typically evenly spaced through the range of x values.

The number of *knots* chosen essentially allows us to control how smooth we would like to model our data. The higher the number of *knots*, the smoother our model will become. Ideally we would like our spline estimate of f defined as \hat{f} to be as close to f as possible and thus choose a smoothing parameter that allows us to do this.

Essentially a GAM is a generalized linear model with a linear predictor including a sum of smooth functions of covariates, where the general model structure can be explained by:

$$y_i = \beta X_i + \sum_{k=1}^m f_k(x_{ki}) + \epsilon_i \quad (4.1)$$

Where β are the parametric coefficients, X_i is a row of the model matrix for any strictly parametric model components, $f_k(x_{ki})$'s are the smooth terms and there are m of these. ϵ_i 's are the errors, which in our case are normally distributed.

However the model may also include interaction terms and factors and be in the form:

$$y_{ij} = \beta X_i + \text{Factor}_j + \sum_{k=1}^m f_k(x_{ki}) + \sum_{r=1}^l f_r(x_{ri}, x_{(r+1)i}) + \epsilon_i \quad (4.2)$$

Where we have l bivariate terms and a factor with j levels.

Coefficients can also vary with one another such that the model is specified in the form, for example:

$$y_{ij} = \beta X_i + \text{Factor}_j + f_1(x_{1i}) \cdot \text{Factor}_j + f_2(x_{2i}) + f_3(x_{1i}, x_{2i}) + f_4(x_{3i}) \cdot x_{4i} + \epsilon_i \quad (4.3)$$

Which would allow $f_1(x_{1i})$ to interact with Factor_j . For instance, if $f_1(x_{1i})$ was a smooth for trend, then allowing $f_1(x_{1i})$ to vary with Factor_j would mean that for each j there would be as many different smooth trends as there is levels for Factor_j . The term $f_4(x_{3i}) \cdot x_{4i}$ allows for a varying coefficient where x_4 varies along x_3 (which could be, for example, time). If needed, smooth components can also be included as random effects.

Essentially GAM models are a combination of both Generalised Linear Models and Additive Models which allow for properties of both. GLM models are flexible generalisations of linear models such that the response variable have distributions other than the normal distribution. Additive Models (AM) are nonparametric, using a one dimensional smoother to build a restricted class of nonparametric regression models.

GAM models retain these properties; the models allow the user to specify a distribution other than the normal distribution and also a link function $g(\cdot)$ relating the expected value of the distribution to the predictor variables if our errors are not normal. The functions $f_j(x_i)$ may be fit both parametric and non-parametrically and therefore GAMs provide potential for better fits than other techniques.

We can use several criterion for model selection, two criteria which can be used for GAM models are the Generalised Cross Validation Score and the Akaike Information Criterion. The *generalised cross validation score* (GCV) is a reasonable approach which was first introduced by Craven and Wahba (1979). This criterion allows us to select the smoothness parameter and compare GCV scores between models.

The GCV score can be described as:

$$GCV = \frac{1}{n} \sum_{i=1}^n \left(\frac{y_i - \hat{f}_i}{1 - \frac{1}{n} \text{tr}(H)} \right)^2 \quad (4.4)$$

Where H is the hat matrix where $H = X(X^T X)^{-1} X^T$, X is our matrix of explanatory variables (and contains the basis of the smooths) and \hat{f} is the estimate of y_i from fitting all the data (Wood (2006)).

The Akaike Information Criterion (AIC) may also be used. The AIC is described as:

$$AIC = 2k - 2\ln(L) \tag{4.5}$$

Where k is the number of parameters and L is the maximised log likelihood function for the estimated model.

Comparisons of models can be judged by both AIC values and GCV scores. Any models comparisons where these values are close, an ANOVA based on an F-test can be used as a deciding factor

In a less general sense, our model has normal errors and we know from previous analysis that changepoints are present within certain series. For series with a changepoint, we can include an *indicator* function. The indicator function is a binary categorical variable which changes from 0 to 1 after the changepoint. This coupled with some informal analysis will allow us to build models logically rather than from the ground up.

All models were built in the statistical software R. All models were fitted using the *mgcv()* package, which has functions for all of the models listed below.

4.1.2 Varying Coefficient Models

A further issue with modelling environmental data is to determine whether the underlying seasonal pattern changes in structure generally and smoothly over time, or whether the structure of our model changes after some point in time. This requires some further flexibility of our statistical model to account for the change in trend or seasonal pattern for example.

Models that are additive in the regressors but allow the coefficients to change smoothly with the value of other variables are called Varying Coefficient Models. The concept of allowing coefficients to vary as a function was covered within section 4.1.1 (GAM models) and even though GAMs can contain varying coefficients, varying coefficient models in this specific section are based on allowing variable to vary with time and or space. Some examples of using varying coefficient models to model time series data include Chen and Tsay (1993) and Cai, Fan and Li (2000).

We consider models which are linear in the regressors but their coefficients allowed to smoothly change with the value of another variable. Suppose we have a random variable Y whos distribution depends on parameter η as well as predictors X_1, X_2, \dots, X_p and $R_1, R_2 \dots, R_p$. We can specify our varying coefficient model in the form

$$\eta = \beta_0 + X_1\beta_1(R_1) + \dots + X_p\beta_p(R_1) \quad (4.6)$$

This model says that $R_1, R_2 \dots, R_p$ change the coefficients of X_1, X_2, \dots, X_n through functions $\beta_1(), \beta_2() \dots, \beta_p()$. In some cases the variables R_j are indistinguishable from the variables X_j and in other cases might be a special variable such as 'time' (Hastie et al. (1993)).

4.1.3 Generalised Additive Mixed Models

A different approach to estimation and inference with Generalised Additive Models is based on representing GAM's as mixed models. This can be used where factors within the model, in our case 'Site', are a random subset from a larger set of values and therefore should not be treated as fixed. A GAMM is a Generalised Linear Mixed Model in which part of the linear predictor is specified in terms of smooth functions of covariates and if for example the terms within a GAMM were to be linear, the GAMM would reduce to a GLMM.

Suppose we have an outcome variable y_i which has length n , p covariates such that $x_i = (1, x_{i1}, \dots, x_{ip})^T$ associated with fixed effects and a vector b of length q containing covariates z_i associated with random effects. We can then specify a GAMM as:

$$y_i = \beta_0 + f_1(x_{1i}) + \dots + f_p(x_{ip}) + z_i^T b + \epsilon_i \quad (4.7)$$

Where each y_i is assumed to be conditionally independent.

Further explanation of the above model can be found in Penheiro & Bates (2000) and Ruppert et al. (2003).

A key feature of this model is that additive nonparametric functions are used to model covariate effects and random effects are used to model correlation between observations. The term which could be included as random is Site. This is because our variable Site does not include every Site within Whitelee's windfarm, it is a subset of a larger number of sites. Our modelling strategy within this section is to logically build a model given our prior knowledge from previous chapters. This model may or may not include seasonally varying coefficients.

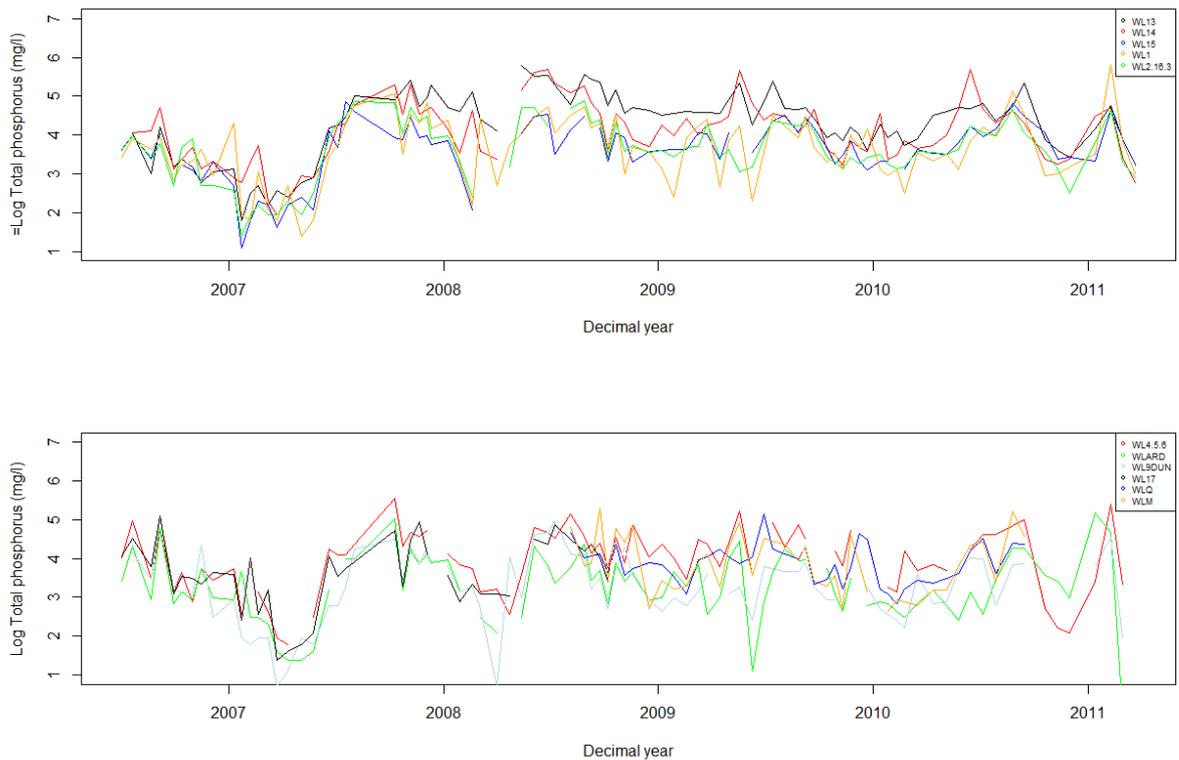
Both generalised additive models and varying coefficient models can be seen as special cases of generalised additive mixed models.

This type of model will not be considered within our analysis because neither the AIC or the GCV can be calculated for this type of model. This means we cannot compare the mixed model directly with the GAM or the VCM.

4.2 Results

4.2.1 Total Phosphorus: Modelling

Within figure 4.1 we can see the overall trend over the years at all 11 sites. There is some variation within the sites although the general overall trend seems consistent such that there is a dip mid-2007 and perhaps a slight negative trend after this point.



Figures 4.1: Time series of Log Total phosphorus

Figure 4.2 shows boxplots of log TP for each month. There does seem to be a seasonal trend which dips in winter and peaks in summer however as the signal is quite weak it is hard to tell. All boxplots are approximately symmetrical and some have a few outliers.

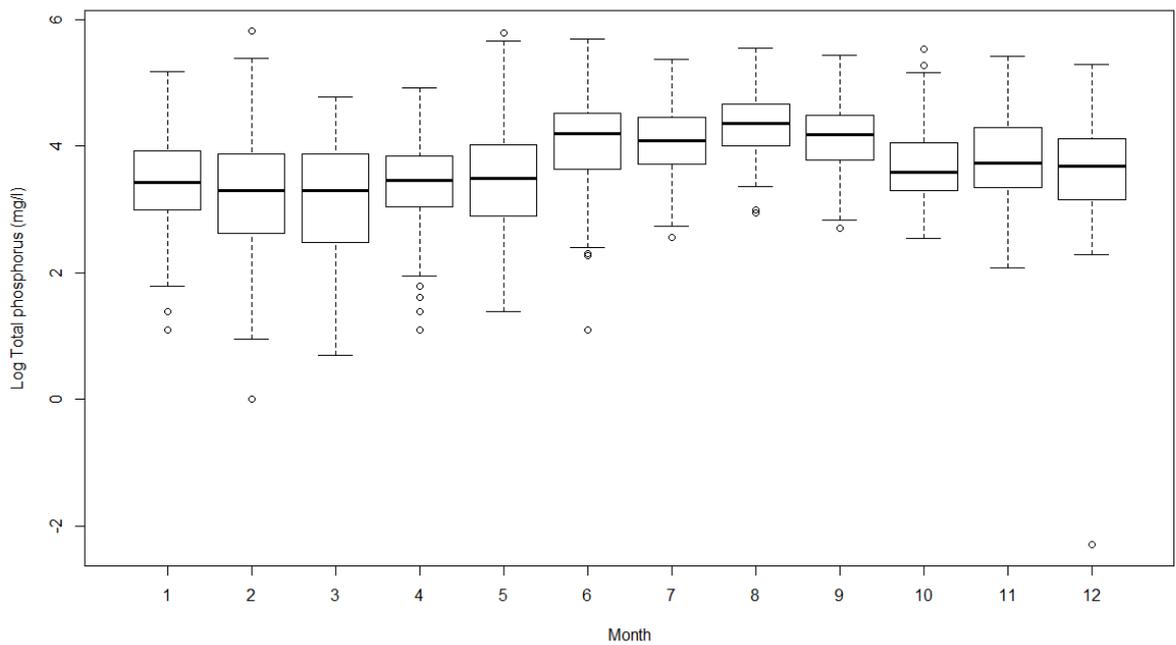


Figure 4.2: Boxplots of Log Total Phosphorus by Month

As described in chapter 1, missing values within the series will be estimated using local linear regression with weights plus variance estimated from a random normal where the variance is representative of the rest of the series.

From work carried out within chapter 3, the observed changepoints can be seen within figure 4.3 and table 4.1 where seasonal adjustment was carried out by modelling the peaks and dips of summer and winter with a harmonic function and then taking the estimates from the original series.

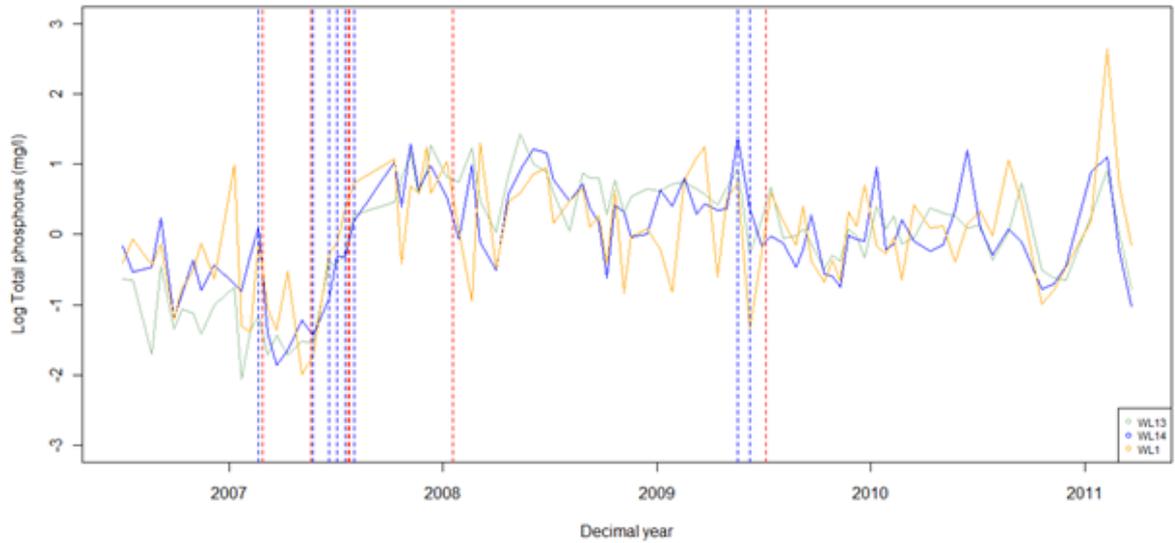


Figure 4.3: Time series of seasonally adjusted log TP (WL13, WL14 and WL1) with superimposed changepoint locations.

Table 4.1: Location of changepoints for stations WL13, WL14 and WL1

WL13	LLR			2007.564		
	B&H		2007.395			
	BSM		2007.395	2007.542		2009.378
WL14	LLR	2007.159	2007.385	2007.564	2008.048	2009.512
	B&H					
	BSM	2007.137	2007.466	2007.542		2009.436
WL1	LLR			2007.564		
	B&H					
	BSM		2007.446			

The vast majority of changes occur around mid-2007 and there is also a small grouping around mid-2009.

As turbines were installed and roads built around mid-2007 we will set an indicator function to account for a change at this time point.

Below in figure 4.4, a plot of log TP with a superimposed model for seasonal trend is shown and in figure 4.5 a plot of log TP with a superimposed model for long term trend is shown. These plots allow us to observe differences between sites by allowing both month and time to vary with site by including interaction terms.

Initially, we can look at whether the seasonal trend is consistent throughout the series. If the mean level between sites crosses at any point, this may be evidence that a change has occurred. This can also be said for the long term trend.

To do this, a simple GAM model can be fit to the data where the only parameters taken into account are Site as a parametric term and Month as a smooth term. The same can be done for long term trend by including a smooth term for Decimal.year instead of month. The overall trend for both seasonal and long term components can be found by excluding the parametric term Site.

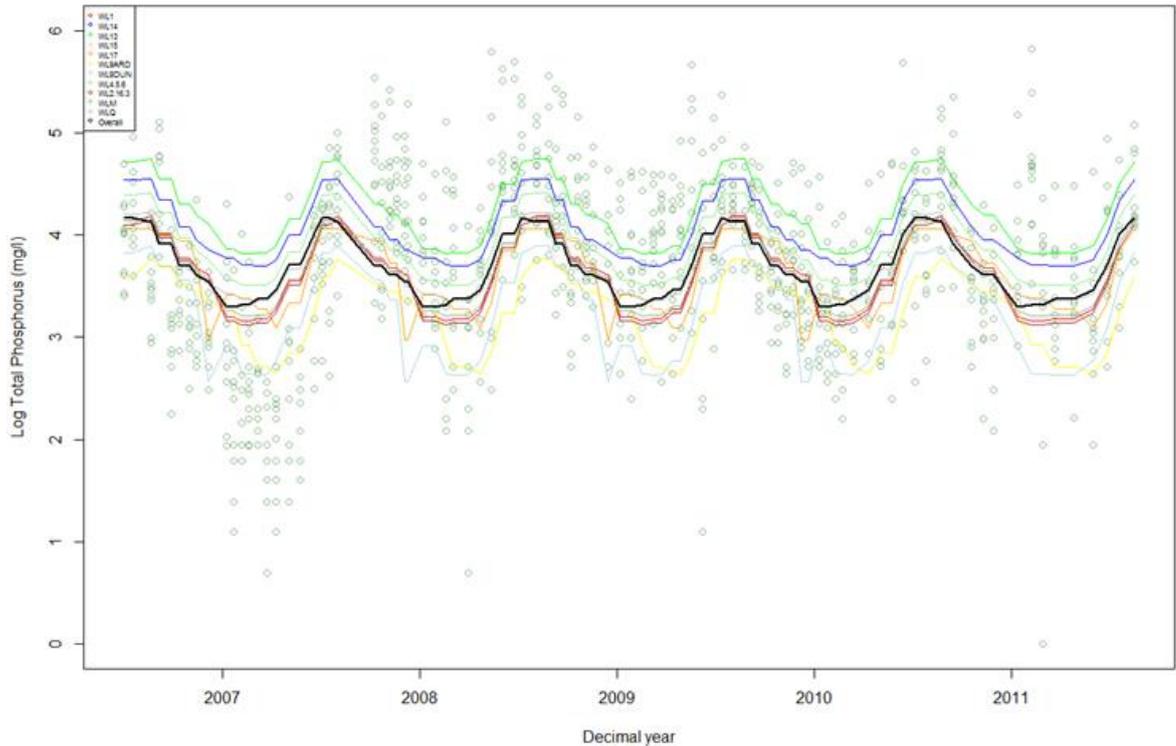


Figure 4.4: Time series of log Total Phosphorus (Superimposed seasonal component model)

Within figure 4.4 the seasonal trend for each site is shown using the model $E(\log(TP_{ijr})) = Site_{ij} + f_1(Month_{ir}) + f_3(Month_{ir}).Site_{ij}$. The interaction between the smooth of month and site allows predictions to be more fluid such that the seasonal pattern can change with each site rather than being fixed. An overall trend can also be seen and is the black prediction. Overall, sites follow the same seasonal pattern as each other and each site has a different mean level. There is little overlap within the plot, curves are relatively parallel.

4.2.2 Generalised Additive Models

Increasingly complicated models were considered within the analysis for the 11 sites each with both parametric and smoothing terms within them. ANOVA p-values are calculated between models where a p-value below 0.05 indicates that the more complex model should be used.

Taking into account previous analysis in chapters 2 and 3, there is evidence of a change point mid-2007. The informal analysis allowed us to observe that the gradient of the smooth curve changes between the left and right at this point and that long term seasonal trend changes slightly between sites.

A good starting point will be to include overall terms for Site and an Indicator function. We will allow a smooth of Decimal year to vary with Indicator to account for the change in gradient, a smooth of Decimal year to vary with Site to allow overall trends between sites to vary and smooths of month to vary with both the Indicator and Site to allow the seasonal trend to change at the changepoint and between sites. This model may have to be reduced if certain terms are not significant.

Table 4.2: GAM models (TP)

Model Description	AIC	GCV	ANOVA (p-value)
$E(\log(TP_{ijk})) = Indicator_{ik} + Site_{ij} \\ + f_1(Decimal.Year_i).Indicator_{ik} \\ + f_2(Month_{ir}).Indicator_{ik} \\ + f_3(Decimal.Year_i, Month_{ir}) \\ + f_4(Decimal.Year_i).Site_{ij} + f_5(Month_{ir}).Site_{ij}$	1841.286	0.3575	---
$E(\log(TP_{ijk})) = Site_{ij} + f_1(Decimal.Year_i) + f_2(Month_{ir}) \\ + f_3(Decimal.Year_i, Month_{ir}) \\ + f_4(Decimal.Year_i).Site_{ij} + f_5(Month_{ir}).Site_{ij}$	1730.866	0.3259	1.797e-05

Surprisingly, our model with the Indicator terms was a worse fit than a model without these terms. This final model was then adjusted by changing the number of knots within each of our smoothing parameters to optimise the model. The second model has a lower AIC and GCV score. Furthermore, the ANOVA shows a p-value of below 0.05 when comparing models, indicating the second model is a better fit.

4.2.3 Varying Coefficient Model

We can adapt and extend our original GAM model to account for a variation in model parameters before and after the change point. This can be achieved by allowing the effects of each of our covariates to vary with the variable *Indicator* and will allow variables to change before and after this point.

The term *Indicator* has been added within the model as a main effect since it is a factor, if it was numeric it would not be added since the resulting smooth is not usually subject to a centering constraint. However if we wished our model to vary with a numeric term a term within the model can be specified as $s(\text{Numeric. var}).\text{Numeric. var}$ which is commonly used for smoothing or nonparametric regression on Y versus X as stated by Hastie and Tibshirani (1993).

A simple model will be considered where decimal year and month as smooth terms vary by the indicator function, allowing the overall trend and seasonal trend to vary at the changepoint.

Table 4.3: Varying coefficient model (TP)

Model Description	AIC	GCV
$E(\log(TP_{ijk_r})) = Site_{ij}$ $: Indicator_{ik} + s(Decimal.year_i).Indicator_{ik}$ $+ s(Month_{ir}).Indicator_{ik}$	1934.284	0.3887

The varying coefficient model shown above has a GCV of 0.3887 and AIC of 1934.284. Both of these values are higher than the final GAM model shown in table 4.5, although this is not surprising as this model does not include some of the terms that the GAM model did. Therefore, this model will not be considered any further.

4.2.4 Final model

Our final model has been chosen as the one with the lowest AIC and GCV score which is our generalised additive model. This will be used to track total phosphorus over time at the 11 different sites at Whitelee.

Smoothing functions will also be presented to show the overall trend, seasonal trend and the interaction between these two variables.

Normality of residuals will also be assessed for the 11 different sites as will the assumption of correlated errors. Note that the EDF is the estimated degrees of freedom for our smooth terms.

Table 4.4: ANOVA table for final model

GAM Object			
Model: $E(\log(TP_{ijk_r})) = Site_{ij} + f_1(Decimal.Year_i) + f_2(Month_{ir}) + f_3(Decimal.Year_i, Month_{ir}) + f_4(Decimal.Year_i).Site_{ij} + f_5(Month_{ir}).Site_{ij}$			
	Parameters	Estimate	p
Parametric Coefficients	α	3.637	< 2e-16 ***
	Site ₂ (WL13)	0.579	1.23e-13 ***
	Site ₃ (WL14)	0.398	2.89e-07 ***
	Site ₄ (WL15)	0.022	0.77495
	Site ₅ (WL17)	0.027	0.72077
	Site ₆ (WL12.16.3)	-0.036	0.63707
	Site ₇ (WL4.5.6)	0.251	0.00117 **
	Site ₈ (WL9ARD)	-0.374	1.38e-06 ***
	Site ₉ (WL9DUN)	-0.402	2.27e-07 ***
	Site ₁₀ (WLM)	0.124	0.10765
	Site ₁₁ (WLQ)	0.129	0.09518 .
			EDF
Smooth Terms	s(Decimal.year, k = 50)	48.620	< 2e-16 ***
	s(Month, k = 12)	8.446	4.33e-06 ***
	s(Decimal.year, Month)	27.000	8.94e-12 ***
	s(Decimal.year): Site ₁ (WL1)	0.916	0.0002 ***
	s(Decimal.year): Site ₂ (WL13)	4.015	2.17e-13 ***
	s(Decimal.year): Site ₃ (WL14)	2.474	5.47e-06 ***
	s(Decimal.year): Site ₄ (WL15)	0.916	0.000182 ***
	s(Decimal.year): Site ₅ (WL17)	5.168	2.60e-05 ***
	s(Decimal.year): Site ₆ (WL2.16.3)	2.276	0.000145 ***
	s(Decimal.year): Site ₇ (WL4.5.6)	2.835	0.000380 ***
	s(Decimal.year): Site ₈ (WL9ARD)	0.916	0.000366 ***
	s(Decimal.year): Site ₉ (WL9DUN)	4.491	0.000386 ***
	s(Decimal.year): Site ₁₀ (WLM)	3.019	0.000756 ***
	s(Decimal.year): Site ₁₁ (WLQ)	0.916	0.000245 ***
	s(Month): Site ₁ (WL1)	1.153	0.658971
	s(Month): Site ₂ (WL13)	0.956	0.671026
	s(Month): Site ₃ (WL14)	6.866	0.198697
	s(Month): Site ₄ (WL15)	1.623	0.419930
	s(Month): Site ₅ (WL17)	7.326	0.057303 .
	s(Month): Site ₆ (WL2.16.3)	0.956	0.665366
	s(Month): Site ₇ (WL4.5.6)	0.956	0.670680
	s(Month): Site ₈ (WL9ARD)	3.272	0.004020 **
	s(Month): Site ₉ (WL9DUN)	7.802	0.009846 **
	s(Month): Site ₁₀ (WLM)	0.956	0.664752
s(Month): Site ₁₁ (WLQ)	3.631	0.120702	

Of the 12 parametric coefficients including the intercept, 6 were significant with p-values less than 0.05. The mean level for sites WL15, WL17 WL2.16.3, WLM and WLQ did not significantly differ from site WL1 which is included within the intercept.

Of the 25 smooth terms, 16 were significant with p-values less than 0.05. Decimal year, with a specified number of knots equal to 50 was highly significant as was the smooth function for month with a specified number of knots equal to 12. The interaction between decimal year and month was significant which indicates that the seasonal trend differs between years. The interaction between decimal year and site was also significant for all levels of the factor site, which indicates that the overall trend differs between sites.

The interaction between month and site is only significant for two sites and a further one is borderline (WL17). This means that the seasonal trend differs between sites only for significant levels of the factor site whereas the rest of the do not have different seasonal patterns between sites.

As this smooth function does have significant terms, we can leave this function within the model.

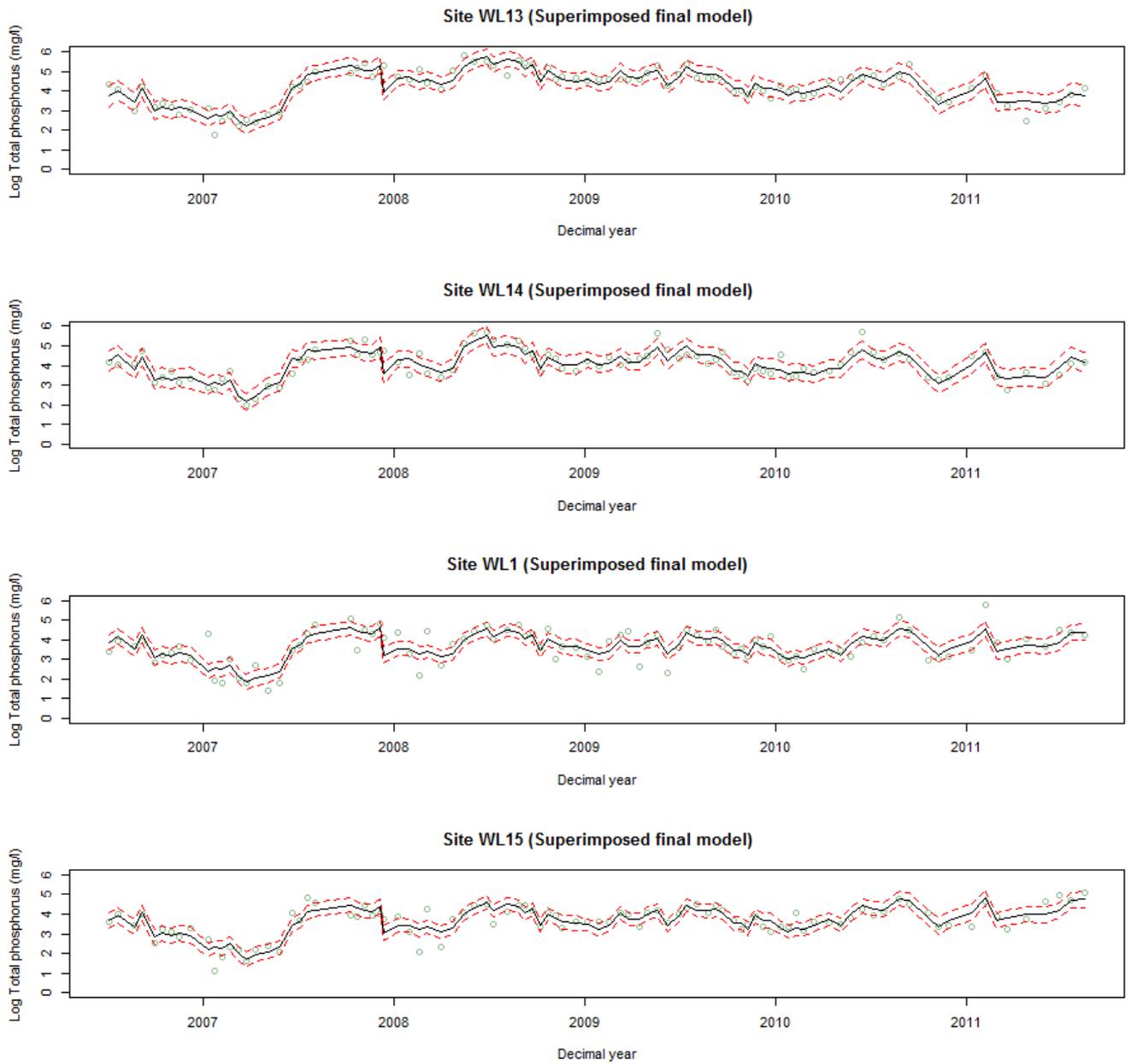


Figure 4.6: Prediction estimates and intervals (TP)

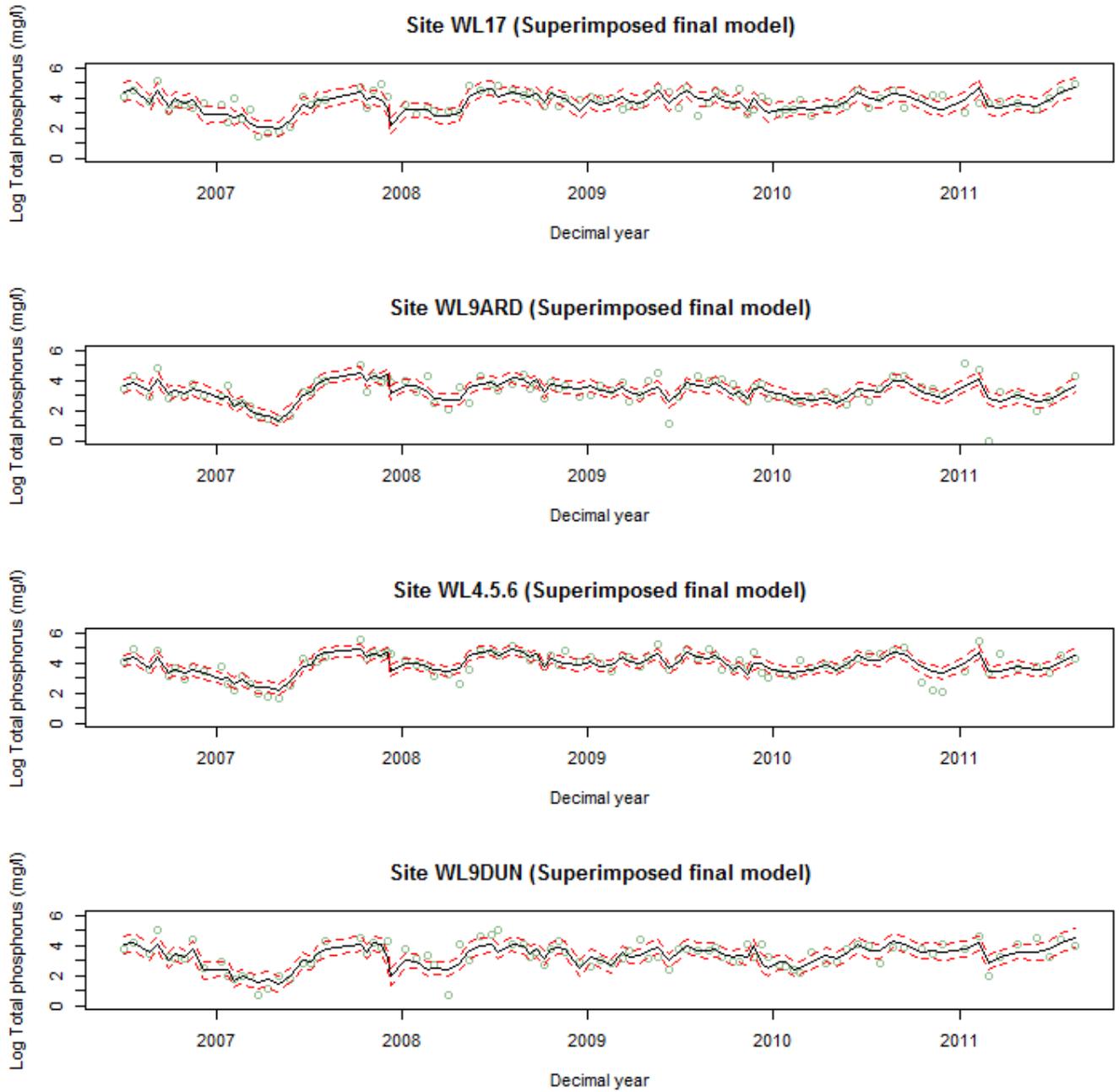


Figure 4.7: Prediction estimates and intervals (TP)

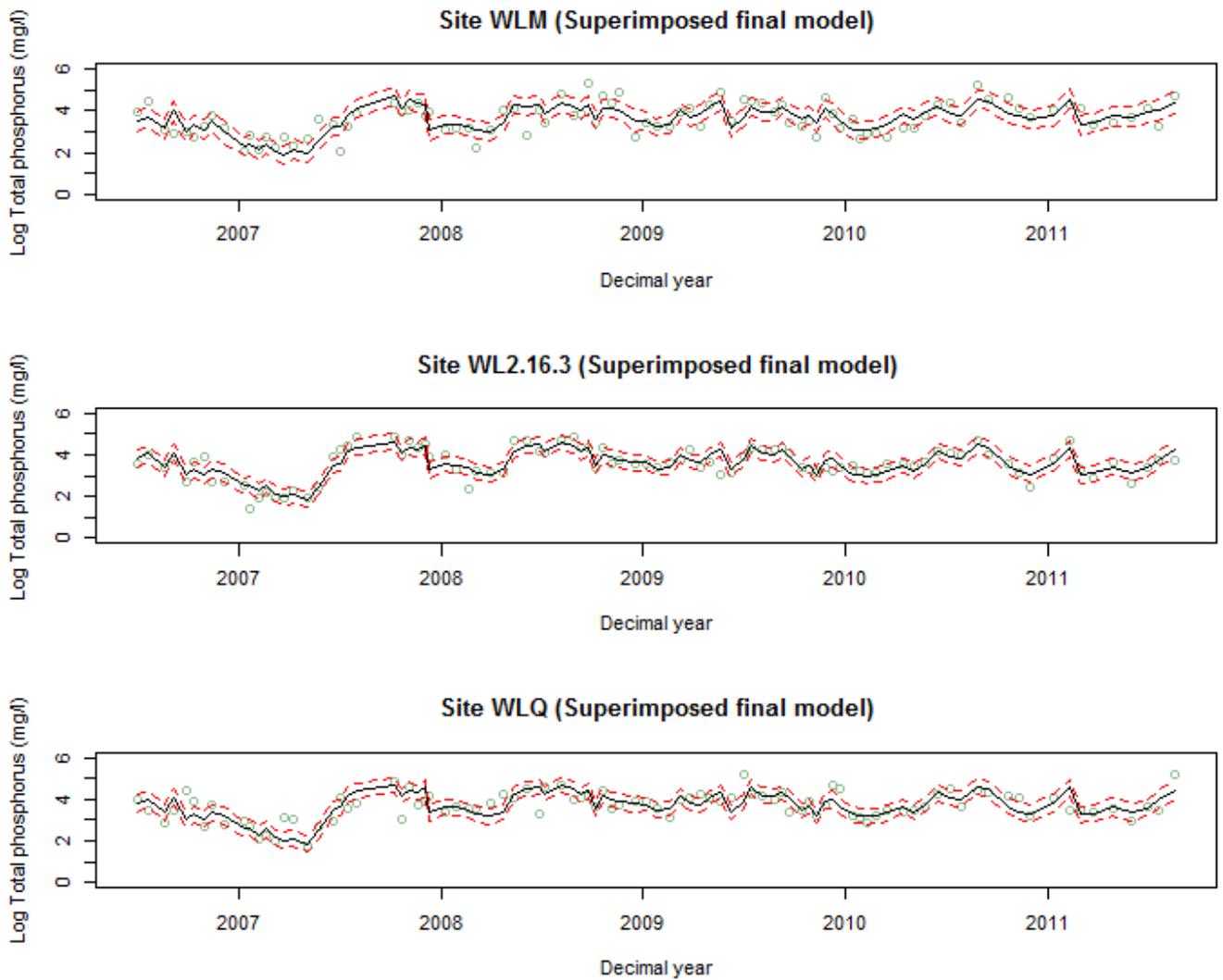


Figure 4.8: Prediction estimates and intervals (TP)

Intervals were calculated by multiplying each point by $1.96 \times SE$ where SE is the standard error calculated as $E = s/\sqrt{n}$ where n is the number of observations and s is the sample standard deviation. It is also true that normality is an assumption of a Generalised Additive Model although this is relaxed in comparison to a Linear Model. This assumption can be informally checked with the use of histograms, and the assumption of uncorrelated errors can be evaluated with the use of an ACF plot.

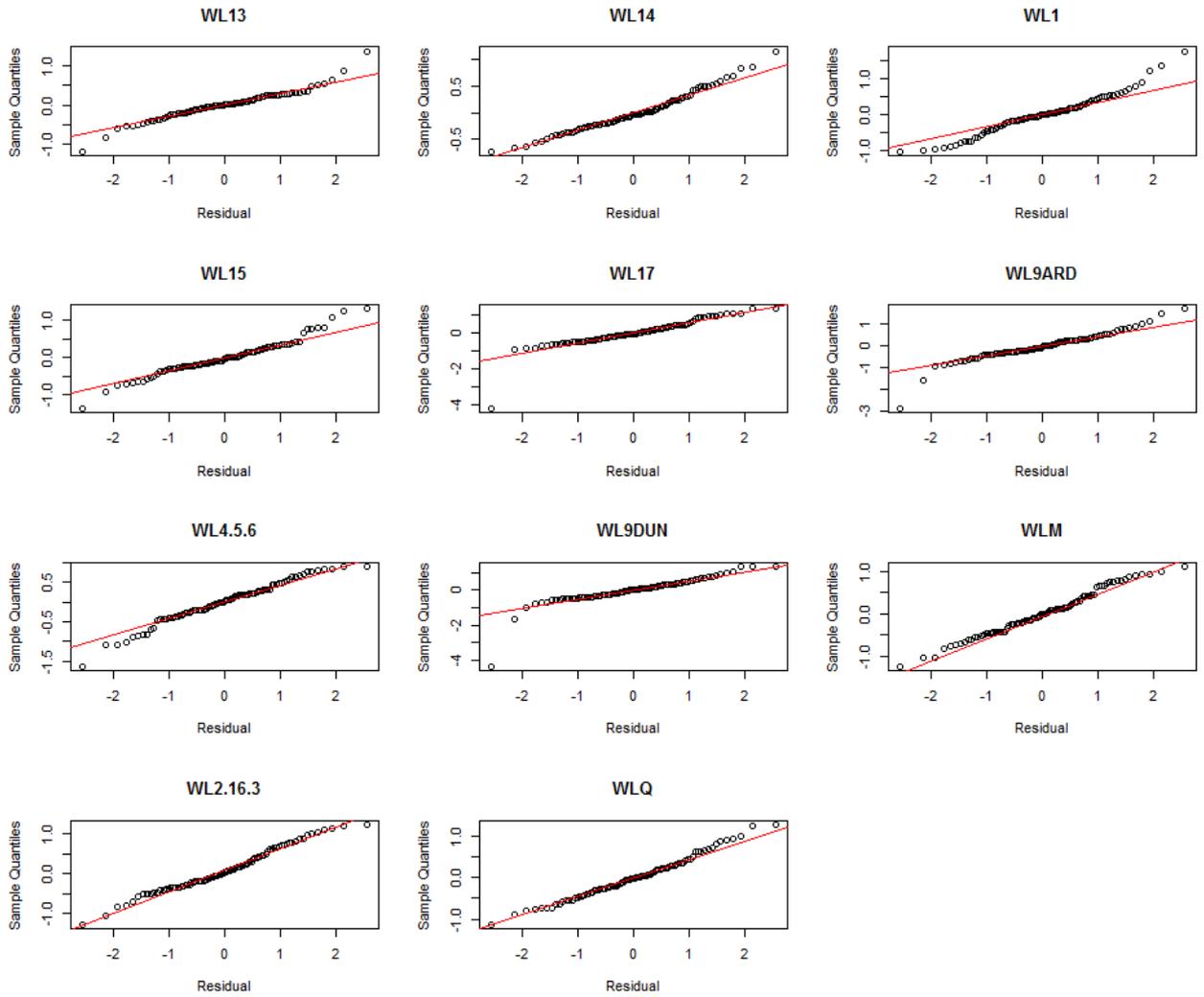
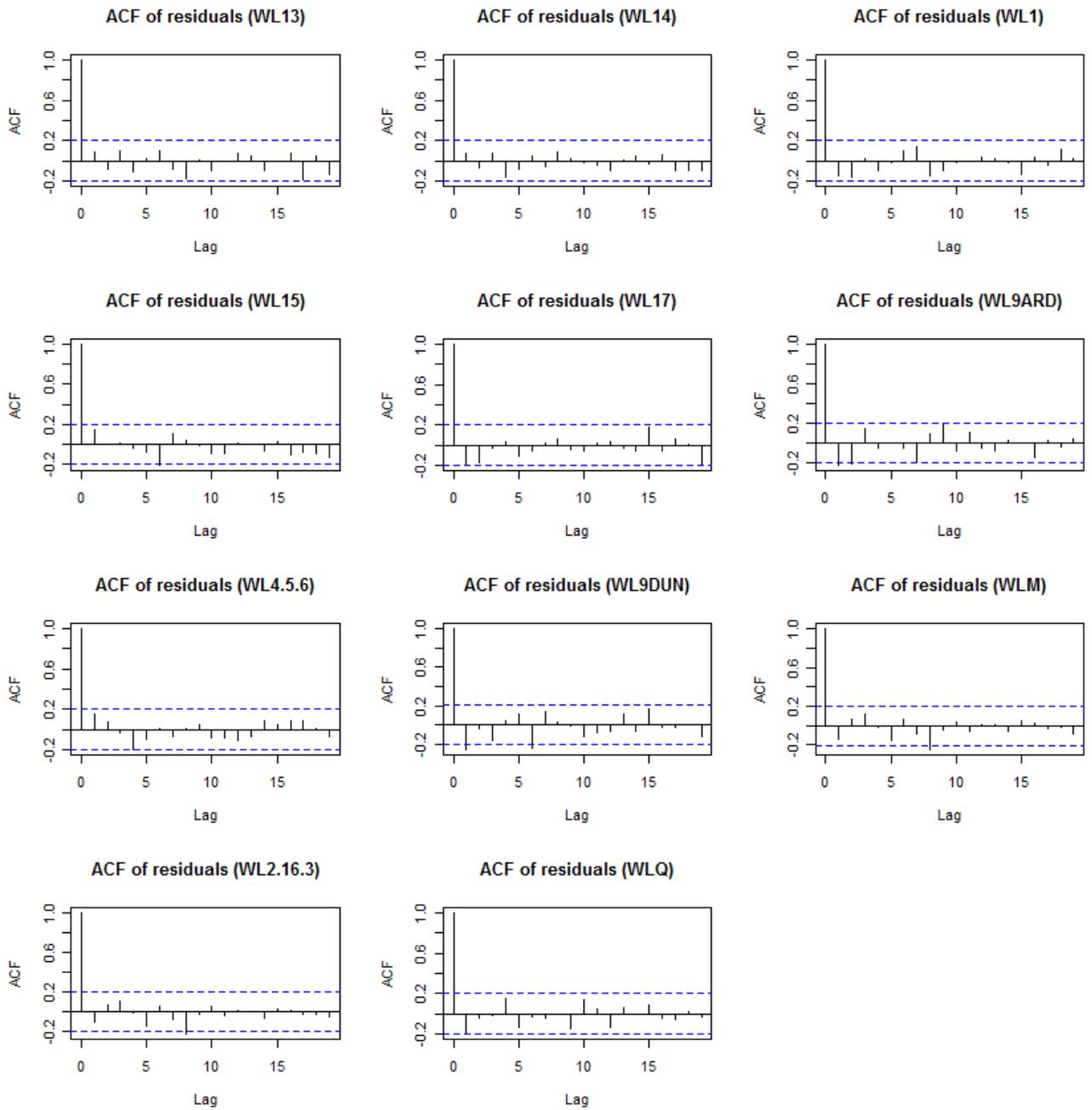


Figure 4.9: QQ normal plots of residuals (TP)

Within figure 4.9 QQ normal plots can be seen for each site. Within each plot, normality can be assumed since almost all points lie on or near the fitted line with no tailing off of points towards the ends. Within WL9DUN there is an outlier however this can be expected.



Figures 4.10: ACF plots of residual values at each site for log Total Phosphorus

The ACF plots shown in figure 4.10 show the autocorrelation between lags within the sites at Whitelee. Within the 11 plots, there are no significant autocorrelation values and we can therefore conclude that the residuals within each site are not correlated with each other.

4.2.5 Conclusion (Total Phosphorus)

From our subjective impression we could determine that log Total Phosphorus had a distinct seasonal trend and a slight negative overall trend. Within each month, data was normally distributed with a few outliers. From previous analysis, a changepoint located in mid-2007 was found, this was incorporated into our modelling structure.

The models which were considered consisted of the terms Site, Indicator, Decimal year and Month. Smooth terms and interactions between smooth variables and non-smoothed variables were also considered. Varying coefficient models were also considered however in the end the GAM model shown below was chosen as our final model:

$$E(\log(TP_{ijk_r})) = Site_{ij} + f_1(Decimal.Year_i) + f_2(Month_{ir}) + f_3(Decimal.Year_i, Month_{ir}) + f_4(Decimal.Year_i).Site_{ij} + f_5(Month_{ir}).Site_{ij}$$

Residuals for this model were normally distributed and ACF plots did not show any significant autocorrelation at lags >1. As these assumptions held, the final model could then be used to model our data.

4.3.1 Total Organic Carbon: Modelling

Figure 4.11 shows the overall series over time where each line represents a different site. It is not clear from the plots whether there is a change mid-2007. There does not seem to be a shift in mean or change in trend at this point.

Seasonal variation is clear from the two plots, with peaks in summer and troughs in winter.

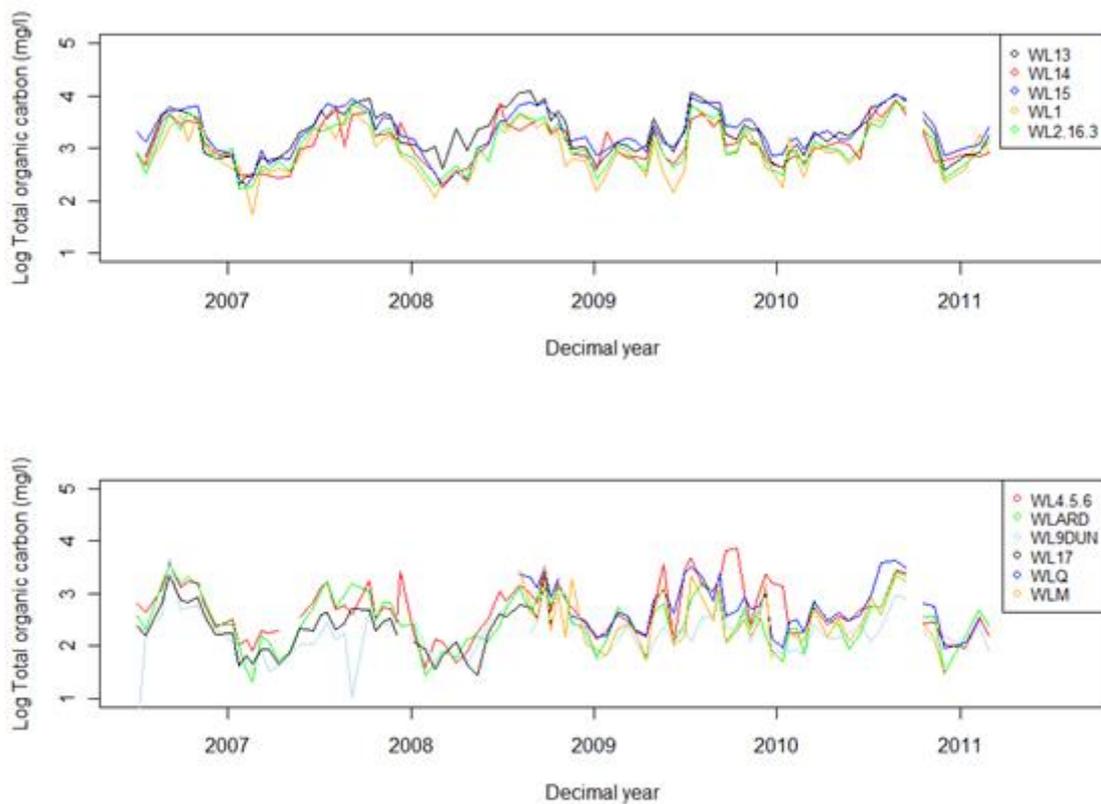


Figure 4.11: Log Total Organic Carbon vs. Decimal Year

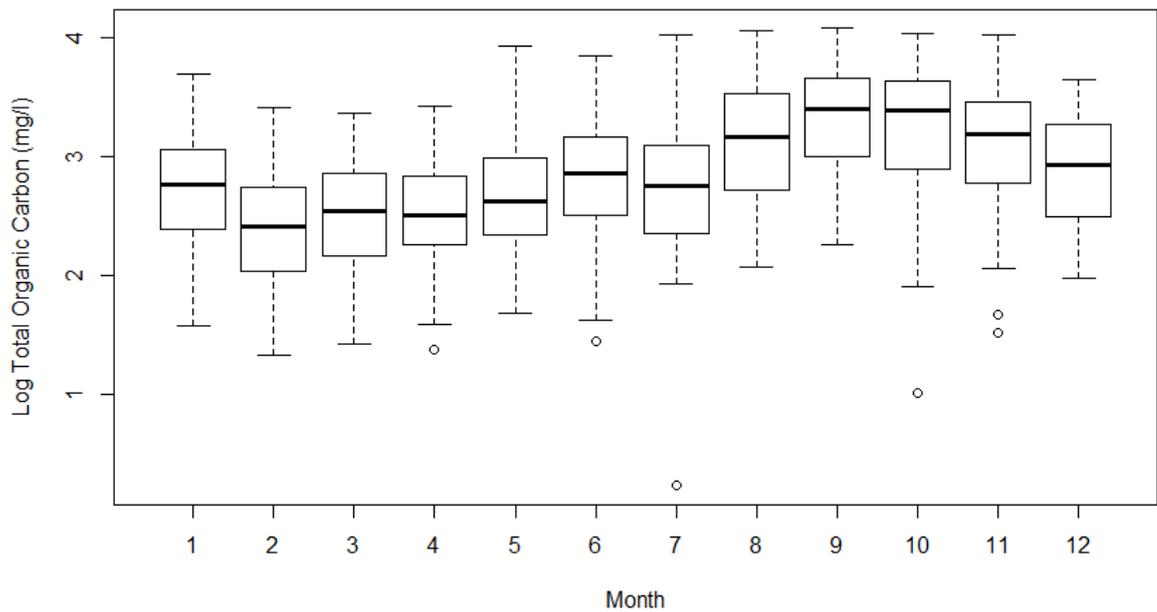


Figure 4.12: Boxplots of Log Total Phosphorus by Month

The boxplot above shows log TOC by each month. The 12 box-and-whisker diagrams within the plot represent each month from January to December. The spread is relatively equal for each plot with a few outliers in July, October and November however the rest of the data is roughly normally distributed.

Clearly there is a seasonal trend within the series with low TOC levels around winter and high TOC levels in late summer.

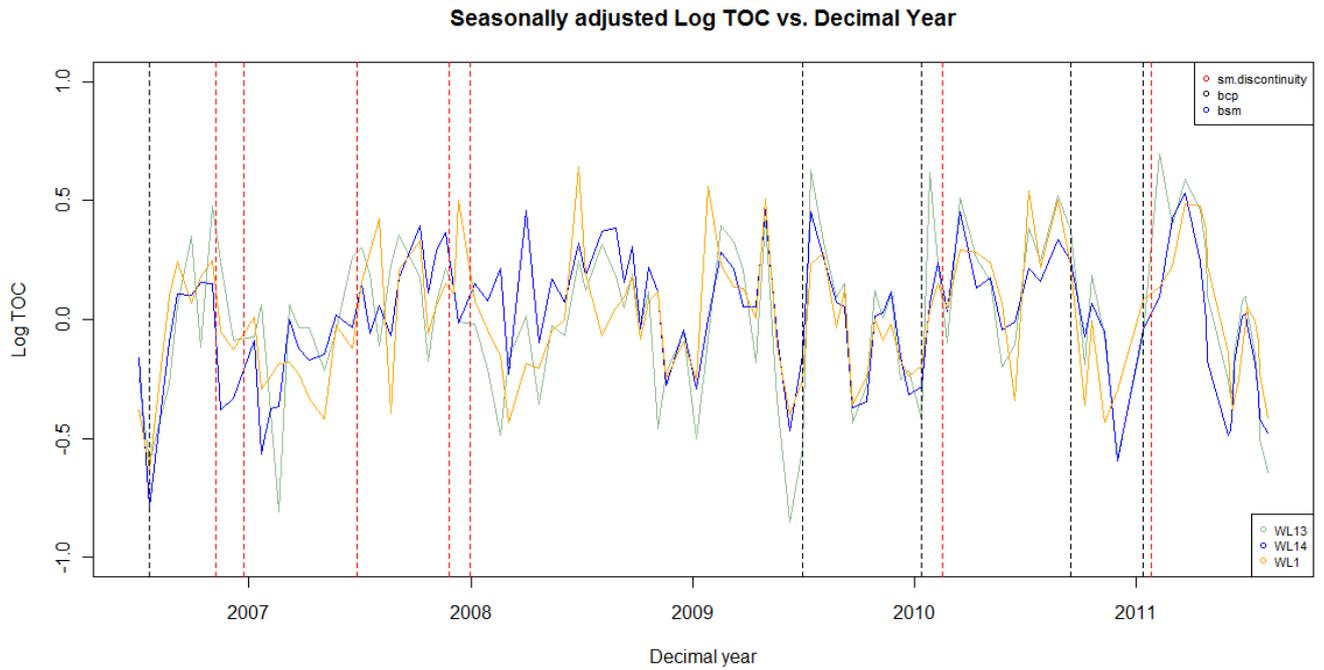


Figure 4.13: Seasonally adjusted log TOC vs. decimal year for stations WL13, WL14 and WL1 with superimposed changepoint locations.

Table 4.5: Location of changepoints for stations WL13, WL14 and WL1

WL13	LLR	2006.852	2007.430	2010.126	2010.126		
	B&H						
	BSM						
WL14	LLR		2007.486	2007.997	2010.115		
	B&H	2006.551				2010.704	
	BSM						
WL1	LLR	2006.977		2007.904			2011.067
	B&H			2009.493	2010.030		2011.030
	BSM						

Changepoint locations are fairly equally spread along the series within figure 4.13 and there does not seem to be any clustering of locations. Of the twelve changepoints, 6 lie to the left

of the midpoint and 6 lie to the right. From this evidence there does not seem to be a change present.

Initially, we can look at whether the seasonal trend is consistent throughout the series. If the mean level between sites crosses at any point, this may be evidence that a change has occurred. This can also be said for the long term trend.

To do this, a simple GAM model can be fitted to the data where the only parameters taken into account are Site as a parametric term and Month as a smooth term. The same can be done for long term trend by including a smooth term for Decimal.year instead of month. The overall trend for both seasonal and long term components can be found by excluding the parametric term Site.

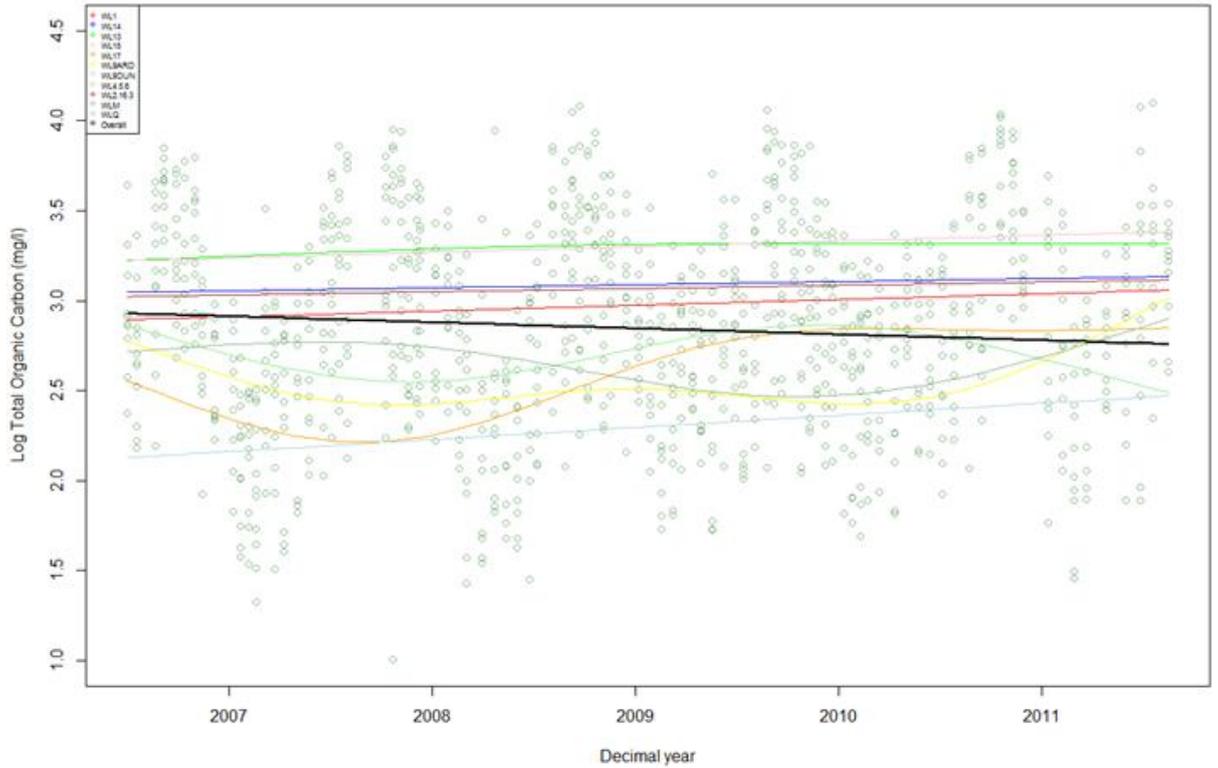


Figure 4.14: Log Total Organic Carbon versus Decimal year (Superimposed long term component model)

Within figure 4.14, we can see the long term trend for each site at Whitelee including the overall trend. Sites WL13, WL15, WL14 and WL1 are all relatively parallel whereas the remaining sites do cross and overlap at certain points throughout the series.

Log Total Organic Carbon (Seasonal trend at each site)

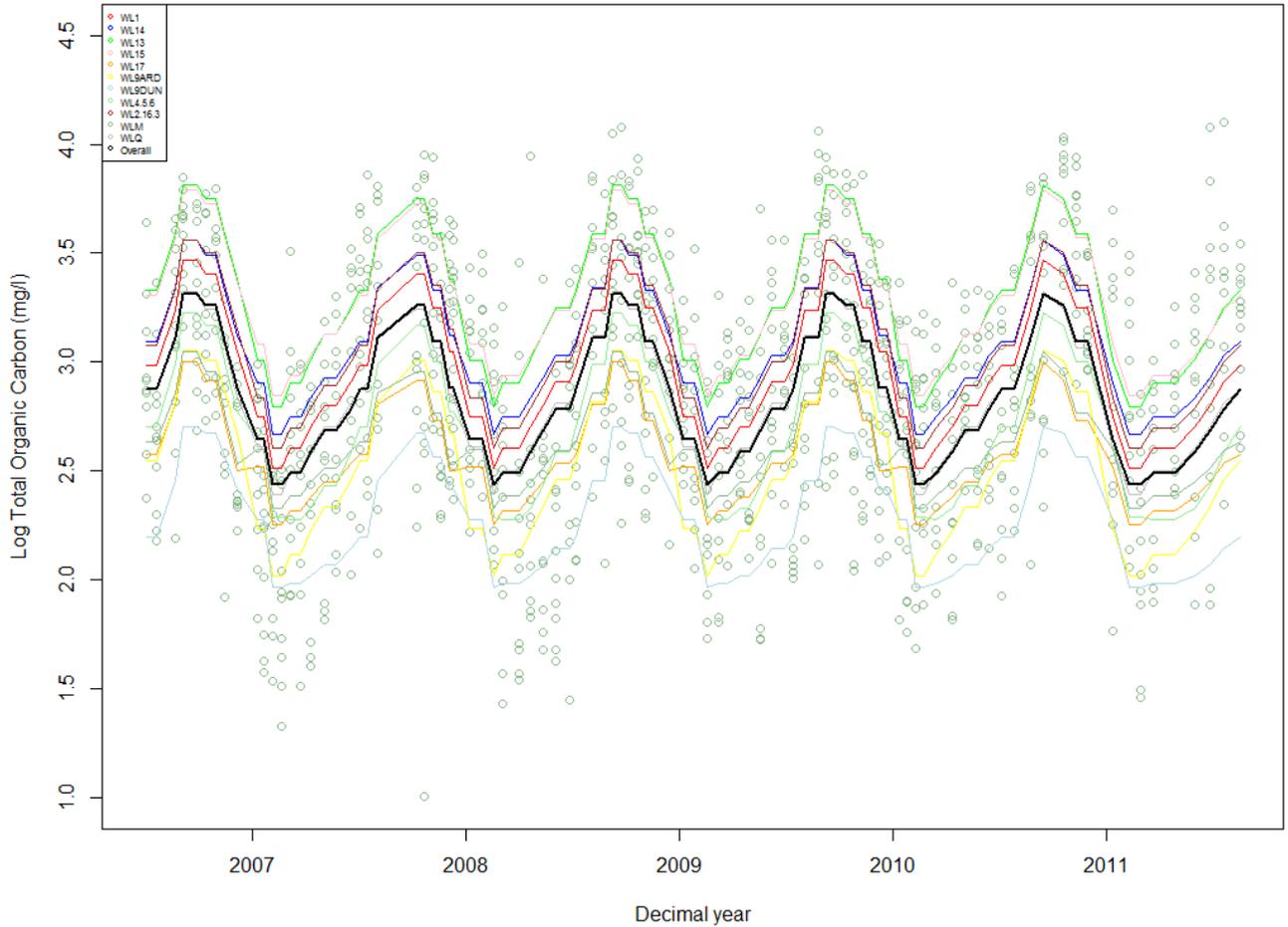


Figure 4.15: Log Total Organic Carbon versus Decimal year (Superimposed seasonal component model)

The seasonal components for each site are plotted above within figure 4.15 where our model is specified as $E(\log(TOC_{ijr})) = Site_{ij} + f_1(Month_{ir}) + f_3(Month_{ir})$. $Site_{ij}$, allowing us to see which seasonal patterns differ. Clearly, the overall seasonal pattern is strong and for individual sites lines generally sit parallel.

4.3.2 Generalised Additive Models

Several models were considered within the analysis for the 11 sites each with both parametric and smoothing terms within them.

From the BACI modelling earlier, there was no evidence to suggest that a changepoint is present within the dataset and from informal analysis no changes can be observed around mid-2007, therefore it does not make sense to include an indicator function to identify this changepoint.

ANOVA p-values are calculated between the current model and the previous model where a p-value below 0.05 indicates that the more complex model should be used.

Table 4.6: GAM models (TOC)

Model Description	AIC	GCV	ANOVA (p-value)
$E(\log(TOC_{ijr})) = Site_{ij} + f_1(Decimal.Year_i) + f_2(Month_{ir}) + f_3(Decimal.Year_i, Month_{ir}) + f_4(Decimal.Year_i).Site_{ij}$	846.170	0.1339	---
$E(\log(TOC_{ijr})) = Site_{ij} + f_1(Decimal.Year_i) + f_2(Month_{ir}) + f_3(Decimal.Year_i, Month_{ir}) + f_4(Decimal.Year_i).Site_{ij} + f_5(Month_{ir}).Site_{ij}$	738.416	0.1214	2.2e-16

It is clear that each additional parameter must be included within the model since both the AIC and the GCV reduces in value as each parameter is added. No indicator term has been added since there was no evidence to suggest a change after mid-2007. We can change the number of knots within our smooth functions to obtain a final GAM model.

Table 4.7: Final GAM model (TOC)

Model Description	AIC	GCV
$E(\log(TOC_{ijr})) = Site_j + f_1(Decimal.Year_i) + f_2(Month_{ir}) + f_3(Decimal.Year_i, Month_{ir}) + f_4(Decimal.Year_i).Site_{ij} + f_5(Month_{ir}).Site_{ij}$	600.5789	0.108

4.3.3 Varying Coefficient Model

We can adapt and extend our original GAM model to account for a variation in model parameters before and after the change point. Since there is not enough evidence to support a change point around mid-2007, specifying an indicator function will not add any value to our statistical model. However, by allowing our parameters to vary with the term month we allow the seasonal pattern to change over time. The term *Month* has a numeric variable and a parameter on its own will not be added since the resulting smooth is not usually subject to a centering constraint. However we can specify this term as $s(\text{Month})$. *Month* which is commonly used for smoothing or nonparametric regression on Y versus X as stated by Hastie and Tibshirani (1993).

Table 4.8: Varying coefficient model (TOC)

Model Description	AIC	GCV	ANOVA (p-value)
$E(\log(\text{TOC}_{ijr})) = \text{Site}_{ij} \cdot \text{Month}_{ir} + s(\text{Month}_{ir}) \cdot \text{Month}_{ir} + s(\text{Decimal. year}_i) \cdot \text{Month}_{ir}$	1048.59	0.1629	---
$E(\log(\text{TOC}_{ijr})) = \text{Site}_{ij} \cdot \text{Month}_{ir} + s(\text{Month}_{ir}) \cdot \text{Month}_{ir} + s(\text{Decimal. year}_i) \cdot \text{Month}_{ir}$	974.47	0.1517	7.439e-15 ***

This model has a GCV of 0.1517 and AIC of 974.47. Both of these values are higher than the final GAM model shown in table 4.8.

4.3.4 Final model

Our final model has been chosen as the one with the lowest AIC and GCV score which is our final generalised additive model. This will be used to track total organic carbon over time at the 11 different sites at Whitelee.

Smoothing functions will also be presented to show the overall trend, seasonal trend and the interaction between these two variables.

Normality of residuals will also be assessed for the 11 different sites as will the assumption of correlated errors.

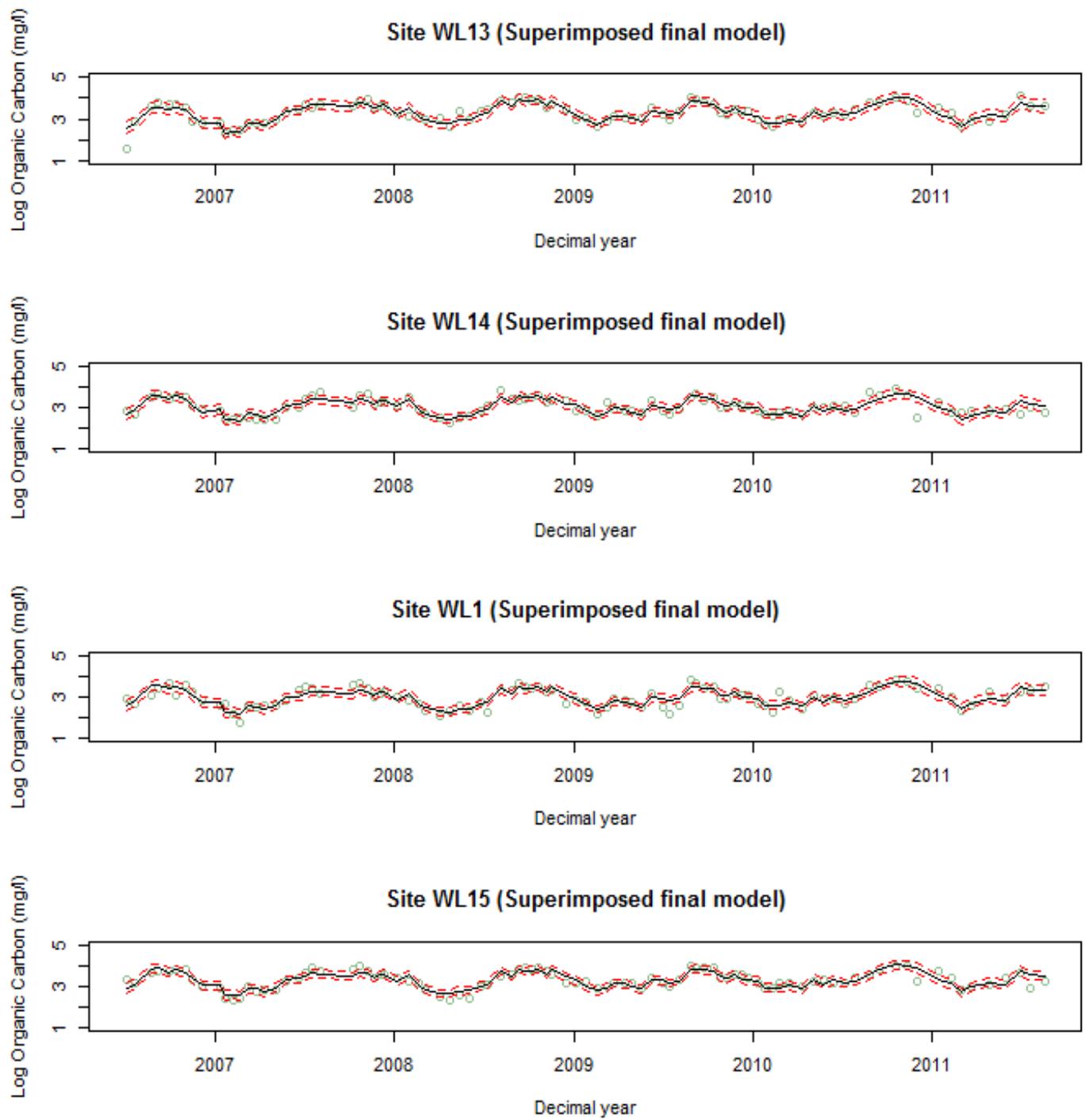
Table 4.9: ANOVA table for final model (TOC)

<i>Model: $E(\log(TOC_{ijr})) = Site_{ij} + f_1(Decimal.Year_i) + f_2(Month_{ir}) + f_3(Decimal.Year_i, Month_{ir}) + f_4(Decimal.Year_i).Site_{ij} + f_5(Month_{ir}).Site_{ij}$</i>			
	Parameters	Estimate	p
Parametric Coefficients	α	2.960	< 2e-16 ***
	Site ₂ (WL13)	0.313	2.75e-12 ***
	Site ₃ (WL14)	0.107	0.01558 *
	Site ₄ (WL15)	0.315	2.32e-12 ***
	Site ₅ (WL17)	-0.386	< 2e-16 ***
	Site ₆ (WL12.16.3)	0.121	0.00625 **
	Site ₇ (WL4.5.6)	-0.243	5.11e-08 ***
	Site ₈ (WL9ARD)	-0.423	< 2e-16 ***
	Site ₉ (WL9DUN)	-0.66	< 2e-16 ***
	Site ₁₀ (WLM)	-0.32	9.49e-13 ***
	Site ₁₁ (WLQ)	-0.18	5.02e-05 ***
		EDF	
Smooth Terms	$s(Decimal.year, k = 50)$	39.677	1.18e-15 ***
	$s(Month, k = 12)$	8.4676	0.003737 **
	$s(Decimal.year): Site_1(WL1)$	0.9164	0.133485
	$s(Decimal.year): Site_2(WL13)$	2.6884	0.013000 *
	$s(Decimal.year): Site_3(WL14)$	1.5558	0.195831
	$s(Decimal.year): Site_4(WL15)$	2.6491	0.099090
	$s(Decimal.year): Site_5(WL17)$	8.3385	3.00e-05 ***
	$s(Decimal.year): Site_6(WL2.16.3)$	3.0423	0.2358
	$s(Decimal.year): Site_7(WL4.5.6)$	6.3690	0.0020**
	$s(Decimal.year): Site_8(WL9ARD)$	7.5349	0.0003***
	$s(Decimal.year): Site_9(WL9DUN)$	8.2496	0.0012**
	$s(Decimal.year): Site_{10}(WLM)$	7.6226	0.0016**
	$s(Decimal.year): Site_{11}(WLQ)$	0.9164	0.0992
	$s(Decimal.year, Month)$	25.7129	7.74e-05***
	$s(Month): Site_1(WL1)$	0.9112	0.8279
	$s(Month): Site_2(WL13)$	1.7594	0.3128
	$s(Month): Site_3(WL14)$	0.9112	0.2602
	$s(Month): Site_4(WL15)$	0.9113	0.5817
	$s(Month): Site_5(WL17)$	3.2564	0.1440
	$s(Month): Site_6(WL2.16.3)$	0.9112	0.9582
	$s(Month): Site_7(WL4.5.6)$	0.9113	0.4040
	$s(Month): Site_8(WL9ARD)$	0.9112	0.2358
	$s(Month): Site_9(WL9DUN)$	1.9677	0.1031
	$s(Month): Site_{10}(WLM)$	5.9860	0.0966
$s(Month): Site_{11}(WLQ)$	6.6616	0.2554	

Even though none of the terms for the interaction between Month and Site are significant, when our model with this term is compared with a model without this term within an ANOVA, the output suggests using the more complicated model which can be seen within table 4.13.

Analysis of Deviance Table						
Model 1: $E(\log(TOC_{ijr}))$ $= Site_{ij} + f_1(Decimal.Year_i) + f_2(Month_{ir}) + f_3(Decimal.Year_i, Month_{ir})$ $+ f_4(Decimal.Year_i).Site_{ij} + f_5(Month_{ir}).Site_{ij}$						
Model 2: $E(\log(TOC_{ijr}))$ $= Site_{ij} + f_1(Decimal.Year_i) + f_2(Month_{ir}) + f_3(Decimal.Year_i, Month_{ir})$ $+ f_4(Decimal.Year_i).Site_{ij}$						
	Resid. Df	Resid. Dev	Df	Deviance	F	Pr(> F)
1	863.16	78.669				
2	890.58	84.680	-27.418	-6.0107	2.4054	8.076e-05 ***

Table 4.10: Comparing models with and without $f(Month_{ir}).Site_{ij}$ (TOC)



Figures 4.16: Prediction estimates and intervals (TOC)

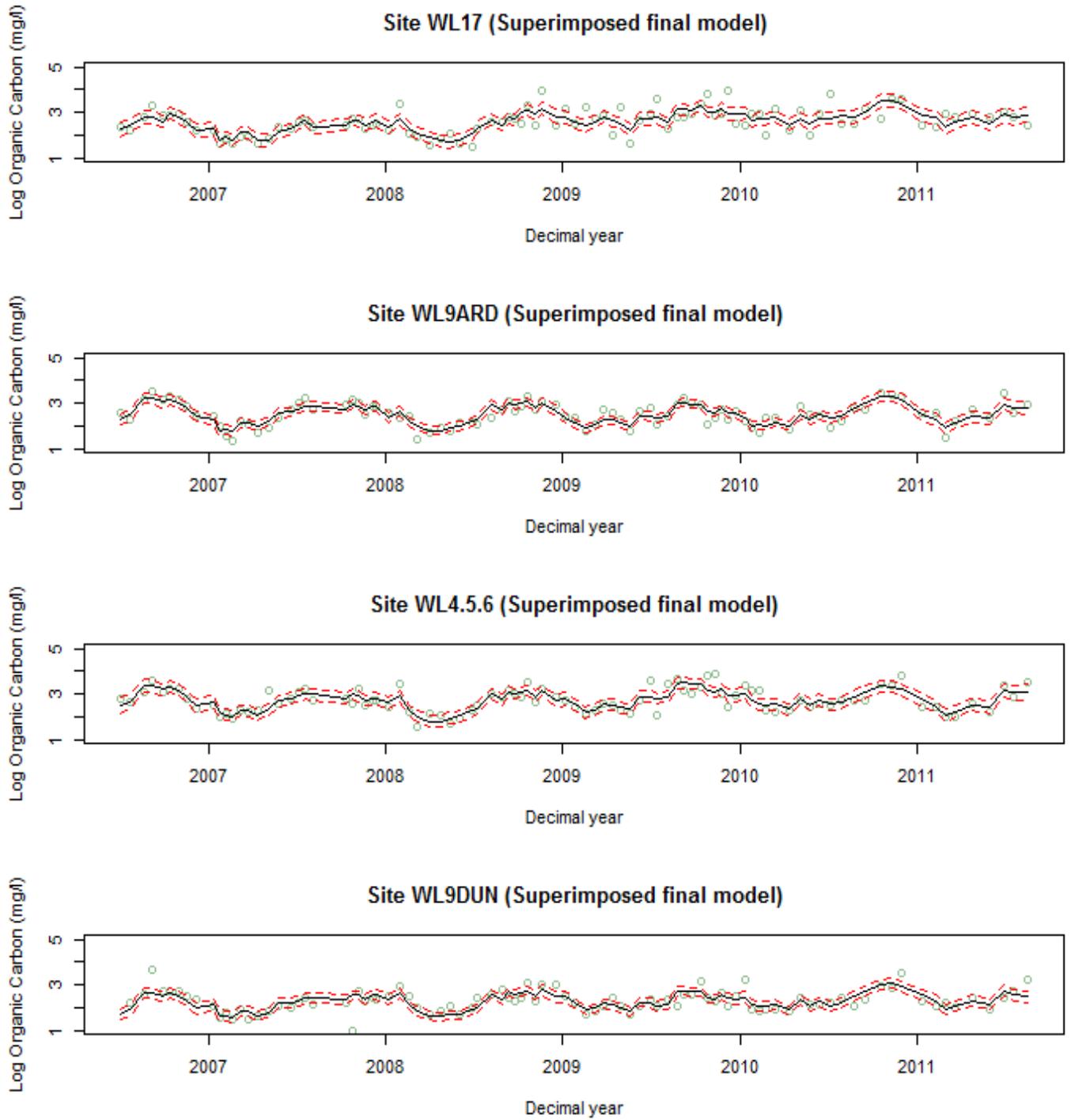


Figure 4.17: Prediction estimates and intervals



Figure 4.18: Prediction estimates and intervals (TOC)

Our final model has been used to predict the 11 TOC series at Whitelee. The model is a good fit to the data with separate mean levels for each series, an overall trend, an overall seasonal pattern and three varying coefficients.

The plotted smoothing functions for decimal year, month and their interaction can be seen within figures 4.15-4.18.

Autocorrelation is also assessed within figure 4.20.

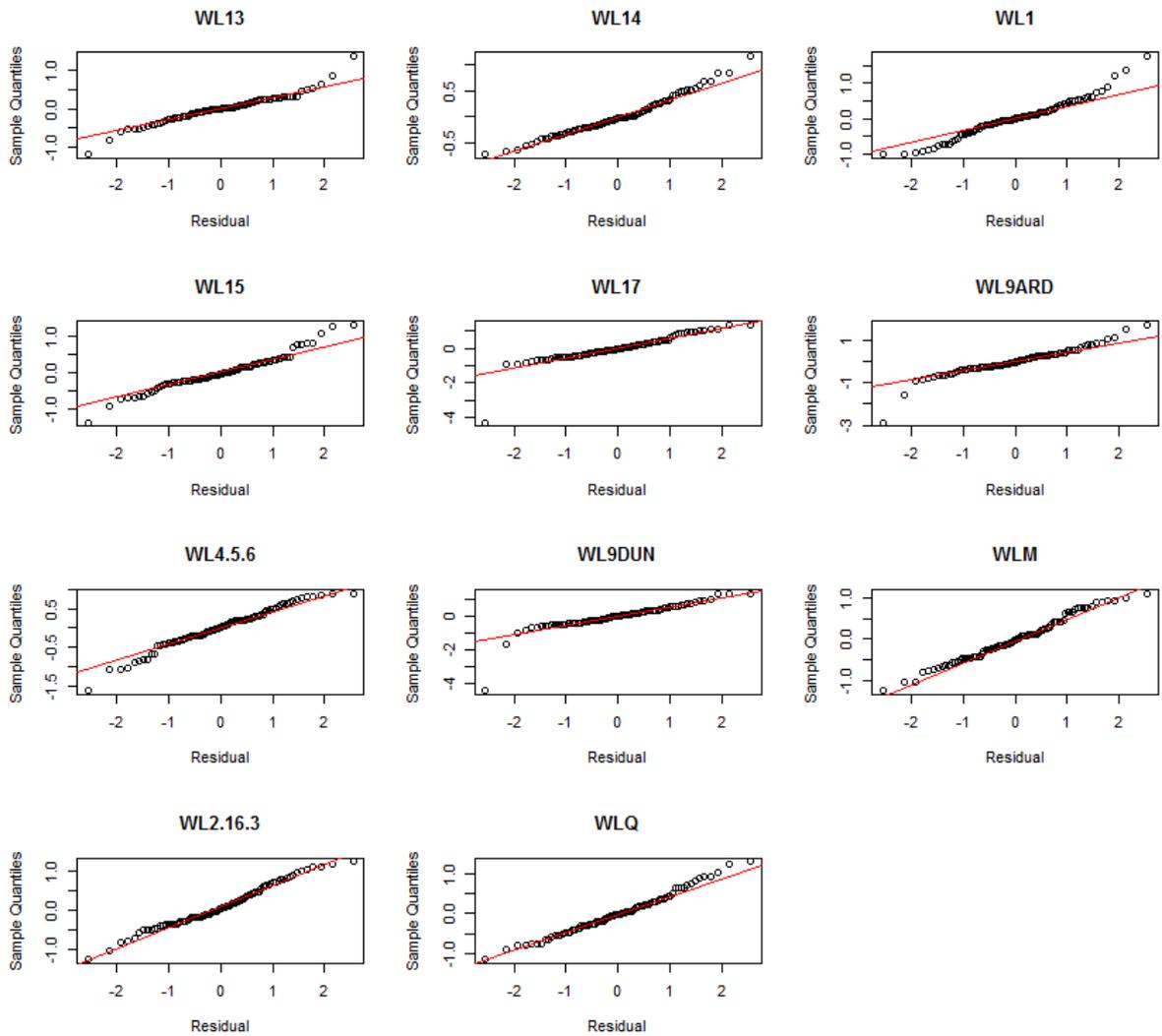
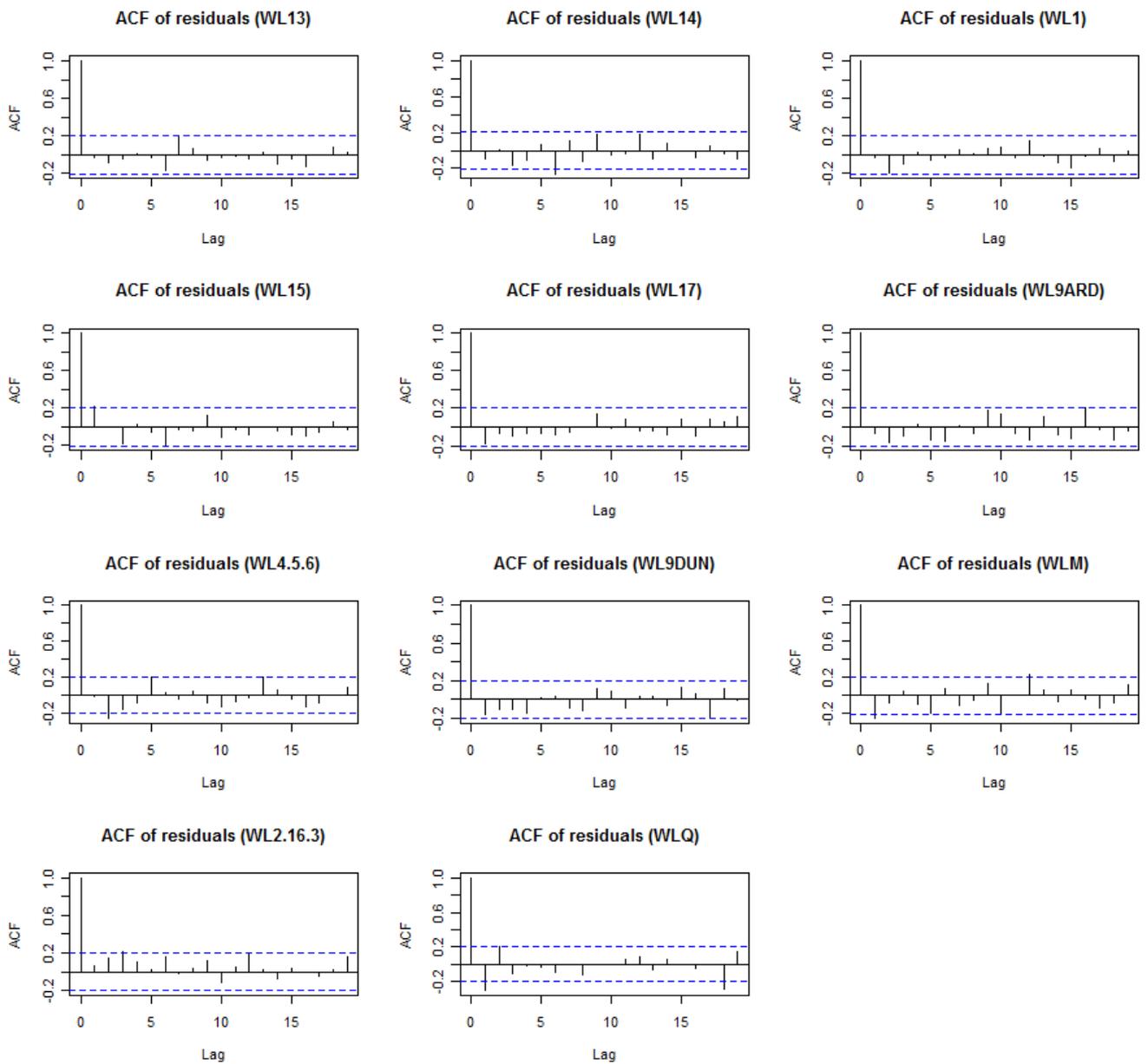


Figure 4.19: QQ normal plots at each site for log Total Organic Carbon

Within figure 4.19 histograms of the residual values between the fitted model and the data can be seen. Residuals for all sites are approximately normally distributed and we can therefore assume that our models are a good fit to the data.



Figures 4.20: ACF plots of residual values at each site for log Total Organic Carbon

The ACF plots shown above show the autocorrelation between residual values at the 11 sites at Whitelee. Within almost all plots the assumption of no autocorrelation between lags can be assumed however at site WLQ and WLM there are significant correlations at lag 1. The auto correlation at these two sites is borderline and can be ignored.

4.4.5 Conclusion (Total Organic Carbon)

Of the three variables in which analysis was carried out, log TOC had the strongest seasonal signal however the long term trend was roughly horizontal for all sites.

From chapters 2 and 3, we determined that no changepoint was present and therefore no Indicator function was present within any analysis.

As with log TP, GAM models and varying coefficient models were considered to model log TOC. However, the final model used after checking assumptions for normal residuals and non-correlated errors is shown below:

$$E(\log(TOC_{ijr})) = Site_{ij} + f_1(Decimal.Year_i) + f_2(Month_{ir}) + f_3(Decimal.Year_i, Month_{ir}) + f_4(Decimal.Year_i).Site_{ij} + f_5(Month_{ir}).Site_{ij}$$

This was then used to model each site individually.

4.4.1 Nitrate Oxide: Modelling

Figure 4.21 shows the overall series over time where each line represents a different site. These series have many missing values however these will be filled in using local linear regression with weights. From the plots alone it is not clear whether there is a regime change around mid-2007.

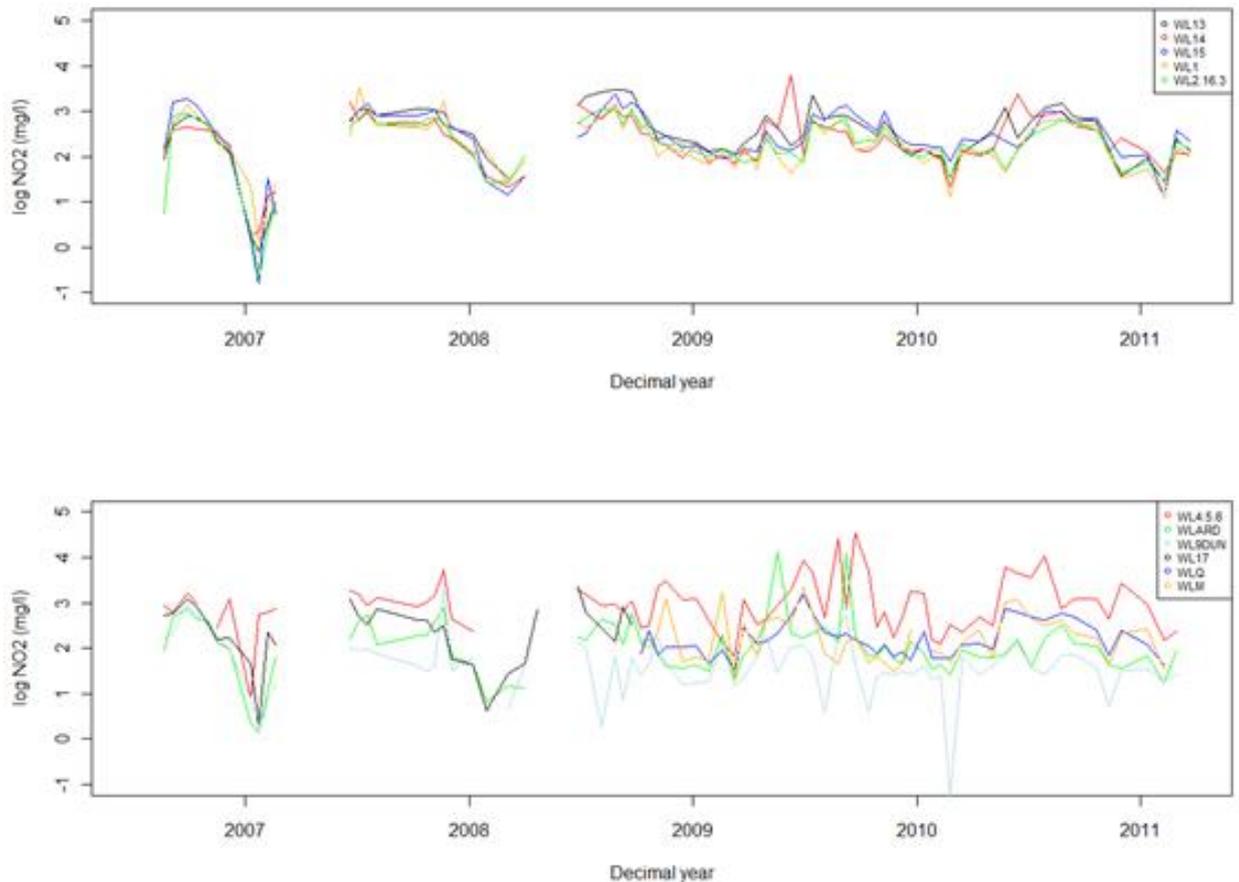


Figure 4.21: Log Nitrate Oxide vs. Decimal Year

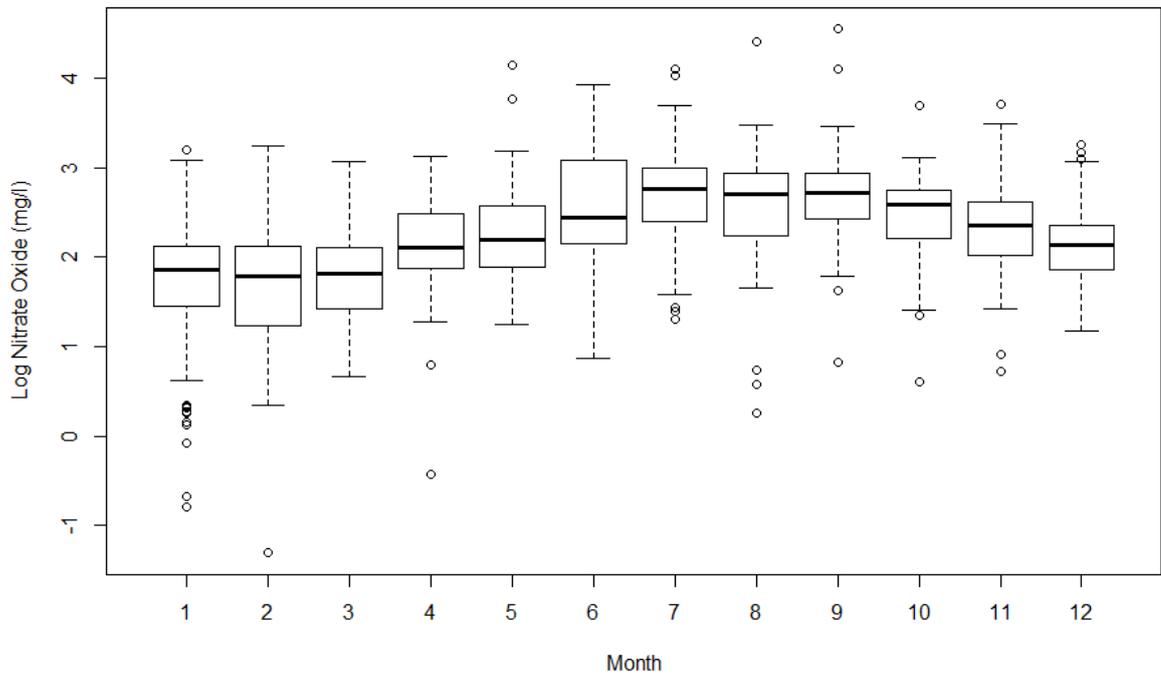


Figure 4.22: Log Nitrate Oxide vs. Decimal Month

Figure 4.22 is a boxplot showing 12 months of the year for log NO₂. Most months are approximately symmetrical with a few residuals. There is a clear seasonal pattern with a peak in summer and a trough in winter.

From this plot alone, we can see it is likely that the term Month will be a significant predictor of log NO₂.

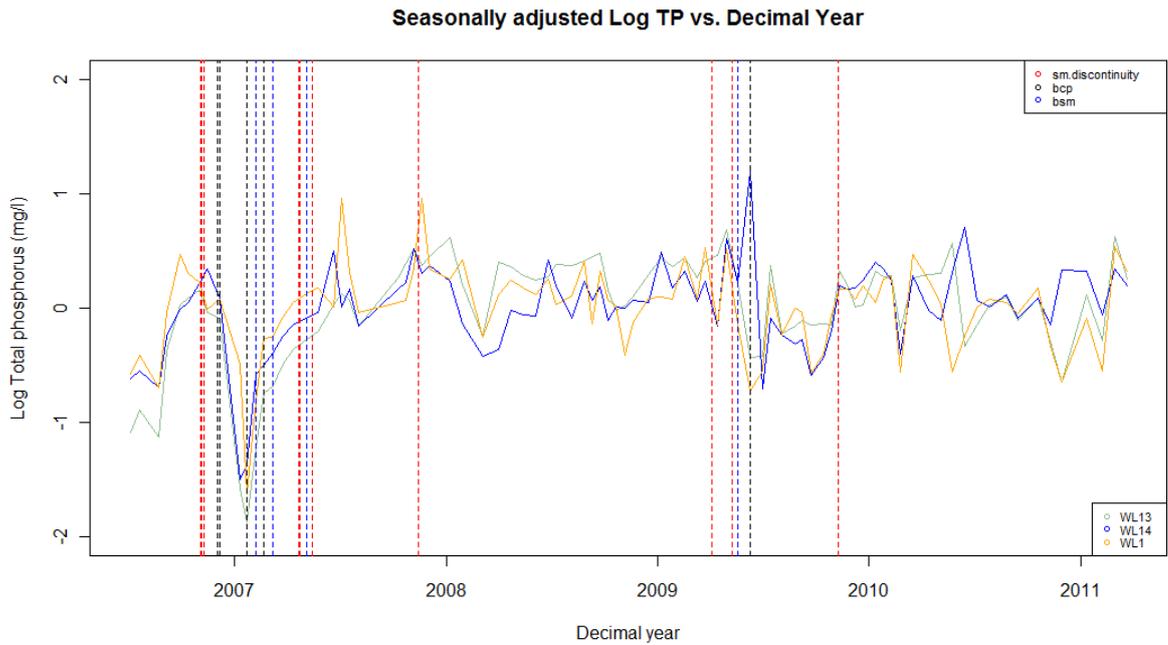


Figure 4.23: Seasonally adjusted log NO2 vs. decimal year for stations WL13, WL14 and WL1 with superimposed changepoint locations.

Table 4.14 and figure 4.23 show the changepoints located by the three tests. Most change points are located around 2007, however there are quite a few which lie around mid-2009. Of the 19 changes which were located, 13 lie around 2007 and 4 lie around mid-2009. An indicator function will be used within the analysis to account for these changes in 2007 and 2009 however as shown previously the GAM models may fit well enough without the need to these functions.

Table 4.11: Location of changepoints for stations WL13, WL14 and WL1

WL13	LLR	2006.854	2007.307		2009.352	2009.838	
	B&H	2006.929	2007.137				
	BSM		2007.34		2009.378		
WL14	LLR	2006.852	2007.367	2007.866	2009.255		
	B&H	2006.929	2007.058		2009.436		
	BSM		2007.181				
WL1	LLR	2006.852	2007.307				
	B&H						
	BSM		2007.099				

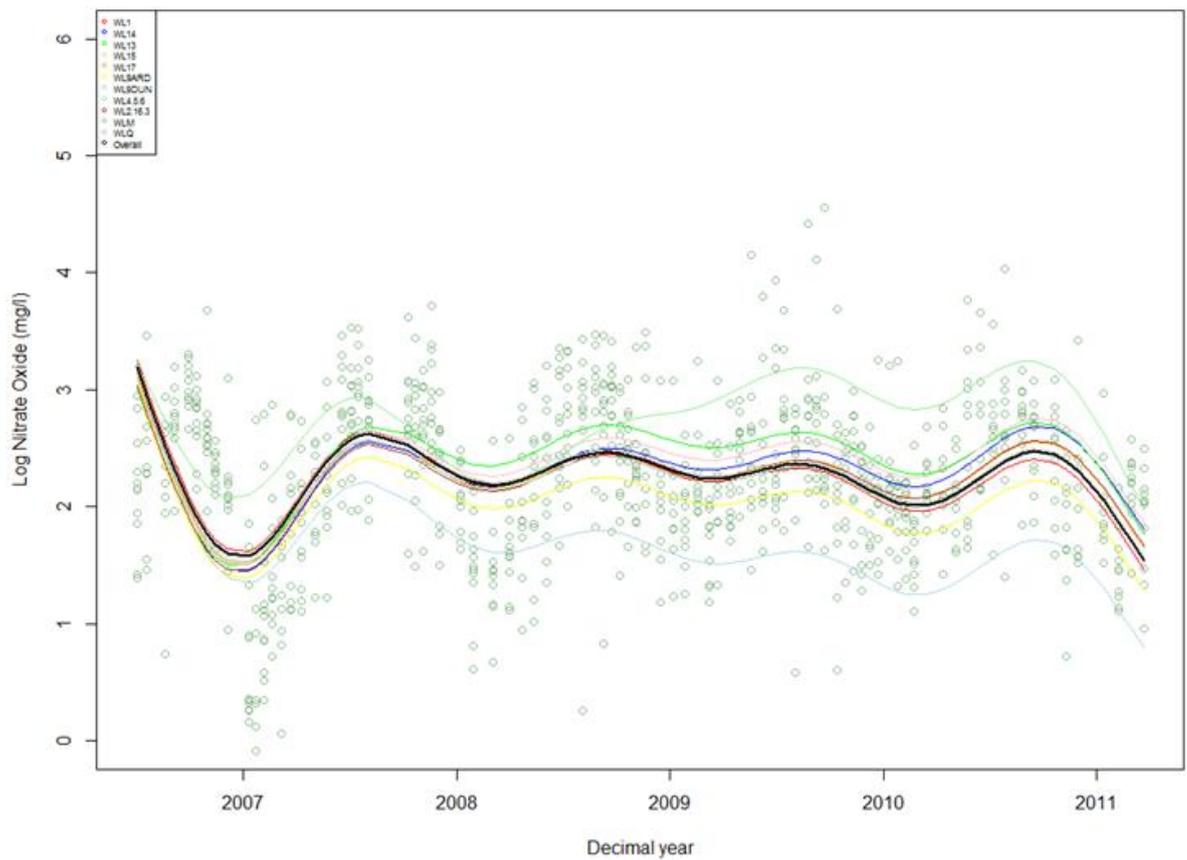


Figure 4.24: Log Nitrate Oxide versus Decimal year (Superimposed long term trend component model)

Within figure 4.24, we can see the long term trend for each site at Whitelee including the overall trend. The model used to describe the data above is described as:

$E(\log(NO2_{ij})) = Site_{ij} + f_1(Decimal.Year_i) + f_3(Decimal.Year_i)$. $Site_{ij}$ such that this model allows each site to have a separate overall mean and their own overall trend.

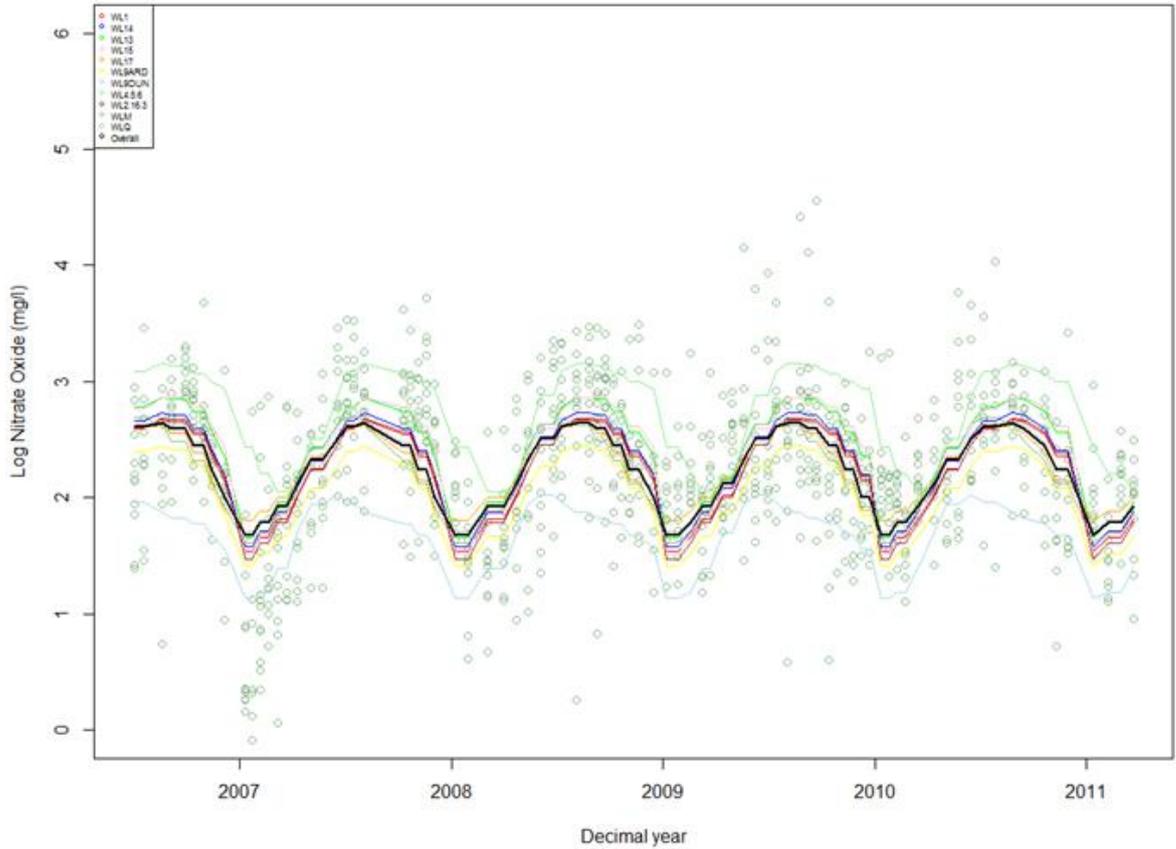


Figure 4.25: Log Nitrate Oxide versus Decimal year (Superimposed seasonal component model)

The seasonal components for each site are plotted above within figure 4.25 where our model allows us to determine which seasonal patterns differ. The model $E(\log(NO2_{ijr})) = Site_j + f_1(Month_{ir}) + f_3(Month_{ir})$. $Site_{ij}$ allows for separate means and separate seasonal patterns via sites.

4.4.2 General Additive Models

Increasingly complicated models were considered within the analysis for the 11 sites each with both parametric and smoothing terms within them. ANOVA p-values are calculated between the current model and the previous model where a p-value below 0.05 indicates that the more complex model should be used. An indicator function has been used to separate the two change points that were detected. One has been set at 2007.1 and one at 2009.3. All 8 models are shown below in mathematical notation:

Table 4.12: GAM models (NO2)

Model Description	AIC	GCV	ANOVA (p-value)
$E(\log(NO2_{ijr})) = Site_{ij} + f_1(Decimal.Year_i) + f_2(Month_{ir})$ $+ f_3(Decimal.Year_i, Month_{ir})$ $+ f_4(Decimal.Year_i).Site_{ij} + f_5(Month_{ir}).Site_{ij}$	761.834	0.171	---
$E(\log(NO2_{ijr})) = Indicator_{ik} + Site_{ij}: Indicator_{ik}$ $+ f_1(Decimal.Year_i).Indicator_{ik}$ $+ f_2(Month_{ir}).Indicator_{ik}$ $+ f_3(Decimal.Year_i, Month_{ir})$ $+ f_4(Decimal.Year_i).Site_{ij} + f_5(Month_{ir}).Site_{ij}$	784.4135	0.176	< 2.2e-16

It is clear that each additional parameter must be included within the model since both the AIC and the GCV reduces in value as each parameter is added. This final model does not include the term *indicator* as the AIC and GCV values are lower than the previous models.

Table 4.13: Final GAM model (NO2)

Model Description	AIC	GCV
$E(\log(NO2_{ijr})) = Site_j + f_1(Decimal.Year_i) + f_2(Month_{ir})$ $+ f_3(Decimal.Year_i, Month_{ir}) + f_4(Decimal.Year_i).Site_{ij}$ $+ f_5(Month_{ir}).Site_{ij}$	616.110	0.146

(Note all models assume $k = 1,2,3$ (accounting for changes in 2007 and 2009))

4.4.3 Varying Coefficient Model

We can adapt and extend our original GAM model to account for a variation in model parameters before and after the change point. Since within the previous model the indicator function did not add any value to the model, we will not assess this as a term to vary our coefficients.

Table 4.14: Varying coefficient model (NO2)

Model Description	AIC	GCV	ANOVA (p-value)
$E(\log(NO2_{ijr})) = Site_{ij} \cdot Month_{ir} + s(Month_{ir}) \cdot Month_{ir} + s(Decimal.year_i) \cdot Month_{ir}$	1511.74	0.2890	---
$E(\log(NO2_{ijr})) = Site_j \cdot Month_{ir} + s(Month_{ir}) \cdot Month_{ir} + s(Decimal.year_i) \cdot Month_{ir}$	1480.76	0.2797	2.94e-07 ***

This has a GCV of 0.2797 and AIC of 1480.761. Both of these values are higher than the final GAM model shown in table 4.15.

4.4.4 Final model

Our final model has been chosen as the one with the lowest AIC and GCV score which is our generalised additive shown in figure 4.16. This will be used to track Nitrate Oxide over time at the 11 different sites at Whitelee.

Smoothing functions will also be presented to show the overall trend, seasonal trend and the interaction between these two variables.

Normality of residuals will also be assessed for the 11 different sites as will the assumption of correlated errors.

Table 4.15: ANOVA table for final model (NO2)

Model: $E(\log(NO2_{ijr})) =$ $Site_{ij} + f_1(Decimal.Year_i) + f_2(Month_{ir}) + f_3(Decimal.Year_i, Month_{ir}) +$ $f_4(Decimal.Year_i).Site_{ij} + f_5(Month_{ir}).Site_{ij}$			
	Parameters	Estimate	p
Parametric Coefficients	α	2.253	< 2e-16 ***
	Site ₂ (WL13)	0.117	0.070708 .
	Site ₃ (WL14)	0.001	0.986498
	Site ₄ (WL15)	0.152	0.019173 *
	Site ₅ (WL17)	-0.026	0.687185
	Site ₆ (WL12.16.3)	-0.043	0.501834
	Site ₇ (WL4.5.6)	0.487	1.71e-13 ***
	Site ₈ (WL9ARD)	-0.238	0.000255 ***
	Site ₉ (WL9DUN)	-0.558	< 2e-16 ***
	Site ₁₀ (WLM)	-0.108	0.095404 .
	Site ₁₁ (WLQ)	-0.052	0.415822
		EDF	
Smooth Terms	$s(Decimal.year, k = 50)$	44.5029	0.000115 ***
	$s(Month, k = 12)$	5.558	0.998338
	$s(Decimal.year): Site_1(WL1)$	0.9975	0.1488
	$s(Decimal.year): Site_2(WL13)$	1.389	0.0684
	$s(Decimal.year): Site_3(WL14)$	0.916	0.0709
	$s(Decimal.year): Site_4(WL15)$	0.916	0.0971
	$s(Decimal.year): Site_5(WL17)$	4.925	0.0628
	$s(Decimal.year): Site_6(WL2.16.3)$	0.916	0.0875
	$s(Decimal.year): Site_7(WL4.5.6)$	7.608	8.06e-06***
	$s(Decimal.year): Site_8(WL9ARD)$	5.302	0.1231
	$s(Decimal.year): Site_9(WL9DUN)$	6.952	0.000286***
	$s(Decimal.year): Site_{10}(WLM)$	0.916	0.1187
	$s(Decimal.year): Site_{11}(WLQ)$	7.8879	0.0253
	$s(Decimal.year, Month)$	18.472	0.655230
	$s(Month): Site_1(WL1)$	0.9167	0.420527
	$s(Month): Site_2(WL13)$	0.9167	0.226705
	$s(Month): Site_3(WL14)$	0.9167	0.387527
	$s(Month): Site_4(WL15)$	0.9167	0.153438
	$s(Month): Site_5(WL17)$	5.507	0.432802
	$s(Month): Site_6(WL2.16.3)$	0.9167	0.253150
	$s(Month): Site_7(WL4.5.6)$	6.352	0.000128 ***
	$s(Month): Site_8(WL9ARD)$	0.995	0.739858
	$s(Month): Site_9(WL9DUN)$	5.412	0.002324 **
	$s(Month): Site_{10}(WLM)$	1.496	0.622926
	$s(Month): Site_{11}(WLQ)$	1.72	0.251862

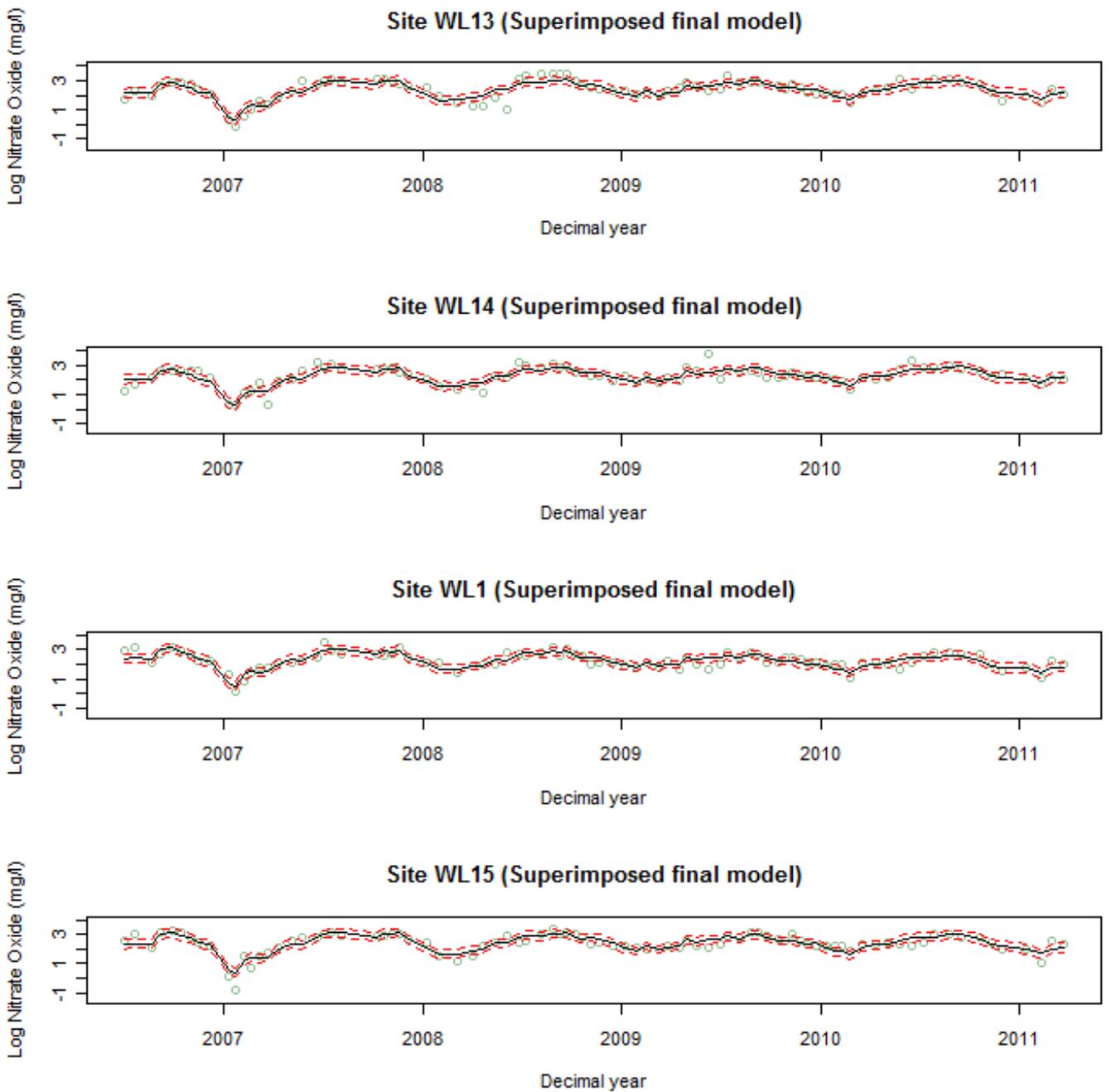


Figure 4.26: Prediction estimates and intervals (NO₂)

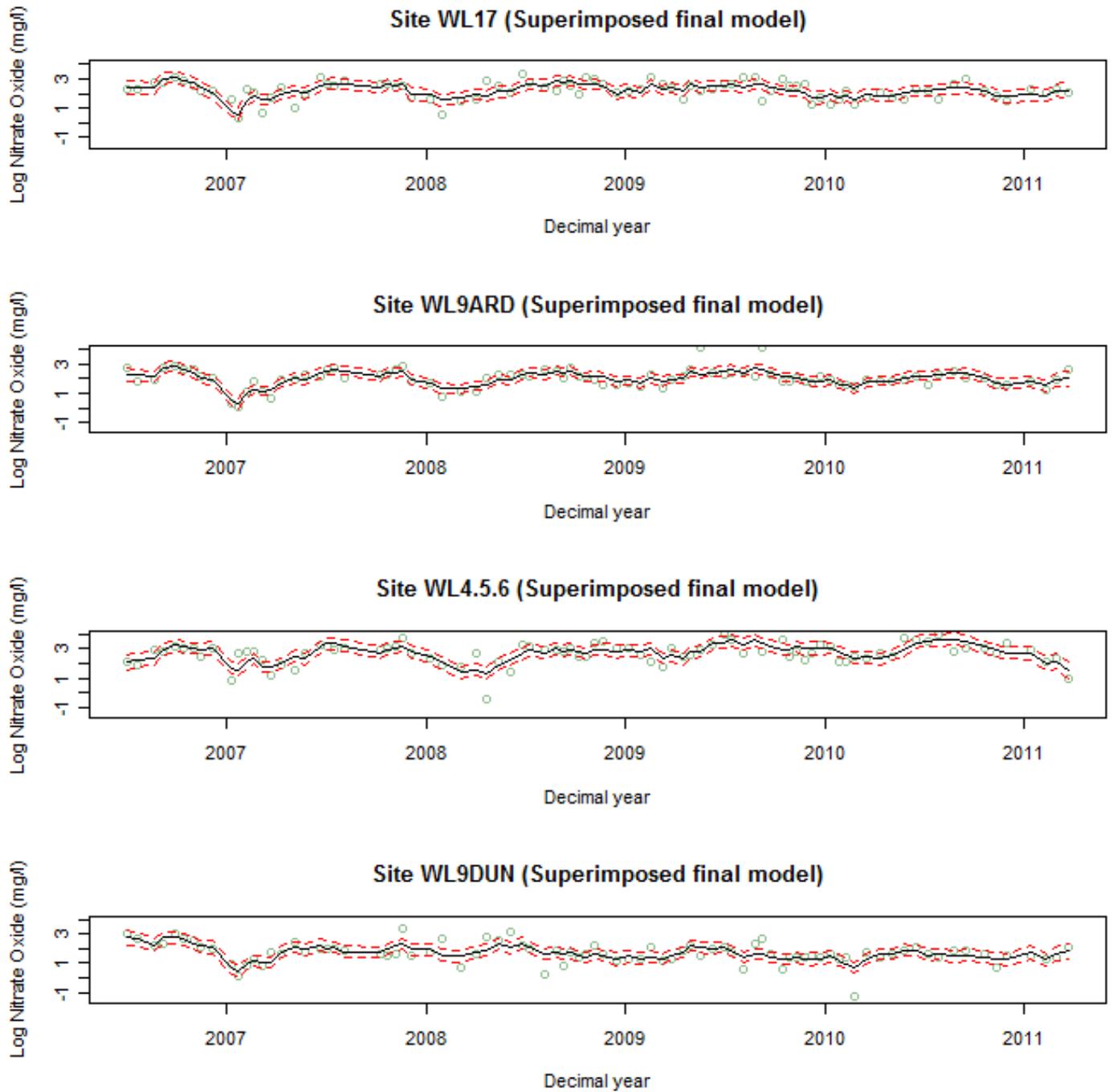
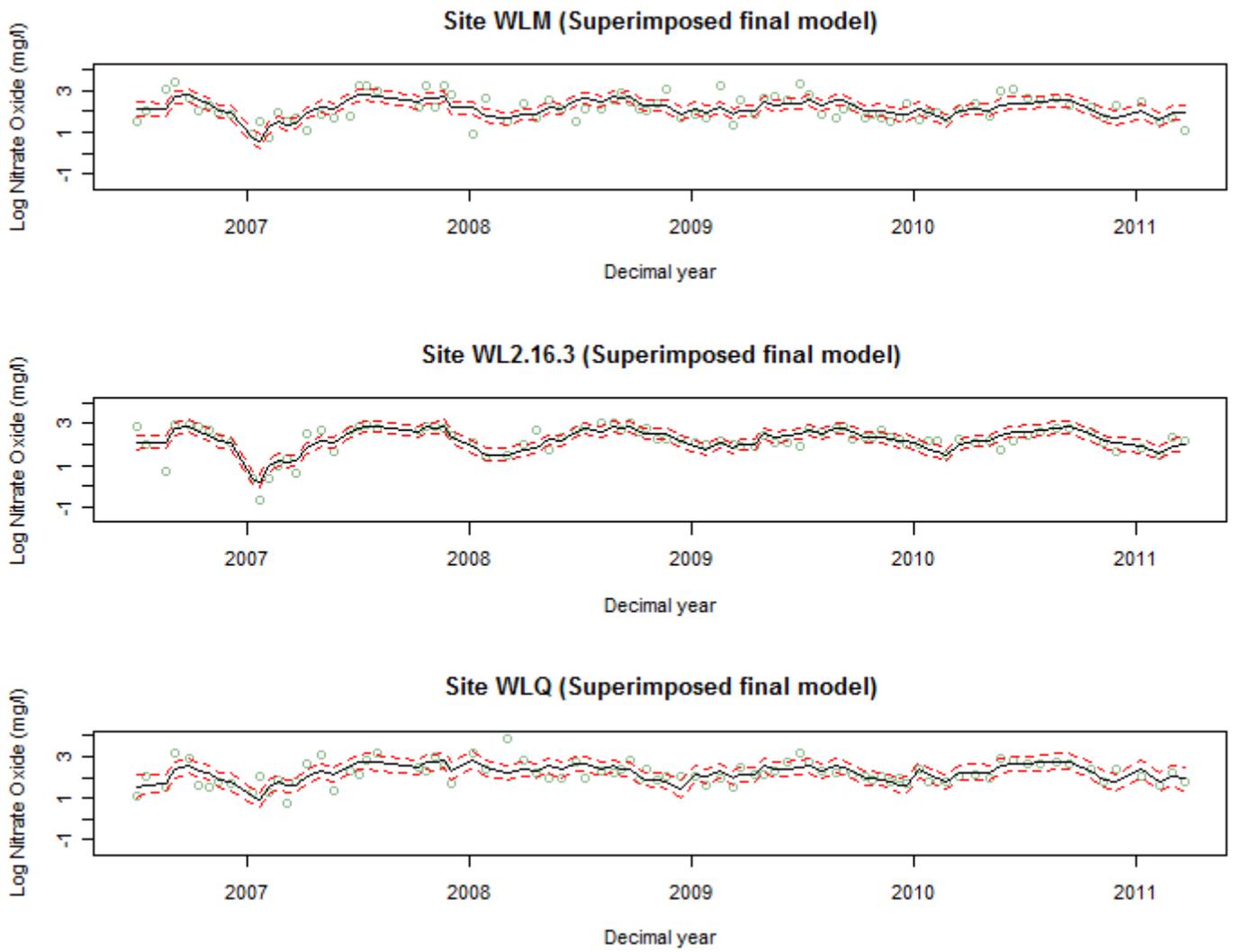
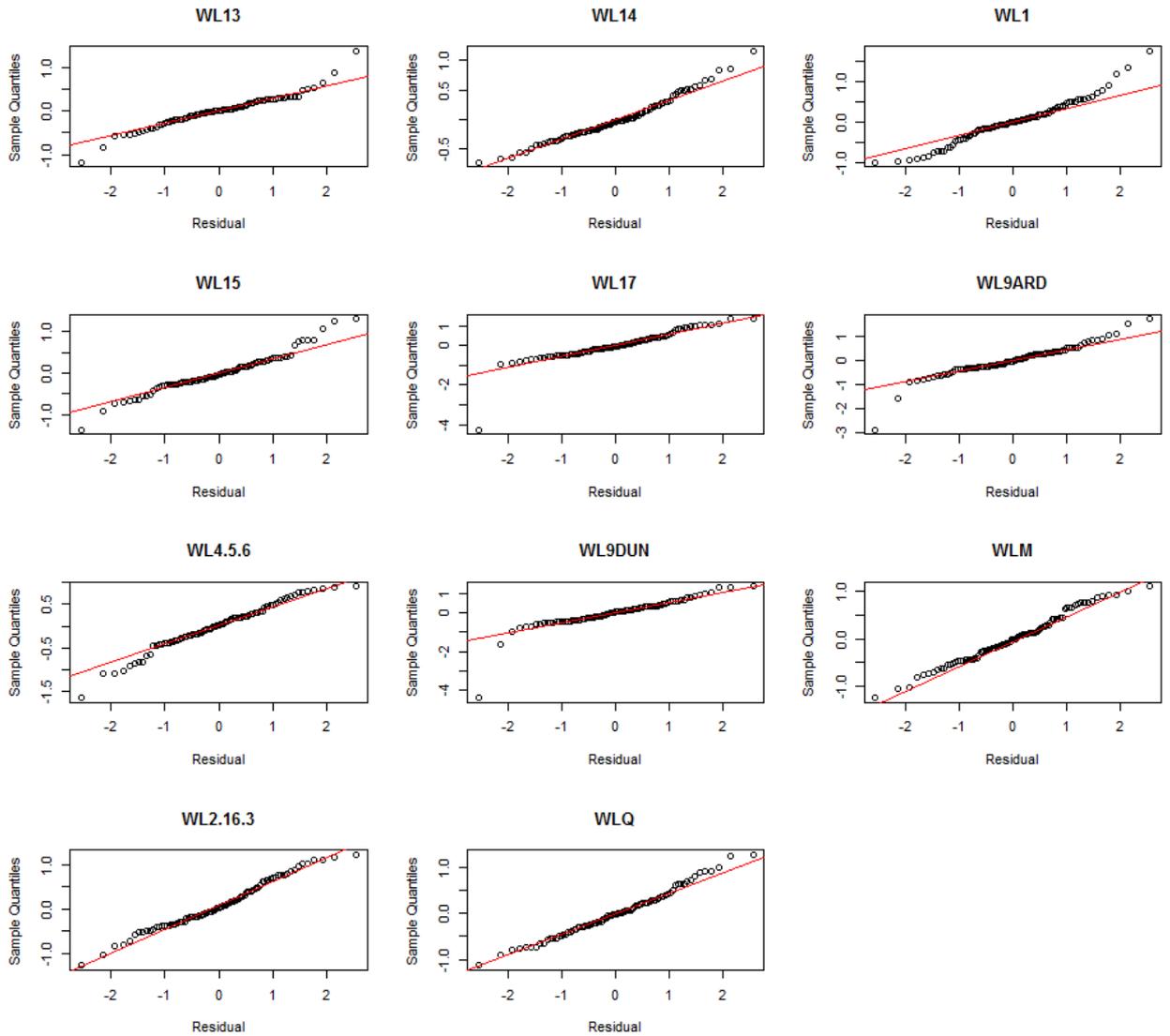


Figure 4.27: Prediction estimates and intervals (NO₂)



Figures 4.28: Prediction estimates and intervals (NO₂)



Figures 4.29: Residual plots at each site for log NO₂

Within figure 4.29 QQ normal plots of the residuals using our final model for NO₂ can be seen. Within almost all plots the points stay on the fitted line. There are a few residuals within a number of plots and the residuals for WL1 do tail off slightly. However, overall we can assume normality from these plots.

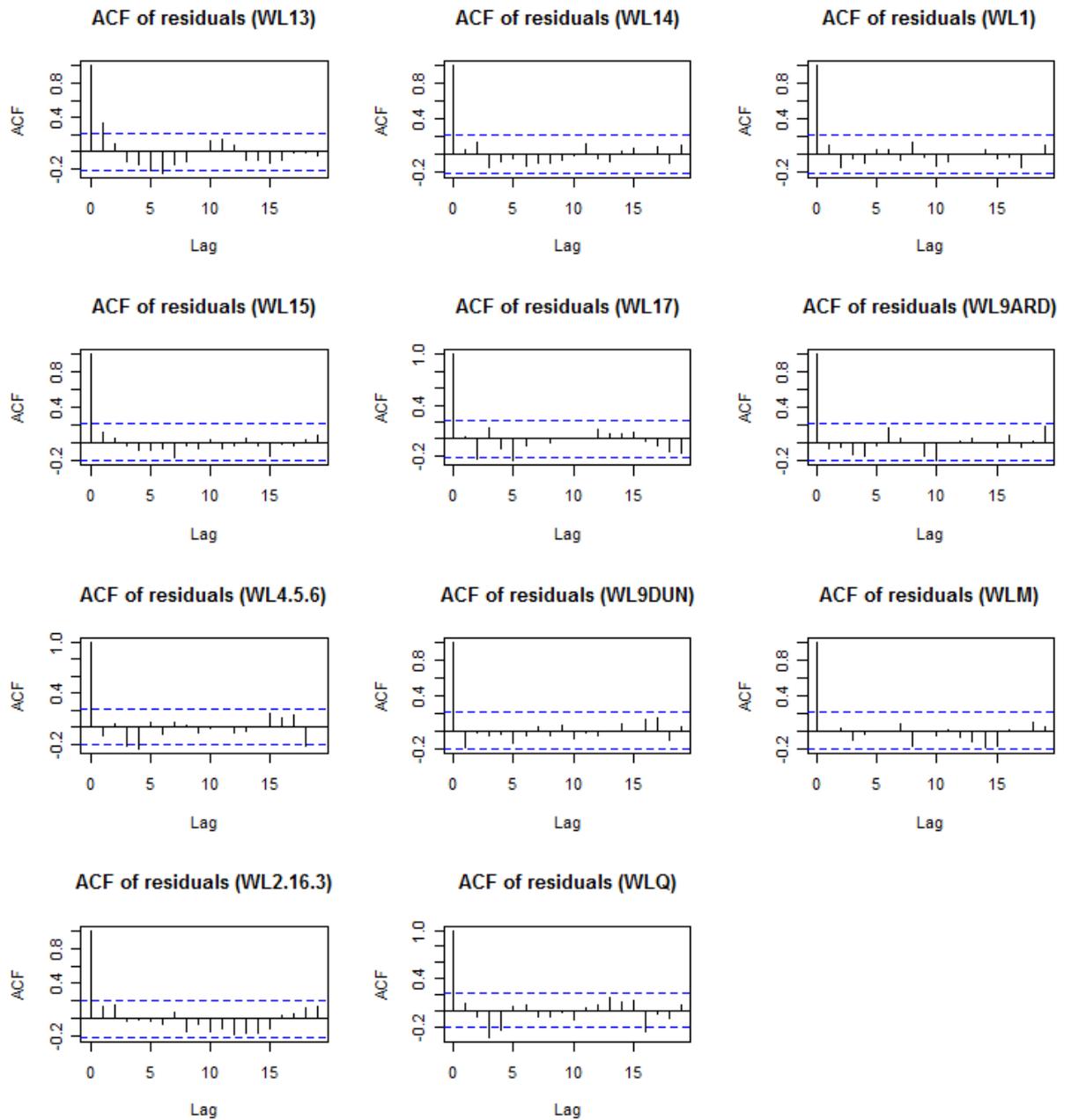


Figure 4.30: ACF plots of residual values at each site for log NO₂

The ACF plots shown above show the autocorrelation between residual values at the 11 sites at Whitelee. Within almost all plots the assumption of no autocorrelation between lags can be assumed however at site WLQ and WLM there are significant correlations at lag 1. The auto correlation at these two sites is borderline and can be ignored. Any other significant autocorrelations can also be ignored as these intervals are based on 95% confidence.

4.4.5 Conclusion (Nitrate Oxide)

From an initial impression and previous analysis, a changepoint does look present within the data with a dip in the data around mid-2007. Seasonal trend is also strong within the series.

As previous analysis allowed us to determine that a changepoint was present, an Indicator function was included within our models using GAMs and varying coefficient models.

Once our model was assessed and adjusted, the final GAM model shown below was chosen:

$$E(\log(NO2_{ijr})) = Site_j + f_1(Decimal.Year_i) + f_2(Month_{ir}) + f_3(Decimal.Year_i, Month_{ir}) \\ + f_4(Decimal.Year_i).Site_{ij} + f_5(Month_{ir}).Site_{ij}$$

However, this model does not include the Indicator function to specify the changepoint.

After checking assumptions for normality of residuals and non-correlated errors, the model was used to predict over each site.

4.6 Conclusion: Modelling

Within this chapter, two modelling approaches were considered to model the Whitelee data. Terms for seasonal trend, sites, months, overall trend and to indicator functions were used to model each series. Interaction terms and smooth terms were also included.

For all three variables, GAM's were preferred over varying coefficient models simply because they provided a better fit to the data. GAMM's were not used because there was no way to compare their fits with GAM's and varying coefficient models.

Changepoints were indicated by either a binary variable or a variable consisting of three levels depending on the number of changepoints present within the data. These 'indicator' functions allowed us to model a mean change within our series at the point where changes lay (these changes were found within chapter 3); however models without these functions actually provided a better fit to the data and therefore they were not needed.

All three final GAM models included terms for overall trend and months. This means that the levels of TP, TOC and NO₂ varied over time, and over each month.

Site was also included as a parameter in all three final GAM models; therefore the levels of the three variables also varied with site meaning different sites around Whitelee had different abundances of each determinant.

Chapter 5

Conclusions

5.1 Overview

The overall aim of this thesis was to explore changepoint detection techniques, both past and present, which are available to statisticians, as well as explore a variety of statistical testing and modelling approaches and see how they can be used within an EIA framework. The techniques which can be used, specifically to environmental time series, depend on the data which are available. The characteristics of each time series need to be examined prior to adopting techniques for example; the presence of seasonal trend, the presence of autocorrelation, the length of the series and the type of discontinuity expected will all impact on the modelling strategy.

The type of discontinuity can vary in magnitude and can be a change in mean, change in trend, change in variance and change in seasonal pattern. More complex discontinuities can include a combination of those listed above. We also need to consider whether the changepoint needs to be evaluated in a temporal or spatial context, and in a BACI setting, whether we have a control site to compare our impact series with.

The data which were used within this thesis were time series from two contexts. Our first dataset obtained from EMEP used sulphur dioxide levels in Austria and England to explore the impact of an international convention controlling discharges. The second dataset included a number of variables at Whitelee windfarm, namely Total Phosphorus (TP), Total Organic Carbon (TOC) and Nitrate Oxide (NO₂) to explore the effects (transient of permanent) of the windfarm development. 11 sites for each of these variables were assessed. Techniques had to be chosen to accommodate the varying characteristics of all of these series.

5.2 Results

Of the tests which were explored within the first few chapters of the thesis, three were chosen to use on both the EMEP and the Whitelee datasets. The performance of these tests was assessed in a simulation study where the type of change (change in mean, change in variance etc.), the magnitude of the change, the correlation, the variance and whether or not a seasonal trend was present were varied. By varying and changing the model characteristics we were able to determine which tests worked best under certain conditions. To summarise, all three tests worked well under different conditions. Our first test, Local linear regression, worked well when a mean change was present however the size of the test was affected by both correlation and seasonal trend. Barry and Hartigans algorithm worked extremely well but the size of the test was drastically affected by the presence of seasonal trend. Binary segmentation worked well under all conditions, even when seasonal trend and correlation were present.

Since two of the three tests were affected by the presence of seasonal trend, harmonic functions were fit to the data and difference series were calculated for the Whitelee time series. The EMEP series were weekly averages so seasonal trend was weak and no alterations were needed.

5.3 EMEP Results

Two series were analysed using changepoint detection techniques from EMEP. Firstly, site AT02 was analysed which is a site located at the south eastern tip of Austria. In total, 11 changepoints were detected, 3 by LLR, 5 by B&H and 3 by BSM. One cluster of 3 changepoints were located in 2003 and another three around the end of 2005.

Analysis on GB02 resulted in a total of 12 changepoints, three were detected by LLR, 4 by B&H and 5 by BSM. Detected changepoints were relatively spread out with no clustering. Two separate changepoints by two separate methods (B&H and BSM) detected changepoints around half way through 1996.

From the analysis on GB02, there does not seem to be any evidence to suggest changepoints. There is possibly a changepoint mid 2006 as two changepoints were detected there however as all other detected changepoints are spread out results do not suggest a significant change in model trend. It is possible that changepoints are located in the AT02 time series, either in 2003 or at the end of 2005 as two separate clusters of changepoints were located around these times.

5.4 Whitelee results

Our first analysis performed on Whitelee was a BACI design experiment. This method consists of applying a linear model to the data with terms for indicating whether the point of interest is before or after the changepoint (named 'indicator'), overall trend, month and whether a site is a control or an impact site (named 'control/impact'). Interaction terms between the indicator term and the control/impact term were also included. The models were then analysed within an ANOVA table and if any of these terms were insignificant, they were removed. Model assumptions were analysed using residuals vs. fitted values for constant variance and histograms were used to assess normality of residuals.

The final model applied to our total phosphorus data consisted of all 5 terms including the interaction term. All of these terms were significant within an ANOVA table and therefore the conclusion was made that mean log TP differed depending on whether it was measured before or after the intervention point and it also differed between control and impact sites.

The final model for TOC included 3 terms, time, month and control/impact. Therefore log TOC varied by whether the site was a control or impact site, but not whether the measurement was before or after intervention.

The final model for NO₂ included 3 terms, month, indicator and control/impact. Therefore log NO₂ varied by site and depended on whether the measurement was before or after intervention.

Chapter 3 consisted of a review of various changepoint detection techniques. Of the techniques reviewed, three were chosen to analyse our datasets for both the EMEP data and the Whitelee data. Local linear regression (LLR), a technique based on normal kernel smoothing, Barry and Hartigans (B&H) algorithm based on Bayesian theory and Binary Segmentation (BSM) based on likelihood were chosen to analyse the datasets. A simulation study was carried out on 6 different scenarios for each of the methods, with each scenario becoming increasingly more complicated.

To summarise, LLR performed well when a change of mean was present and also performed satisfactory when seasonal trend was present however the kernel bandwidth had to be

reduced when seasonal trend was present. B&H performed very well when seasonal trend was not present. When seasonal trend is present within our datasets, a linear model with harmonic components was used to remove this trend. BSM performed very well under all conditions.

Once all three methods were applied to log Total Phosphorus, 16 changepoints were detected between the three methods with 13 grouped closely in mid-2007 and 3 grouped closely in mid-2009.

15 changepoints were detected within our log Total Organic Carbon data however these changepoints were relatively spread out with no clustering.

For log Nitrate Oxide, 19 changes were detected. 14 were located around 2007 and 5 were located mid-2009.

Modelling was carried out on the Whitelee data using Generalised Additive Models and Varying Coefficient models. Changepoints detected from chapter 3 were used within our models to indicate where intervention points were found.

Both modelling techniques used terms for overall trend, month, site, and also indicator terms. Smooth functions of these terms and interactions between them were also used. Comparisons between models were made with both the AIC and GCV, if there was any doubt whether a model should be selected over another, an ANOVA between the two models was calculated.

The final model selected for modelling log TP was a GAM model included terms for site, smooths for decimal year and month and a smooth interaction between decimal year and month as well as smooths for decimal year and month which vary by site. All of these terms were significant within an ANOVA table. For both log TOC and log NO₂, the final models included terms for site, a smooth term for the overall trend and month as well as a smooth term for the interaction between the overall trend and month and terms which allowed both the overall trend and month to vary with site. There was no indicator term within these models.

Overall, the models applied to the three determinands fitted the data well. Indicator terms were not needed in any of the three models since the models tracked the data closely enough without the term. After these models were built, prediction estimates and 95% confidence intervals were calculated and plotted for each site. Normality of residuals was assessed using QQ plots and ACF plots allowed us to check for autocorrelation in the lags.

Through our BACI analysis, changepoint analysis and modelling through various techniques, we can draw together what has been found to identify the effect of the wind farm development on the three determinands that have been assessed. If the 'Indicator' function within the BACI analysis was significant, then this indicates a significant change in mean at the point in time that the indicator function changes value from 0 to 1. BACI analysis also allowed us to quantify the magnitude of this difference. Furthermore, clustering of changepoints within our changepoint analysis section also indicates a significant change in model terms.

Our analysis for Total Phosphorus suggests a changepoint in mid-2007. The significance of the 'Indicator' function within the BACI analysis and the clustering of changepoints detected by LLR, B&H and BSM in mid-2007 indicates a change at this point in time. There does not seem to be any evidence to suggest a changepoint within our series for Total Organic Carbon. The BACI analysis concluded that the 'indicator' term was insignificant and no clusters of changepoints were found when our three changepoint detection methods were applied. Within the NO₂ series, it is likely that there is a changepoint mid-2007. The BACI analysis concluded that there was a significant change in mean before and after intervention. The changepoint analysis showed clusters in 2007 and a small cluster around mid-2009.

To conclude, there were changepoints identified mid-2007 in both the log Total Phosphorus series and the log Nitrate Oxide series since BACI analysis showed a significant change in mean at this point and changepoint analysis identified changes at this point in time. The log Total Organic Carbon series did not have a changepoint at mid-2007 since there was no significant change in mean identified from BACI analysis and there was no clustering of changepoints at this time.

5.5 Discussion and limitations

One of the main problems within BACI analysis was determining which sites should be considered as impact sites and which should be considered as control sites. BACI analysis should be designed from the outset and sites should be set as control or impact sites prior to the study beginning. This could not be done in this case as the data had already been collected; therefore control sites were selected as those which had the lowest percentage of deforested area around the site.

Missing values were a major concern within the analysis, particularly within the EMEP sites where large chunks of data were missing. Local linear regression was chosen to estimate missing values, however within the GB02 series much of the data towards the end of the series was missing and therefore much of the data towards the end of this series was estimated.

Another problem is the fact that within our changepoint analysis, it is likely that many detected changepoints were falsely identified. For example, in the TOC series changepoints were detected relatively spread out without any clustering. Further analysis could include adjusting the threshold probability for B&H, the smoothing parameter for LLR and the penalty for BSM so that it is less likely that a change is falsely detected.

References

- Abraham, B. (1980). Intervention Analysis and Multiple Time Series. *Biometrika.*, 67, 73-78.
- Alexanderson, H. A., (1986). A homogeneity test applied to precipitation data. *J. Climatol.*,6, 661–675.
- Bagshaw, M. and Johnson, R. (1975). The effect of serial correlation on the performance of CUSUM tests. *Technometrics.*, 17,73-80.
- Beaulieu, C., Chen, J. and Saramiento, JL. (2010). Change point analysis as a tool to describe past climate variations. -“Climate predictions: the influence of nonlinearity and randomness”. *The Royal Society.*, 370, 1228-1249.
- Beaulieu, C., Saramiento, JL., Mikaloff Fletcher, SE., Chen, J., Medvigy D. (2011). Identification and characterization of abrupt changes in the land uptake of carbon. *Global Biogeochemical Cycles.*, 26
- Beaulieu, C., Ouarda T., Seidou O. (2010). A Bayesian normal homogeneity test for the detection of artificial discontinuities in climatic series. *International Journal of Climatology.*, 30, 2342–2357.
- Bender, E.A., Case, T.J. & Gilpin, M.E. (1984). Perturbation experiments in community ecology: theory and practice. *Ecology.*, 65, 1-13.
- Bernhardson, C.S. (1975). Type I Error Rates When Multiple Comparison Procedures Follow a Significant F Test of ANOVA. *Biometrics.*, 31, 229-232.
- Bernstein, B.B. and Zalinski, J. (1983). An optimum sampling design and power tests for environmental bi-ologists. *J. Environ. Manage.*, 16, 35-43.
- Boland, P.J. (1984). A Biographical Glimpse of William Sealy Gosset. *The American Statistician.*, 38, 179-183.
- Bowman, A.W. and Azzalini, A. (2003). Computational aspects of nonparametric smoothing with illustrations from the sm library. *Computational Statistics and Data Analysis* 42, 545-56.
- Bowman, & Azzalini. (2006). *Applied Smoothing Techniques for Data Analysis*. Oxford Science Publications.
- Bowman, A.W., Pope, A. & Ismail B. (2006). Detecting discontinuities in nonparametric regression curves and surfaces. *Statistics & Computing*. 16:377-390.
- Box, G. E. P. and Tiao, G. C. (1975). Intervention Analysis with Applications to Economic and Environmental Problems. *J. Am. Statis. Assoc.*, 70, 70-79.

- Cai, Z., Fan, J. and Li, R. (2000). Efficient estimation and inferences for varying coefficient models. *Journal of the American Statistical Association*. 95, 888-902.
- Chen, R. and Tsay, R.S. (1993). 'Functional coefficient autoregressive models', *Journal of American Statistical Association*, 88, 298-308.
- Clarke, K.R. (1993). Non-parametric multivariate analyses of changes in community structure. *Aust. J. Ecol.* 18, 117 – 43.
- De-Lacy, M.C., Reguzzoni, M., Sanoso, F., and Venuti, G. (2008). The Bayesian detection of discontinuities in a polynomial regression and its application to the cycle-slip problem. *Journal of Geodesy.*, 82, 527-542.
- Downton, M.W. and Katz, R.W. (1993). A Test for Inhomogeneous Variance in Time-averaged Temperature Data. *J. Climate.*, 6, 2448–2464.
- Eisenhart, C. (1979). On the Transition from "Student's" z to "Student's" t. *The American Statistician.*, 33, 6-10.
- Eubank, R. and Speckman, P. (1994). Nonparametric estimation of functions with jump discontinuities. Change point problems. *Institute of Mathematical Statistics Lecture Note Series.*, 23, 130-144.
- Ferguson, C.A., Bowman, A.W., Scott, E.M., and Carvalho, L. (2007). Model comparison for a complex ecological system. *Journal of the Royal Statistical Society: Series A (Statistics in Society)*, 170 . pp. 691-711. ISSN 0964-1998
- Fisher R.A. (1925). Theory of statistical estimation. *Proc. Cambridge Philos. Soc.*, 22,700-725.
- Gérard-Marchant, Pierre G. F., Stooksbury, DE., Seymour, L. (2008). Methods for Starting the Detection of Undocumented Multiple Changepoints. *J. Climate*, 21, 4887–4899.
- Green, P (1995). Reversible jump markov chain monte carlo computation and Bayesian model determination. *Biometrika.*, 82, 711-732.
- Green, R.H. (1979). Sampling design and statistical methods for environmental biologists. Wiley, New York.
- Gregoire, G and Hamrouni, Z. (2002a). Change points estimation by local linear smoothing. *Journal of Multivariate Analysis.*, 83, 56-83.
- Hall P. and Titterington D.M. (1992). Edge-preserving and peak preserving smoothing. *Technometrics.*, 34, 429–440.
- Hastie, T. J., Tibshirani, R. (1993). Varying-coefficient models. *Journal of the Royal Statistical Society.* 4, 757-796.

- Hastie, T. J., Tibshirani, R. and Buja, A. (1994). "Flexible Discriminant Analysis by Optimal Scoring." (Bell Labs Technical report; postscript).
- Hewitt J.E., Thrush S.E., Cummings V.J. (2001). Assessing Environmental Impacts: Effects of Spatial and Temporal Variability At Likely Impact Scales. *Ecological Applications.*, 11,1502–1516.
- Hinkley, D. V. (1970) Inference About the Change-Point in a Sequence of Random Variables, *Biometrika.*, 57, 1-17.
- Hinkley, D.V. (1969). Inference about the intersection in two-phase regression. *Biometrika.*, 56, 495.
- Hipel, K.W. and McLeod, A.I. (1989). Intervention analysis in environmental engineering. *Environmental Monitoring and Assessment.*, 13, 185-201.
- Hirsch, R.M. & Slack, R.J. (1984). A Nonparametric Trend Test for Seasonal Data With Serial Dependence. *WATER RESOURCES RESEARCH.*, 20, 727-732.
- Hogan T.D. (1984). An Intervention Analysis of the effects of legalized abortion upon U.S fertility. *Population Research and Policy Review.*, 3, 201-218.
- Hurlbert, S.J, (1984). Pseudoreplication and the design of ecological field experiments. *Ecol. Monogr.*, 54, 187-211.
- Inclan, C and Tiao, G. (1994). Use of Cumulative Sums of Squares for Retrospective Detection of Changes of Variance. *Journal of the American Statistical Association.*, 89, 913-923.
- Killick R., Fearnhead P. & Eckley I.A. (2011). Optimal detection of changepoints with a linear computational cost. [arXiv:1101.1438](https://arxiv.org/abs/1101.1438)
- Knott, N.A. (2009). Contemporary ecological threats from historical pollution sources: impacts of large-scale resuspension of contaminated sediments on sessile invertebrate recruitment. *Journal of Applied Ecology.*, 46, 770-781.
- Lepage, Y. (1971). A combination of Wilcoxon's and Ansari-Bradley's statistics. *Biometrika.*, 58, 213–217.
- Lombard F. (1987). Rank tests for changepoint problems. *Biometrika* 74(3): 615–624.
- Lund, R & Reeves, J (2002). Detection of Undocumented Changepoints: A Revision of the Two-Phase Regression Model. *Journal of climate.*, 15, 2547-2554.
- Muller H.G. (1992). Change-points in nonparametric regression analysis. *Ann. Statist.*, 20, 737–761.
- Murakami, H (2011). A nonparametric location-scale statistic for detecting a change point. *Int J Adv Manuf Technol.*, 61, 449-455.

- Murray, H (2012). Assessing the impact of windfarm-related disturbance on streamwater carbon, phosphorus and nitrogen dynamics. A case study of the Whitelee catchments. PhD thesis, University of Glasgow.
- Musco, L., Terlizzi, A., Licciano, M. (2009). Taxonomic structure and the effectiveness of surrogates in environmental monitoring: a lesson from polychaetes. *Mar. Ecol. Prog. Ser.*, 383, 199-210.
- Nam, C.F.H., Aston J.A.D. and Johansen, A.M. (2012). Quantifying the Uncertainty in Change Points. *Journal of Time Series Analysis.*, 33, 807-823.
- Ning, W., Pailden, J., Gupta, A. (2012). Empirical Likelihood Ratio Test for the Epidemic Change Model. *Journal of Data Science.*, 10, 107-127.
- Oaten A.S. & Bence J.R. (2001). Temporal and Spatial Variation in Environmental Impact Assessment. *Ecological Monographs.*, 71, 305-339.
- Oaten A.S. (2002). Impact assessment. In A. H. El-Shaarawi and W.W. Piegorsch, editors. *Encyclopedia of Environmetrics*. 2.
- Oaten A.S. (2003). On Rejection Rates of Paired Intervention Analysis: Comment. *Ecological Society of America.*, 84, 2795-2799.
- Oaten A.S., Murdoch W., Parker K.R. (1986). Environmental Impact Assessment: "Pseudoreplication" in Time?. *Eco. Soc. America.*, 67, 929-940.
- Osenberg, C.W., Schmitt, R.J., Holbrook, S.J., Abu-Saba, K.E., Flegal, A.R. (1994) Detection of Environmental Impacts: Natural Variability, Effect Size and Power Analysis. *Ecological Applications.*, 4, 16-30.
- Page, E. (1954). Continuous inspection schemes. *Biometrika.*, 41, 100-115.
- Page, E. (1955). A test for a change in a parameter occurring at an unknown point. *Biometrika.*, 42, 523-527.
- Pinheiro, J.C., and Bates, D.M. (2000) "Mixed-Effects Models in S and S-PLUS", Springer
- Perneger T. V. (1998). What's wrong with Bonferroni adjustments. *BMJ.*, 316, 1236-1238.
- Perreault, L., Bernier, J., Bobee, B., Parent, E. (2000). Bayesian change-point analysis in hydrometeorological time series. Part 1. The normal model revisited. *Journal of Hydrology.*, 235, 221-241.
- Pettitt, A. (1980). A simple cumulative sum type statistic for the change-point problem with zero-one observations. *Biometrika.*, 67, 79-84.
- Pettitt, A.N. (1976). A two-sample Anderson-Darling rank statistic. *Biometrika.*, 63, 161-168.

Pettitt, AN (1979) A non-parametric approach to the change point problem. *Appl Stat.*, 28, 126–135.

Queiroz, N., Lima, FP., Riberio, PA., Pereira, SG. (2006). Using asymmetrical designs for environmental impact assessment of unplanned disturbances. *Hydrobiologia.*, 555, 223-227.

Quessy, JF., Favre, AC., Said, M., Champange, M. (2011). Statistical inference in Lombards smooth change model. *Environmetrics.*, 22, 882-893.

Reeves, J. (2006). A review and comparison of change point detection techniques for climate data. *Journal of Applied Metrology and Climatology.*, 46, 900-914.

Rodionov, S. and Overland, J.E. (2005). Application of a sequential regime shift detection method to the Bering Sea ecosystem. *Journal of Marine Science.*, 62, 328-332

Rodionov, S. N., (2005): A brief overview of the regime shift detection methods. *Large-Scale Disturbances (Regime Shifts) and Recovery in Aquatic Ecosystems: Challenges for Management toward Sustainability*, UNESCO–OSTE/BAS Workshop on Regime Shifts, Varna, Bulgaria, UNESCO–ROSTE/BAS, 17–24.

Rodionov, S.N., (2005b): A sequential method for detecting regime shifts in the mean and variance. In: *Large-Scale Disturbances (Regime Shifts) and Recovery in Aquatic Ecosystems: Challenges for Management Toward Sustainability*, V. Velikova and N. Chipev (Eds.), UNESCO-ROSTE/BAS Workshop on Regime Shifts, 68-72.

Rotondi, R. (2002). On the influence of the proposal distributions on a reversible jump MCMC algorithm applied to the detection of multiple change-points. *Computational Statistics & Data Analysis.*, 40, 633-655.

Ruppert, D., Wand, M.P. and Carroll, R.J. (2003). *Semiparametric Regression*. Cambridge

Ruxton G.D. (2006). The unequal variance *t*-test is an underused alternative to Student's *t*-test and the Mann–Whitney *U* test. *Behav. Ecol.*, 17, 688–690

Scott, A. J. and Knott, M. (1974) A Cluster Analysis Method for Grouping Means in the Analysis of Variance, *Biometrics* 30(3), 507–512

Sen, A. & Srivastava, M. S. (1975). On tests for detecting a change in mean. *Annals of Statistics* 3., 98–108.

Shapiro S. S. and Wilk M. B. (1965). An Analysis of Variance Test for Normality (Complete Samples). *Biometrika.*, 52, 591-611.

Smith E.P. (2002). BACI design. *Encyclopedia of Environmetrics.*, 1, 141–148.

- Smith E. P., Pontasch, KW., Cairns, J. (1990) Community similarity and the analysis of multispecies environmental data: a unified statistical approach. *Water Res.* 24, 507 – 14.
- Smith, E.P., Orvos, D. & Cairns, J. (1993). "Impact assessment using the before-after control-impact model: Comments and concerns." *Canadian Journal of Fisheries and Aquatic Science.*, 50, 627-637.
- Sollow, A.R. (1987). Testing for climate change: An application of the two-phase regression model. *Journal of climate and applied meteorology.*, 26, 1401-1405.
- Speckman, P. (1994). Detection of change points in nonparametric regression, University of Missouri-Columbia
- Terlizzi A., Franschettie, S. (2005). Multivariate and univariate asymmetrical analyses in environmental impact assessment: a case study of Mediterranean subtidal sessile assemblages. *Marine Ecology Progress Series.*, 289, 27-42.
- Underwood A.J. (1992). Beyond BACI: the detection of environmental impacts on populations in the real, but variable, world. *J. Exp. Mar. Biol. Ecol.*, 161, 145-178.
- Underwood, A.J. (1994). On beyond BACI: Sampling designs that might reliably detect environmental disturbances. *Ecol. Appl.*, 4(1): 3-15
- Underwood, A.J. (1989). The analysis of stress in natural populations. *Biol. J. Linn. Soc.*, 37, 51- 75.
- Underwood, A.J. (1991a). Beyond BACI: Experimental designs for detecting human environmental impacts on temporal variations in natural populations. *Aust. J. Mar. Freshwater Res.*, 42, 569-587.
- Underwood, A.J., (1991b). Biological monitoring for human impact: how little it can achieve. In, *Proceedings of the 29th Congress of the Australian Society of Limnology*, Jabiru, NT, 1990, edited by R.V. Hyne, Office of the Supervising Scientist, Alligator Rivers Region, Australian Government Printing Service, pp. 105-123.
- Vincent, L.A. and Gullet D.W. (1999). CANADIAN HISTORICAL AND HOMOGENEOUS TEMPERATURE DATASETS FOR CLIMATE CHANGE ANALYSES. *Journal of Climatology.*, 19, 1375-1388.
- Wang, X.L. (2003). Comments on "Detection of undocumented change points: A revision of the two-phase regression model." *J. Climate.*, 16, 3383–3385.
- Welch, B.L. (1938). The significance of the difference between two means when the population variances are unequal. *Biometrika.*, 29, 350-362.
- Western, B, Kleykamp M. (2004). A Bayesian Change Point Model for Historical Time Series Analysis. : Princeton University
- Wood, S.N. (2006) *Generalized Additive Models: An Introduction with R.* CRC/ Chapman and Hall, Boca Raton, Florida.

Wyse, J., Friel, N. (2010). Simulation-based Bayesian analysis for multiple change points. arXiv:1011.2932v1

Wyse J., Friel, N., Rue, HJ. (2011). Approximate simulation-free Bayesian inference for multiple change point models with dependence within segments. *Bayesian Analysis.*, 6, 501-528.

Yashchin, E. (1993). Performance of CUSUM control schemes for serially correlated observations. *Technometric.s*, 35, 37-52.

Zhao, X., Chu, PS. (2008). Bayesian Multiple Changepoint Analysis of Hurricane Activity in the Eastern North Pacific: A Markov Chain Monte Carlo Approach. *Journal of Climate.*, 19, 564-578.

Zhao, X., Chu, PS. (2010). Bayesian Changepoint Analysis for Extreme Events (Typhoons, Heavy Rainfall, and Heat Waves): An RJMCMC Approach. *J. Climate.*, 23, 1034–1046.